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FLORIDA INTERNATIONAL UNIVERSITY

Landscape Pattern- Ridge, Slough, and Tree Island Mosaics

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INTRODUCTION

Status and history of the Ridge-Slough Mosaic

The Florida Everglades is a large subtropical wetland with diverse hydrologic, edaphic, and vegetative characteristics. Historically, a significant portion of this system was a slow moving river originating from the Kissimmee River floodplain, flowing into the vast but shallow Lake Okeechobee, and draining south-southwest over extensive peatlands into Florida Bay (McVoy et al. 2011). Human-induced alterations to the hydrologic regime, including reduction, stabilization, and impoundment of water flow through diversion and compartmentalization of water via canals and levees have degraded pre-drainage vegetation patterns and microtopographic structure (Davis and Ogden 1994, Ogden 2005, McVoy et al. 2011).

The Everglades peatland emerged 5,000 years ago with the stabilization of sea level at approximately current elevations (Loveless 1959, Gleason and Stone 1994). This, combined with subtropical rainfalls, allowed a vast mass of water to slowly flow over a limestone bedrock platform 160 km long and 50 km wide at a near uniform descent totalling about 6 m, ultimately reaching Florida Bay (Stephens 1956, Gleason and Stone 1994, McVoy et al. 2011). Vegetation quickly colonized the area, and peat, in the absence of adequate respiration, accumulated on the limestone bedrock to a depth of 3-3.7 m (Gleason and Stone 1994, McVoy et al. 2011). The “River of Grass” referenced by Douglas (1947) alludes to the dually intertwined processes of the historic riverine nature of the Everglades and the vast sawgrass (*Cladium jamaicense*) communities that have dominated the landscape for about the last 1,000 years (Bernhardt and Willard 2009).

McVoy et al. (2011) identify eight major historic landscapes that comprised the greater Everglades: Custard Apple Swamp, Sawgrass Plains, Ridge and Slough, Peat Transverse Glades, Rockland Marl Marsh, Marl Transverse Glades, Perrine Marl Marsh, and the Ochopee Marl Marsh. Of these, the Ridge and Slough (c. 6,000 km²) encompassed slightly over 50% of the total extent (c.11,000 km²). Prior to hydrologic modification, this landscape provided biotic communities with distinct elevational niches that were organized in a characteristic pattern parallel to flow (Fig. 1). Ridges, comprised almost totally of dense stands of sawgrass, were located in areas of higher topographic relief (and therefore lower water depths), whereas sloughs containing white water lily (*Nymphaea odorata*), other macrophytes, and periphyton were at lower elevation and therefore deeper water (Loveless 1959, Ogden 2005, McVoy et al. 2011). A third community, the wet prairie, was comprised of *Eleocharis cellulosa* (spikerush), *Panicum hemitomon* (maidencane), and *Rhynchospora tracyi* (beakrush), and usually occupied the physical zone between the boundary of ridges and sloughs, in areas of intermediate water depths (Loveless 1959, Ogden 2005). The Ridge and Slough landscape was interspersed with high elevation tree islands which support hardwood communities, and portion of which remain fully emergent throughout the year (Loveless 1959, Sklar and van der Valk 2002, van der Valk and Warner 2009).

As in all wetlands, the hydrologic regime is a critical factor influencing the distribution and composition of vegetation in the greater Everglades (Gunderson 1994, Ross et al. 2003, Armentano et al. 2006, Zweig and Kitchens 2008, Todd et al. 2010). Local variation in hydrologic conditions resulting from microtopographic differentiation is essential for the maintenance of the distinct vegetation community boundaries that were a feature of the pre-drainage Ridge and Slough

landscape (Loveless 1959, Ogden 2005, McVoy et al. 2011). In the current managed Everglades system, the pre-drainage, patterned mosaic of sawgrass ridges, sloughs and tree islands (Fig. 1) has been substantially altered or reduced largely as a result of human alterations to historic ecological and hydrological processes (Larsen et al. 2011).

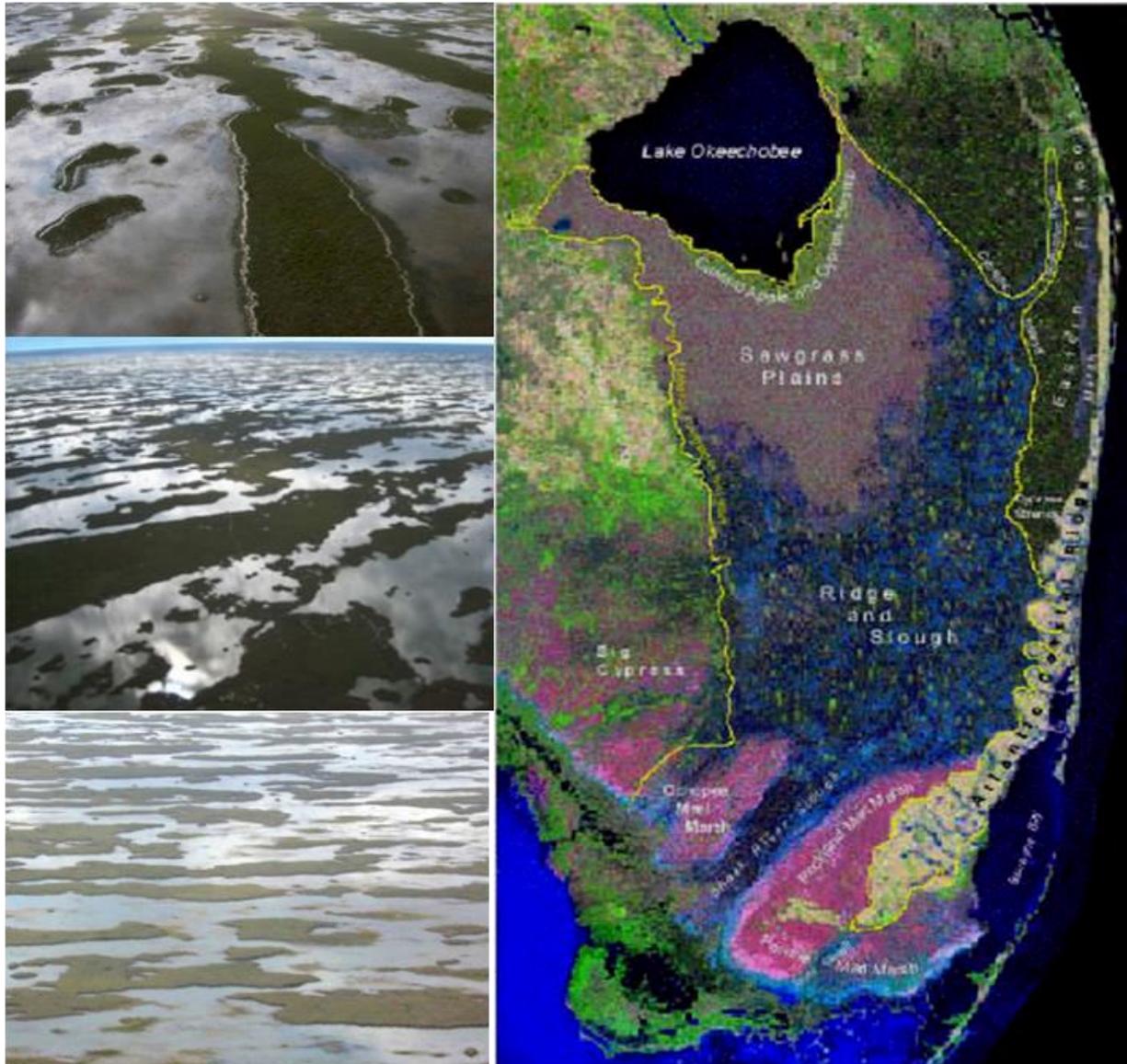


Figure 1: Aerial images and historic distribution of the ridge-slough landscape. (Left) Linear, flow-parallel orientation of ridges and sloughs under conserved conditions. (Right) Distribution of ridge and slough and other landscape types prior to major hydrologic

The Ridge and Slough community has undergone dramatic structural, compositional and functional changes since anthropogenic modification of the hydrologic regime began in the early 20th century (Davis and Ogden 1994, Team 2003, Ogden 2005, Larsen et al. 2011, McVoy et al. 2011). Average water levels across the Ridge and Slough landscape are estimated to be 30-45 cm lower between the pre-modification era and 1955, but have since rebounded slightly through concentrated restoration efforts (McVoy et al. 2011). Nonetheless, sloughs in many areas of the Everglades frequently “dry-down” by the end of the dry season, but in the pre-modified landscape it is believed that sloughs dried very rarely, supported by an estimation of average depth at 90 cm (McVoy et al. 2011). These changing hydrologic regimes result from some combination of altered water flows and loss of microtopographic structure.

Lower hydroperiods have severely restricted the range and abundance of many slough species, including the historically ubiquitous white water lily, and even more drastically restricted spatterdock (*Nuphar advena*), a species adapted to the deepest water (McVoy et al. 2011). Similarly, hydrologic modification has led to an expansion among species better adapted to low water, such as sawgrass and southern cattail (Urban et al. 1993, Davis and Ogden 1994, Gunderson 2001, McVoy et al. 2011). Woody vegetation may have been uncommon in the ridge community prior to hydrologic modification (Loveless 1959, McVoy et al. 2011), but wax myrtle (*Myrica cerifcra*) and coastal plain willow (*Salix caroliniana*) now frequently inhabit ridges in drained areas (McVoy et al. 2011). Ross et al. (2003) found that subtle but significant differences in hydroperiod define the boundaries between ridge, slough, and wet prairie communities in Shark Slough, ENP, but these boundaries are eroding with hydrologic modification. Similarly, working in a marl prairie environment in Taylor Slough, ENP, Armentano et al. (2006) found that changes in the hydrologic regime over periods as brief as three years had induced concurrent changes in vegetation composition and community patterning. Zweig and Kitchens (2008, 2009) found vegetation communities in southern Water Conservation Area 3A (WCA 3A) are influenced by both current and historic hydrologic conditions, and vegetation responses to hydrologic modification varied among species. Loss of historically distinct vegetation communities thus appears to be a characteristic of ridges and sloughs to hydrologic modification.

Hydrologic modification also has consequences for the landscape-scale structure of the ridge-slough mosaic (Fig. 2). Where hydroperiods have been reduced, ridges have invaded marsh areas (Science Coordination Team 2003, Ogden 2005), and much of the slough component of the landscape, particularly in the Everglades National Park, has been usurped by both wet prairie and ridge (Davis and Ogden 1994, Olmsted and Armentano 1997, Richards et al. 2011). Areas of reduced flow have lost the elongated ridge-slough topography, while areas with excessively long flooding duration have experienced a decline in the prevalence of ridges and tree islands (Sklar et al. 2004, Ogden 2005). Remaining ridges have lost rigidity, structure, and directionality (or anisotropy; Wu et al. 2006, Larsen et al. 2007, Watts et al. 2010), and elevation differences between ridges and sloughs have become less distinct (Fig. 3; Watts et al. 2010, McVoy et al. 2011). However, while many characteristics of degrading Everglades landscapes are known, to date no system-wide and systematic studies have addressed their spatial distribution, hydrologic constraints, and covariation.

Self-organization of the Ridge and Slough Landscape

The characteristic ridge and slough mosaic has been theorized to be a self-organized landscape maintained by autogenic processes that balance ridge expansion and slough persistence (Larsen et al. 2007, Givnish et al. 2008, Larsen and Harvey 2010, Watts et al. 2010, Cohen et al. 2011, Heffernan et al. 2013).

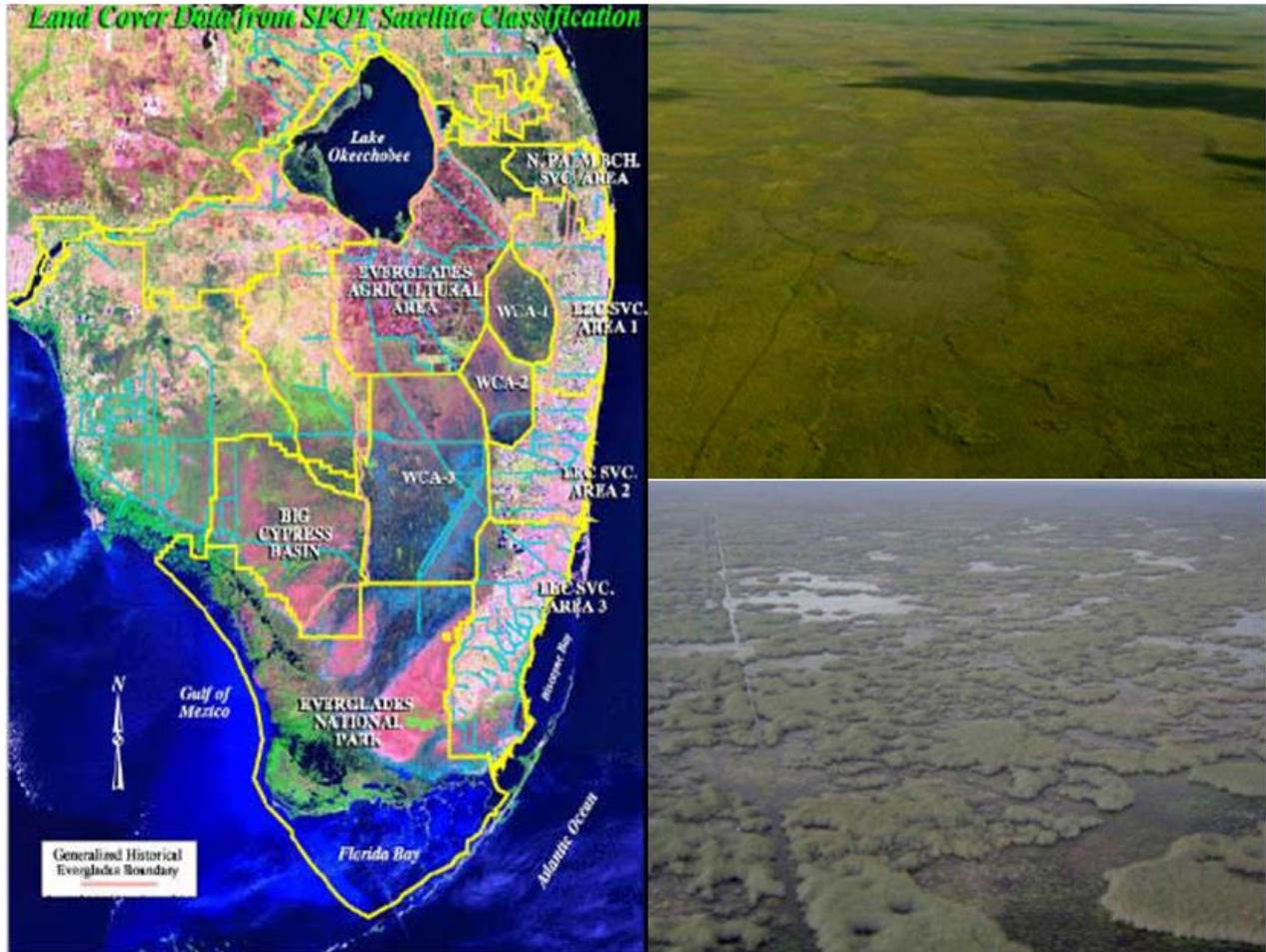


Figure 2: Present configuration of the greater Everglades, and associated changes in ridge-slough structure. (Left) The contemporary Everglades is subdivided into distinct management basins subject to varied uses and management objectives. These basins are bounded and separated by hydrologic engineering structures including canals and levees. (Right, top) Where hydrologic modification has reduced water levels and hydroperiod, historic ridge-slough landscapes have lost topographic structure and become dominated by sawgrass (*Cladium jamaicense*). (Right bottom) Where impoundment has raised water levels and lengthened hydroperiods, ridges senesce and the landscape takes on a characteristic 'moth-eaten' pattern.

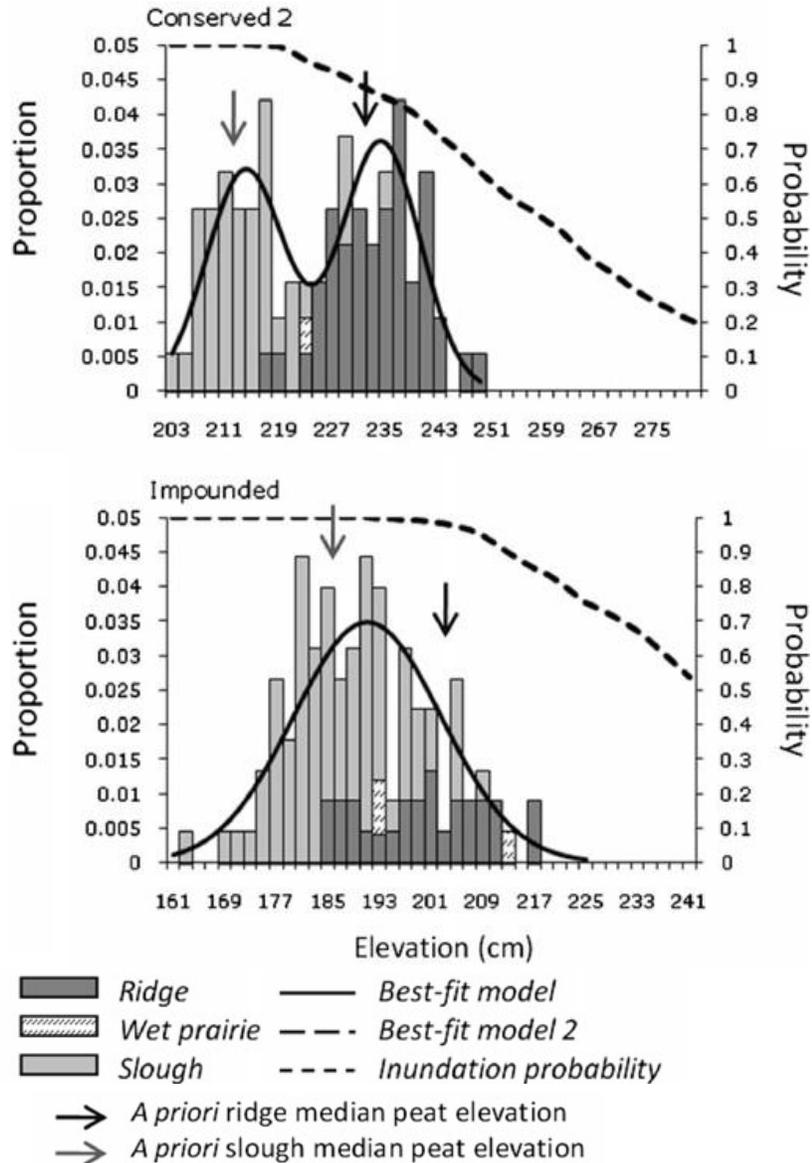


Figure 3: Examples of conserved (top) and degraded (bottom) microtopographic structure. Conserved landscapes are characterized by high topographic heterogeneity and bi-modal elevation distributions. Degraded landscapes have lost these characteristics. From Watts et al. 2010. Shadings indicate vegetation communities, and arrows indicate their median elevation. Solid line indicates best fit model of density vs. elevation. Dashed line indicates probability of inundation over preceding 10 years at each elevation.

Decoupling of soil elevations from underlying bedrock topography in areas of relatively conserved landscape pattern suggests that historic microtopography and landscape structure have arisen largely from internal feedbacks between vegetation, hydrology, and soil elevations. Whether local geologic features have acted as nucleation sites for ridge initiation remains unresolved. In either case, pressure exerted by the hydrologic regime has been theorized to promote the stable existence of the ridge and slough communities (Givnish et al. 2008, Watts et al. 2010, Cohen et al. 2011). Plant production provides raw material for the development of peat, and may increase as soil elevation allows for high productivity of recalcitrant organic matter by sawgrass. Peat depth is maintained by decomposition of biomass, and the loss of peat through aerobic respiration (Craft et al. 1995, Borkhataria et al. 2011). Ridges accumulate biomass faster than sloughs, but shallower water depths promote more rapid decomposition that roughly balances higher gross peat production (Larsen and Harvey 2010, Cohen et al. 2011). The production-respiration equilibrium is regulated within both community types at nearly equal rates over long time periods, keeping ridges and sloughs from forming mountains and valleys. Community shifts in microtopographic range when the hydrologic regime changes may help maintain vegetation zonation, and thus potentially feedback on microtopographic structure (SCT 2003, Larsen and Harvey 2010, Cohen et al. 2011, D'Odorico et al. 2011).

Regular self-organized spatial patterning of ecological systems results from spatially-dependent feedbacks whose strength and sign vary with distance (Rietkerk and Van de Koppel 2008). While strong local feedbacks can generate discrete patches on the landscape (Silliman et al. 2005, D'Odorico et al. 2011), the structure of such mosaics is generated by distal negative feedbacks, by which the presence of organisms at one location inhibits their establishment or persistence at some distance (van de Koppel et al. 2005). In some cases, organisms concentrate a limiting resource such as water or soil nutrients (Rietkerk et al. 2002, Rietkerk et al. 2004b, Ludwig et al. 2005, Eppinga et al. 2009), facilitating local expansion while limiting suitability of locations outside of occupied patches. In others, plants or animals may locally mitigate a stressor such as temperature, salinity, or shear stress but exacerbate the stressor outside the area of biotic influence (Bader et al. 2007, Weerman et al. 2010). The structure of resulting spatial patterns depends on the spatial extent of the inhibitory feedback. If the inhibitory feedback acts at intermediate scales, the result is regular patterning (Rietkerk and van de Koppel), while global feedbacks produce scale-free or fractal patterning of biota (Kefi et al. 2007). The diversity of potential distal feedbacks in any given ecosystem requires the development of distinctive predictions and measurements that discriminate among plausible mechanisms (Eppinga et al. 2010; Acharya et al. 2015, Casey et al. 2015).

When distal negative feedbacks are coupled to strong local positive feedbacks, spatial patterning can exhibit global bi-stability (Fig. 4), meaning that alternative equilibria may exist at the scale of entire landscapes. In such cases, either regular patterning or unstructured, homogenous states may exist under the same set of environmental conditions. As a result, transitions between patterned and homogenous states may be sudden, and trajectories of recovery may exhibit hysteresis (meaning that transitions in one direction occur at a different threshold condition than transitions in the other direction (Rietkerk et al. 2004a). While some researchers have argued that regular spatial patterning always exhibits global bi-stability across some range of conditions (Rietkerk et al. 2004a), others have illustrated that regular patterning can arise via stochastic processes that do not produce such landscape-scale transitions (D'Odorico et al. 2007).

Nonetheless, the potential for hysteretic responses of patterning to environmental drivers presents a challenge to their conservation and restoration (van de Koppel et al. 2002, Suding et al. 2004, Suding and Hobbs 2009), particularly since the loss of patterning can have important implications for ecosystem function (Rietkerk et al. 2004a, D'Odorico et al. 2006) and habitat value (Ogden 2005).

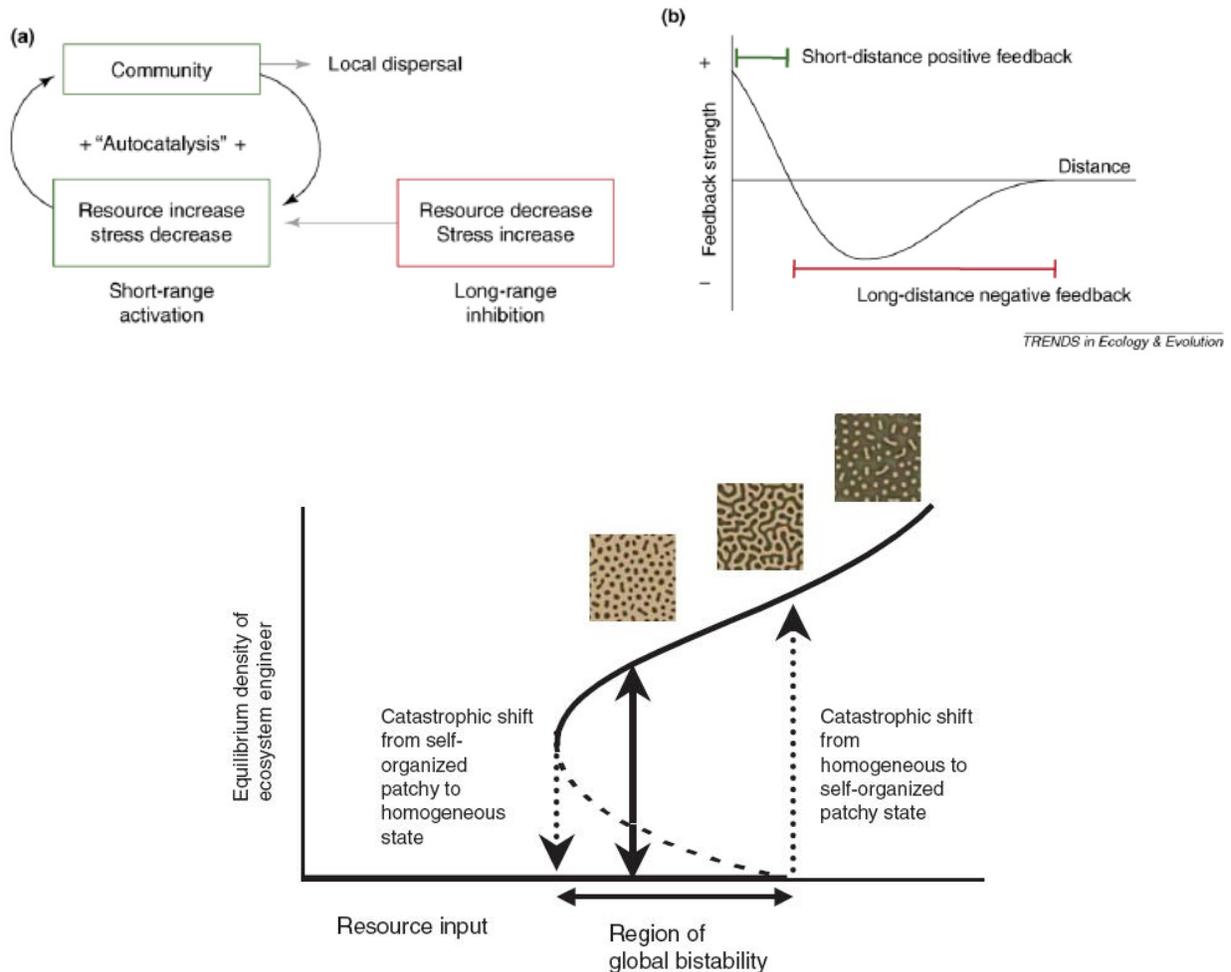


Figure 4: Feedback processes that generate regular pattern and landscape-scale alternative stable states, as is hypothesized to occur in the Everglades Ridge and Slough. (a and b) Biota can, in some cases, create positive local feedbacks via resource increases and stress decreases, but as a consequence reduce the suitability of habitat at greater distances. (c) In response to varied resource inputs (or stressors), patterned landscapes may undergo catastrophic transitions to a homogenous state. Across some range of conditions, both the patterned and homogenous states may be stable, in which case the degraded condition may resist restoration. (a) and (b) from Rietkerk and van de Koppel 2008; (c) From Rietkerk et al. 2004.

The emergence of flow-parallel patterning in the Everglades suggests that the spatial feedbacks that create pattern must act anisotropically (i.e., differently with direction), and that water flow is an important component of those feedbacks (Heffernan et al. 2013, Acharya et al. 2015). However, the specific mechanisms that create flow-parallel ridges remain unresolved, as multiple plausible mechanisms have been suggested, including sediment entrainment and deposition (Larsen et al. 2007, Larsen and Harvey 2010), transpiration-driven nutrient concentration (Ross et al. 2006, Cheng et al. 2011), and hydrologic competence (Givnish et al. 2008, Watts et al. 2010, Cohen et al. 2011, Heffernan et al. 2013, Acharya et al. 2015). While the relative importance of and interactions between these mechanisms remains an active area of research, observations of pattern loss in response to hydrologic management, nutrient enrichment, and other disturbances points to the disruption of those feedbacks as a primary cause of landscape degradation (Sklar et al. 2004).

The coupling of microtopography, hydrology, vegetation composition and productivity, and their responses to hydrologic modification and other disturbance can create challenges in disentangling causal relationships and diagnosing trajectories of change. One objective of this monitoring study is to assess whether microtopographic structure, vegetation community composition, or relationships between these variables serve as leading indicators of pending change in other landscape characteristics. While it is known that altered microtopography affects vegetation structure sometime after hydrologic modification (Ross et al. 2003, Givnish et al. 2008, Zweig and Kitchens 2008, Zweig and Kitchens 2009), vegetation changes may also influence microtopography (Cohen et al. 2011, Larsen et al. 2011). Watts et al. (2010) hypothesize that topographic changes are more rapid than those of vegetation structure. They argue that drainage and stabilization of the Everglades hydrologic regime leads to more rapid peat loss through aerobic bacterial respiration in higher elevation ridges compared to sloughs, flattening landscape scale topography. Simultaneously, but over much longer timeframes, drained and stabilized hydrologic regimes facilitate ridge expansion through the more drained sloughs, resulting in vegetation structure homogeneity (Larsen and Harvey 2010). Then, after some time period, both topography and vegetation structure equilibrate to a relatively homogenous landscape. Nonetheless, the relative timescales of changing vegetation and topographic structure are not well understood.

Simultaneous assessment of microtopographic structure and vegetation community composition provide one means to assess the relative time scales over which these characteristics respond to hydrologic modification (Fig. 5), i.e., whether landform or vegetation is more resistant to change in water regime. If vegetation change precedes topographic change, then we should observe a subset of sampled landscapes in which vegetation community composition is relatively conserved, but topographic structure is degraded. If topographic change precedes vegetation change, we will observe only the converse configuration.

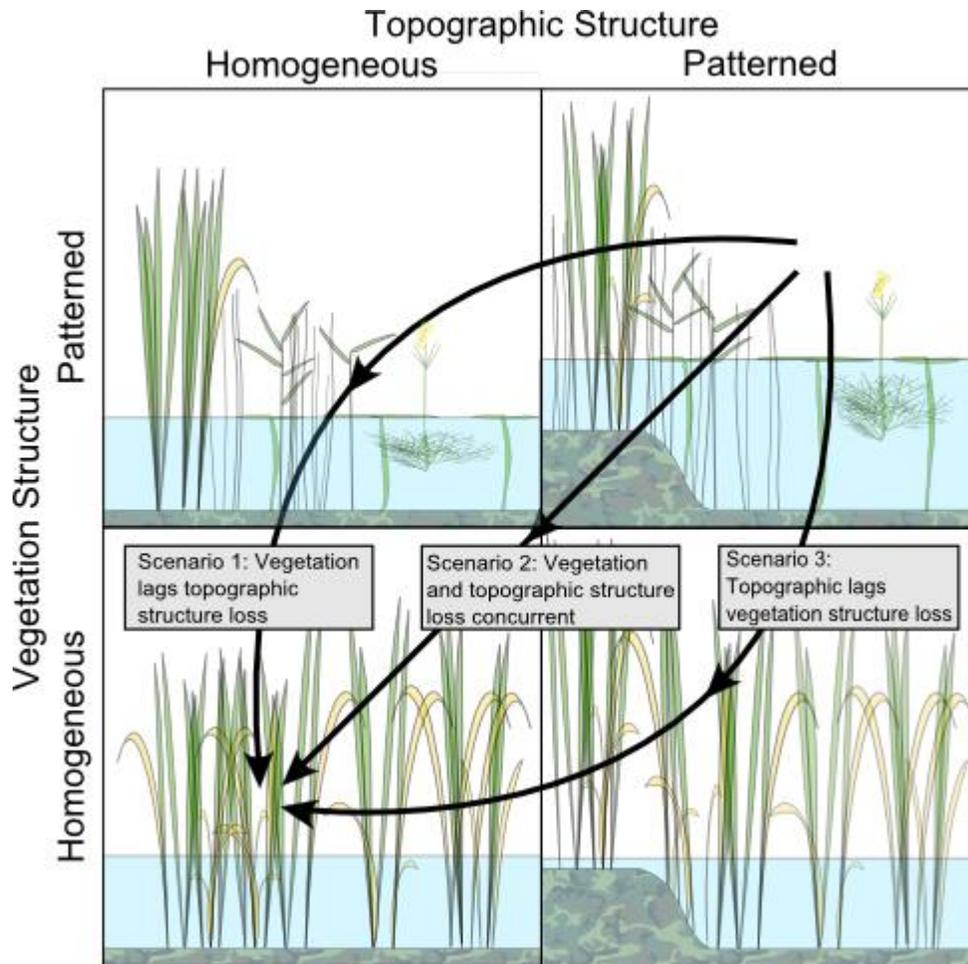


Figure 5: Possible pathways of microtopographic and vegetative degradation in the ridge-slough landscape. In one scenario (uppermost arrow) topographic structure is reduced after modification of the hydrologic regime, followed by a lagged response from the vegetation structure; alternatively (lowermost arrow) vegetation patterning may degrade initially in response to modification of the hydrologic regime, followed by a lagged response of topographic patterning; finally (middle arrow) microtopographic flattening and vegetation homogenization may occur, but both lag behind modification of the hydrologic regime. Depending on which pattern accurately describes pathways of ridge-slough degradation, either vegetation or microtopography may serve as a leading indicator of change in the other characteristic. Pathways of degradation may differ depending on the nature of hydrologic alteration and other disturbances.

Monitoring rationale and goals

The Water Resources Development Act (WRDA) of 2000 authorized the Comprehensive Everglades Restoration Plan (CERP) as a framework for modifications and operational changes to the Central and Southern Florida Project needed to restore the South Florida ecosystem. Provisions within WRDA 2000 provide for specific authorization for an adaptive assessment and monitoring program. A CERP Monitoring and Assessment Plan (MAP; RECOVER 2004, 2006) has been developed as the primary tool to assess the system-wide performance of the CERP by the Restoration Coordination and Verification (RECOVER) program. The MAP presents the monitoring and supporting research needed to measure the responses of the South Florida ecosystem to CERP implementation. Investigators are encouraged to refer to this document for details on the methods and procedures outlined below. These documents can be accessed from the following web sites: http://www.evergladesplan.org/pm/recover/recover_map.aspx and http://www.evergladesplan.org/pm/recover/recover_map_part2.aspx.

The MAP also presents system-wide performance measures representative of the natural and human systems found in South Florida that will be evaluated to help determine CERP success. These system-wide performance measures address the responses of the South Florida ecosystem that the CERP is explicitly designed to improve, correct, or otherwise directly affect. A separate document, the Development and Application of Comprehensive Everglades Restoration Plan System-wide Performance Measures (RECOVER 2007), has been prepared by RECOVER and provides the scientific, technical, and legal basis for the performance measures. This document and performance measure documentation sheets can be downloaded from: http://www.evergladesplan.org/pm/recover/eval_team_perf_measures.aspx.

The general goals of restoration are to stem, and possibly reverse, degradation of the ridge-slough-tree island landscape by redirecting flows now released unused to coastal waters across the surface of this landscape (USACE and SFWMD 1999). The CERP MAP, Parts 1 and 2, presented the overarching monitoring framework for guiding restoration efforts throughout the entire process (RECOVER 2004, 2006). This requires not only a comprehensive assessment of the current state of the ecosystem and assessment of restoration endpoints (targets), but also ongoing monitoring and evaluation throughout the process that will aid the implementing agencies in optimizing operational procedures and project designs. The work described below represents the first system-wide landscape monitoring project. This monitoring effort supports the Greater Everglades Wetlands module of the MAP and is directly linked to the monitoring or research component identified in that module as number 3.1.3.6.

This monitoring project seeks to provide information necessary for the evaluation of the efficacy of the Comprehensive Everglades Restoration Program (CERP), as delineated in the Water Resources Development Act (WRDA) of 2000. The work described provides indices of system-wide applicability of performance measures related to the response of the ridge-slough mosaic, tree islands, and other landscape features of the central Everglades to the restoration of historic hydrologic conditions, with the goal of informing the adaptive management of Everglades restoration as outlined in the CERP Monitoring and Assessment Plan (RECOVER 2004).

The primary objective of this monitoring project is to assess the condition of landscapes within the Greater Everglades Wetlands ecosystem. This effort focuses on the condition of wetlands (including tree islands) within the historic distribution of the ridge and slough landscape and provides a baseline to detect changes/trends in the patterns and vegetation communities of these systems as a result of water management operations, restoration initiatives and episodic events such as droughts, fire and hurricanes. The secondary objective is to integrate knowledge regarding landscape patterning, soil dynamics and community structure and composition with hydrologic data provided by Everglades Depth Estimation Network (EDEN) and other sources. Particular attention is paid to how these dynamics might: 1) be affected by restoration and 2) relate to CERP hypotheses from the MAP.

The specific objectives of this work are:

- 1) To determine extant reference conditions for each of the performance measures described below (including variability of those measures in time and space).
- 2) To establish the present status of landscape performance measures throughout the central Everglades, particularly in areas of historic ridge-slough landscape patterning, identify spatial and temporal trends of those performance measures, and quantify their relationships to the present hydrologic regime.
- 3) To detect unanticipated changes in ecosystem structure and processes that result from hydrologic management or manipulation, CERP restoration activities, or climatic variation
- 4) To provide data in support of scientific studies of inter-relationships among vegetation, microtopography, and hydrologic regime that may provide insight into the causes of unanticipated ecosystem responses.

The monitoring work is designed to address the needs identified in the Greater Everglades wetlands performance measures: (1) Wetland Landscape Patterns – Ridge-Slough Community Sustainability; and (2) Wetland Landscape Patterns - Marl Prairie Cape Sable Sparrow Habitat. The program specifically addresses the Greater Everglades Wetland Landscape and Plant Community Dynamics hypotheses: (1) ridge and slough microtopography in relation to organic soil accretion and loss; (2) ridge and slough landscape pattern in relation to microtopography; and (3) plant community dynamics along elevation gradients as water depths and thus hydroperiods change (RECOVER 2006).

METHODS

Overview of approaches

Monitoring efforts for 2010-15 consisted of three core components: (1) mapping vegetation features from aerial photographs, (2) aerial surveys for classification of tree island type, and (3) ground surveys of water depth and plant community structure (in both tree islands and surrounding marsh), which were used to quantify aspects of the hydrologic regime, determine relationships between vegetation structure and water depth, quantify the distribution and spatial structure of peat elevations, and ground-truth broader-scale maps based on remote sensing and aerial surveys. These activities were linked both logistically and analytically (Fig. 6). For example, vegetation mapping from photographs were supported by aerial marsh reconnaissance that was supplemented by tree island characterization activities. Mapping accuracy can be determined from vegetation observations made during surface pattern sampling. Perhaps most importantly, analysis of pattern based on multiple variables (vegetation and soil micro-topography) at multiple scales (aerial photos, helicopter reconnaissance, ground surveys) will maximize the likelihood of change detection, allow inference about interrelationships among stressors and response variables, and present an integrated picture of the pre-restoration structure of the Greater Everglades Wetland Ecosystem. The ability to compare multiple measures of landscape condition may provide a more robust understanding of spatial patterns and temporal trajectories of landscape degradation.

This study takes advantage of a previously established framework for representative sampling of the entire Everglades landscape. This Generalized Random-Tessellation Stratified sampling network (GRTS; Phillipi 2007) divides the Everglades landscape into a grid of 2x5 km landscape blocks (primary sampling units [PSUs]), with the 5 km edge of each PSU aligned parallel to the historic water flow. A spatially-stratified random sample of 80 PSUs were selected for sampling over a 5 year period (n=16 per year), and each year a subset of these were drawn to achieve a spatially balanced sample of the modern Everglades compartments (Everglades National Park (ENP), Water Conservation Area 3A North (WCA 3A N), Water Conservation Area 3A South (WCA 3A S), Water Conservation Area 3B, Water Conservation Area 2, and the Loxahatchee National Wildlife Refuge (LNWR); Fig. 7). In Year 3, budget and permitting constraints limited sampling in ENP. During Years 3 and 4, monitoring efforts were also shifted to additional PSUs outside the original sampling scheme, with the purpose of documenting pre-restoration reference conditions within the footprint of the DECOMP Physical Model (PSU DPM, between WCA3AS and WCA3B), and downstream of the raised section (accomplished and planned) of Tamiami Trail (PSU 513 and Blue Shanty Area in northern ENP). Mechanistic analyses for this study focus on 62 PSUs (from sampling years 1-5) found within the historic distribution of the Ridge and Slough landscape (McVoy et al. 2011). These PSU's represent the full range of contemporary hydrologic regimes, and their vegetative and microtopographic structure range from well-conserved to severely degraded (Wu et al. 2006, Watts et al. 2010, Nungesser 2011).

This Year-5 report includes results of marsh landscape structure in 62 PSUs sampled over five years (2010-2015) in the first 5-year cycle.

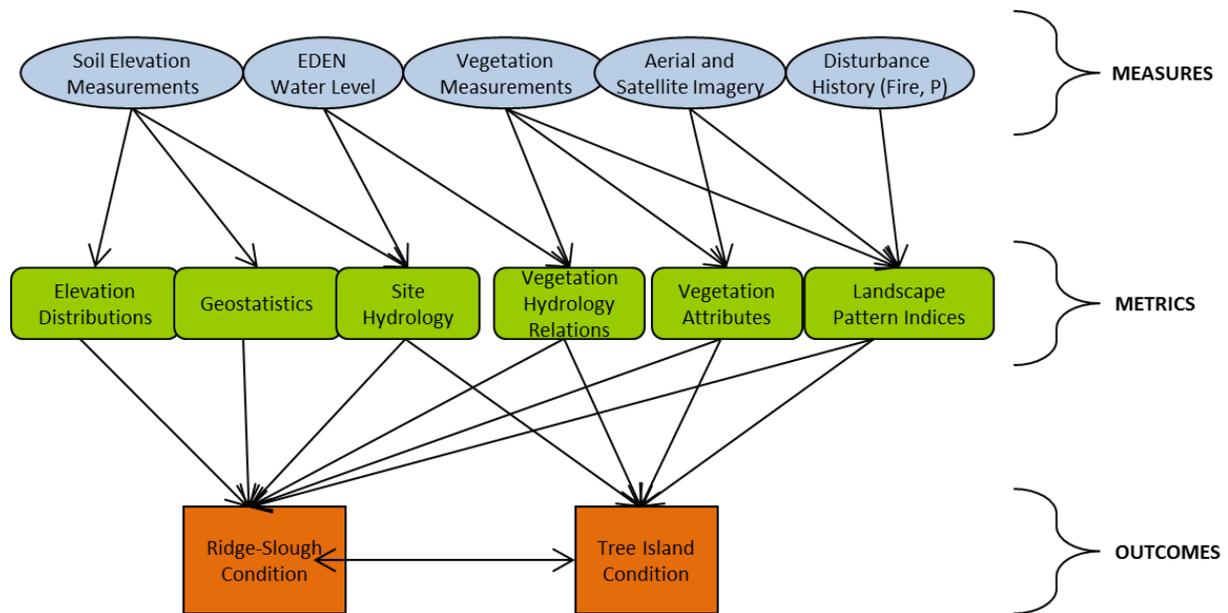


Figure 6: Relationships among direct measurements to be collected, metrics derived from those measurements, and assessment outcomes.

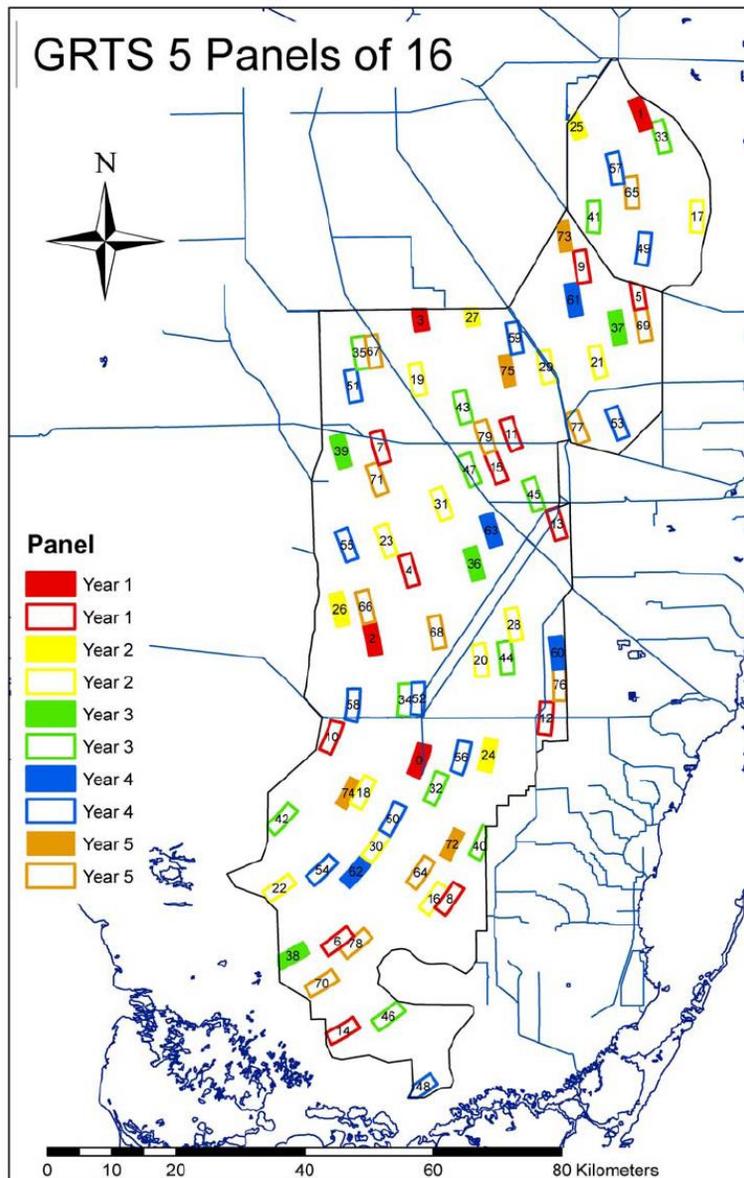


Figure 7: Map of all PSUs for landscape sampling (from Phillipi 2007). Colors indicate years for sampling of individual PSUs. A selection of tree islands were sampled in solid landscape blocks in Years 1 and 2.

Mapping of marsh and tree island vegetation

Developing tools for remote detection of ecological condition is of obvious utility for large area surveillance of restoration progress. Some research has already been done to discern landscape pattern metrics that may be of use in long term monitoring efforts (Wu et al. 2006, Nungesser 2011, Casey et al. 2015), and one objective of this component of monitoring is to evaluate these existing metrics. A second objective is to develop landscape metrics that are predicated more on the mechanisms of landscape pattern maintenance (i.e., flow and longitudinal slough connectivity). A third objective is to compare field metrics of ridge-slough landscape condition with those obtained from imagery to determine patterns of association and critical transition thresholds that may be useful for restoration assessment.

The narrow objective of the vegetation mapping component of this project is to map and characterize the vegetation in each PSU using the classification scheme developed for CERP vegetation mapping. However, the incorporation of a mapping component in the Ridge – Slough-tree island mosaic project facilitates achievement of a broader goal: the development of a better understanding of the relationships between hydrology and other drivers of vegetation pattern. It does so by allowing hypotheses about these relationships to be addressed at an intermediate scale – much coarser than the plot data collected in the land surface elevation component of this project, and considerably finer than the vegetation mapping component for the entire Everglades.

Vegetation maps of each PSU have been produced, with a minimum mapping unit of 200m² for non-woody vegetation and 36 m² for tree islands (See Appendix II). The mapping effort or area mapped within each PSU varied by year. PSU mapped during the first year were mapped to their full 2 km x 5 km extent while those mapped in the 2nd and 3rd year were limited to a central 2 km x 2 km portion within each PSU (see Fig. 9 for locations of PSU's sampled in Years 1, 2 and 3). The base map of these vegetation maps consisted of either color infrared CERP aerial imagery from 2003 or 2009, for PSU 4 and 5 and PSU 0 and 14, respectively, or NAIP 2010 color infrared aerial imagery for the other 44 PSU and the DPM area. Features within each PSU were classified according to the classification system developed by a consortium of south Florida vegetation scientists (Rutchev et al. 2006). Most community types have been distinguished to Level 3 of this classification system

Each tree island within a given PSU has been mapped and characterized. Tree island >36 m² in size within each PSU were identified and digitized. Based on helicopter aerial surveys, each tree island was classified into one of the following seven types:

- i. Hardwood hammocks (forests or woodlands with >50% relative cover of upland & transitional tree species)
- ii. Bayhead (closed crown forest with >50% relative cover of swamp and transitional tree species)
- iii. Bayhead Swamp (open woodland community with >50% relative cover of swamp and transitional tree species)
- iv. Willow head or strand
- v. Cypress dome or strand
- vi. Exotic-dominated tree island

The maps of the PSUs, including the one in DPM area were already included in the Year-1, 2 and 3 reports (Ross et al. 2013, 2015a, 2015b). In subsequent analysis, maps of each PSU will be used to determine the 5 indices utilized by Wu et al (2006) to evaluate the integrity of ridge-slough patterning: Lacunarity index (LI), Average Length of Straight Flow (ALS), Average Width of Slough (ASW), Percentage of Ridge Area (PRA), and Average Length-Width Ratio (ALW). Responsiveness of map-derived indices of vegetation pattern degradation will be compared to statistical descriptors of peat surface patterning and multivariate descriptors of hydrologic regime (see below) in order to determine the circumstances under which each approach provides a more robust and sensitive indication of landscape integrity.

Sampling and analysis of Marsh Vegetation

Field measurement of soil elevation and vegetation composition

The approach adopted in this monitoring component is strongly informed by recent studies of relationships between hydrologic regime and peat surface structure in eight landscape blocks (structurally similar to current PSUs) stretching from WCA 3AN to ENP and encompassing historic ridge and slough landscapes currently subject to drained, inundated, and stabilized hydrologic regimes (Watts et al. 2010). The central finding of these efforts is that the frequency distribution and spatial structure of peat surface elevations provides a directionally-sensitive indicator of ridge-slough landscape response to hydrologic regime, in that each of these metrics responds differently to inundation, drainage, and flow stabilization. Importantly, metrics based on peat surfaces appear to diverge in some cases from metrics of the landscape structure of vegetation in identifying conserved and degraded patterning. The present analysis adds measures of vegetation community distinctiveness and fidelity to elevation as core metrics. Geostatistical measures of plant community dissimilarity and microtopographic variation will provide additional measures of landscape structure, but are not included in this analysis.

Wherever possible, field sampling of the ridge-slough landscape was done via airboat, during periods when sufficient water was present to obtain a reliable measure of water depth at all locations. As such, no dry weather sampling was conducted. For PSUs situated in Everglades National Park, sites were accessed by airboat or helicopter, as allowed by permitting and budgetary constraints.

Prior to the sampling of each PSU, the 2 x 5 km area was subdivided in 80 equal area zones (250 m x 500 m) and a sampling cluster was located at a random location within each (Fig. 8). At each cluster, samples were then collected at the center and at two randomly selected distances between 3 and 35 m in two cardinal directions, east and north. Sampling at each location commenced with setting a 1-m² quadrat on the ground, centered on the target point, to delineate the sampling boundary. Within each quadrat, water depth was measured using a meter stick with a 10-cm diameter hard plastic foot anchored to one end; the foot ensures that water depth is measured to the soil surface. Field training of sampling personnel ensured that a standardized amount of pressure was applied to the foot such that the measurement of water depth was uniform across time and space. Water depths were measured with a precision of 0.5 cm. In addition, we determined depth to bedrock at each node.

Vegetation characterization within each quadrat consisted of identifying all taxa present to species level, estimating cover of each using a Braun-Blanquet scale (1 – 1-5%, 2 – 5-25%, 3 – 25-50%, 4 – 50-75%, 5 – 75-95%, and 6 – 95-100%). Based on these vegetation measurements, the vegetation within a 25 m radius of each sampling location was assigned to a community category (ridge, slough, tree island, wet prairie, cattail). In some PSUs, species cover was estimated as percentage cover of the plot area at either 1%, 5% or at 10% intervals; values from Braun-Blanquet scales were converted to these values for data analysis. Where quadrats span a transition from one community type to another, we assigned points to mixed categories (e.g., ridge/wet prairie). In Years 3-5, field classifications were adjusted to better inform mapping from aerial imagery, and more directly related to community classifications adopted by Rutchey et al. (2006). As a result, direct comparison of the quantitative characteristics of field assigned communities is not possible across all years.

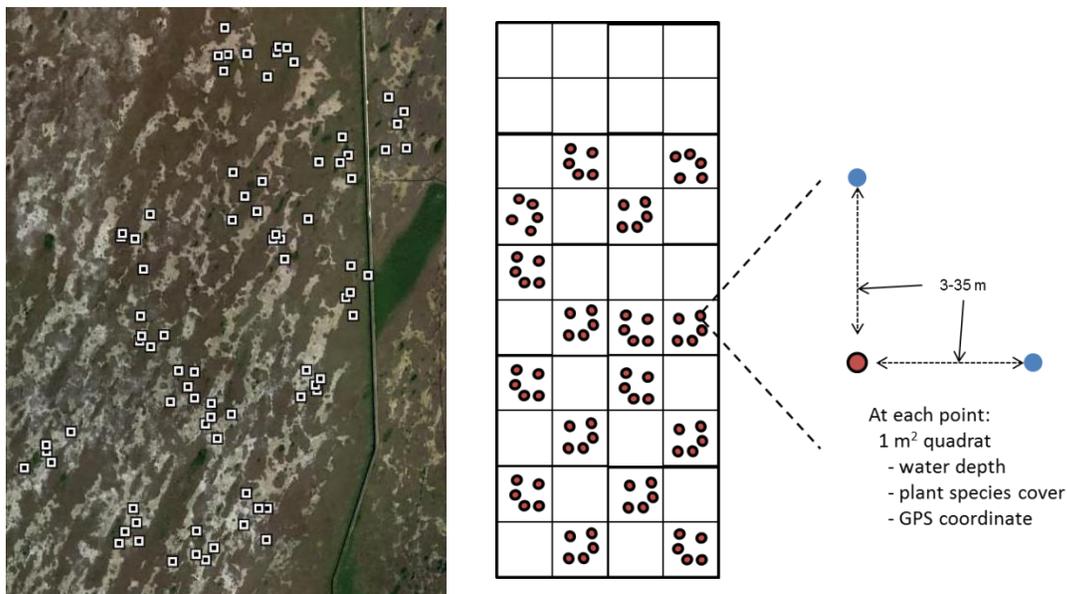


Figure 8: Locations of sampling clusters (red dots) within 2x5 km primary sampling units (PSUs); the location of clusters within each 500 x 500 m zone is assigned randomly. At each cluster, 3 sampling locations (green dots) are visited; sites are situated at the center of each cluster, and at a random distance between 3 and 35 m in the direction of the PSU azimuth and in the orthogonal direction. Measurements at each site include location, vegetation community composition and water depth.

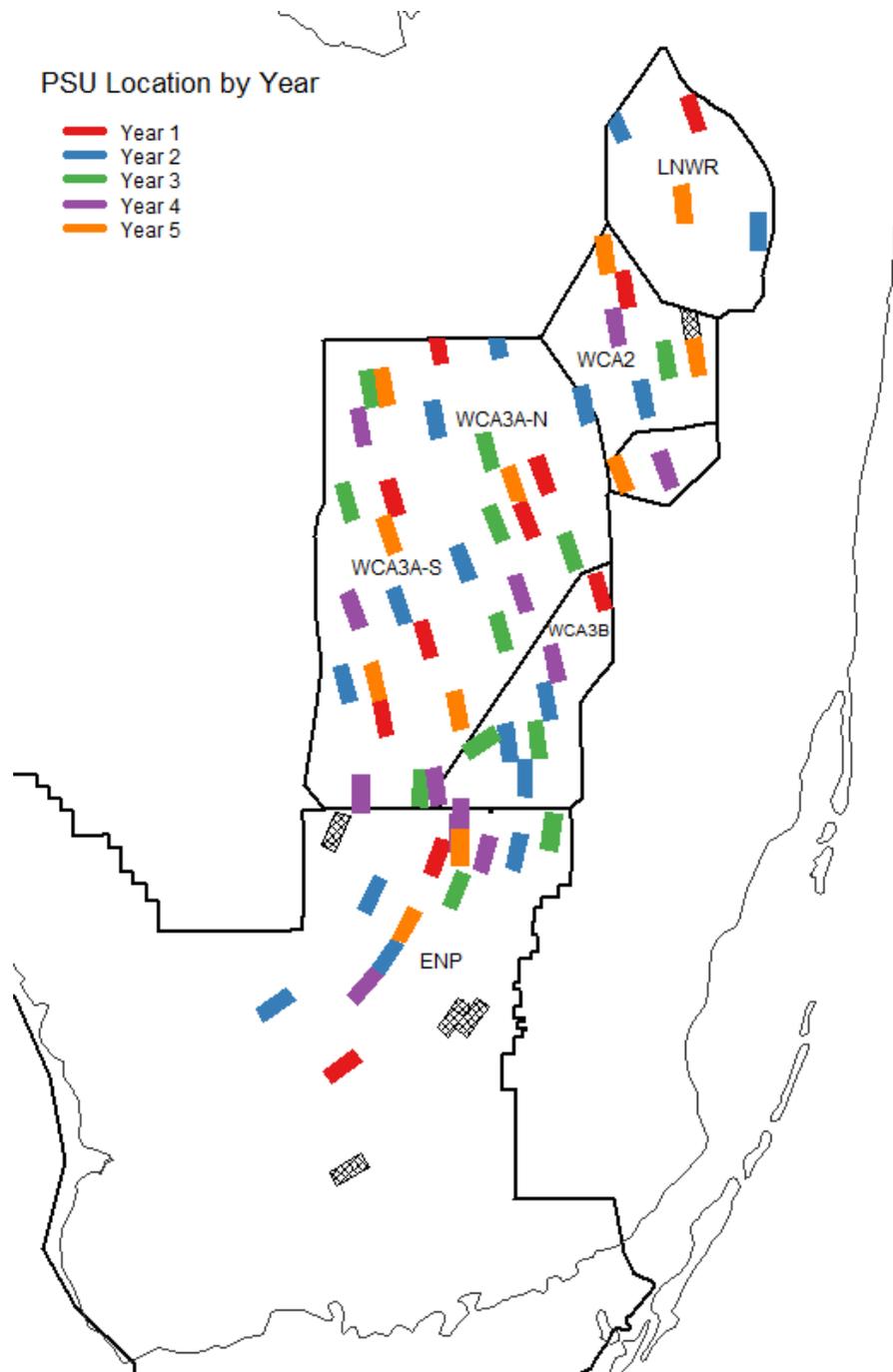


Figure 9: Map of the greater Everglades landscape showing the 62 study sites sampled within the historic ridge-slough landscape during years 1-5 of this project. Each site contains up to 240 sampling points in a spatially-stratified design. Year 1 PSU's are in red, Year 2 PSU's are in blue, Year 3 PSUs are in green, Year 4 PSUs are in purple, Year 5 PSUs are in orange. Hatched PSUs were sampled during years 1 and 2, but fall outside of the historic ridge-slough landscape, and are not included in most cross PSU analyses.

Table 1: Characteristics of PSUs sampled to date.

PSU	Cycle (Year)	Dates Sampled	Classification		Location			
			Area [†]	Historic Ridge-Slough	Centroid Easting (UTM)	Centroid Northing (UTM)	Azimuth (Degrees)	UTM Zone
0	1	3/20, 3/22, 3/27/2012	ENP	Y	532345.51	2842696.30	19	17
1	1	9/18/2009	WCA1	Y	566677.85	2942982.08	341	17
2	1	11/9, 11/23, 11/24/2009	WCA3AS	Y	525056.59	2861614.12	349	17
3	1	9/4/2009	WCA3AN	Y	532505.33	2910966.94	354	17
4	1	7/28,7/29/2009	WCA3AS	Y	530756.35	2872127.60	344	17
5	1	NA*	WCA2	Y	566325.52	2914610.64	354	17
6	1	10/24,10/28/2009	ENP	Y	519649.37	2814585.30	39	17
7	1	12/8/2009	WCA3AN	Y	526262.38	2891226.13	345	17
8	1	11/30, 12/12, 12/16/2011	ENP	N	537019.49	2821237.51	30	17
9	1	8/17, 8/18/2009	WCA2	Y	557549.62	2919280.24	352	17
10	1	4/19, 5/4, 5/5/2012	ENP	N	518729.07	2846327.59	339	17
11	1	8/5, 8/6, 8/11/2010	WCA3AN	Y	546603.34	2893273.01	342	17
81	1	12/13, 12/14, 12/16, 12/17/2010	WCA3B	Y	544130.08	2853456.03	360	17
13	1	8/24, 8/27, 9/3/2010	WCA3AN	Y	553652.16	2879348.07	344	17
14	1	9/9, 9/14, 10/7/2011	ENP	N	520452.78	2800699.28	348	17
15	1	6/17, 6/18, 7/30/2010	WCA3AN	Y	544263.57	2888174.08	340	17
16	2	12/19/2011, 1/4, 1/6/2012	ENP	N	534551.56	2821237.18	31	17
17	2	2/2/2010	WCA1	Y	575467.53	2927079.79	350	17
18	2	5/18,5/25, 6/5, 6/7/2010	ENP	Y	523582.48	2837739.76	25	17
19	2	9/30, 10/8, 10/12/2010	WCA3AN	Y	532020.89	2901747.79	350	17
20	2	9/30, 10/3/2011	WCA3B	Y	541840.16	2858248.34	353	17
21	2	3/3/2010	WCA2	Y	560020.33	2904486.44	348	17
22	2	11/4, 11/9/2011	ENP	Y	510586.67	2822844.43	346	17
23	2	9/23, 9/26, 9/28/2011	WCA3AS	Y	527209.63	2876687.70	342	17
24	2	2/28, 3/1, 3/6/2012	ENP	Y	543033.61	2843539.09	13	17
25	2	2/3/2010	WCA1	Y	556804.01	2940955.57	342	17
26	2	9/10, 9/17, 9/23/2010	WCA3AS	Y	519957.43	2866106.03	346	17
27	2	10/12, 10/21/2011	WCA3AN	Y	540532.06	2911393.98	356	17

* Extremely high cattail density made all sampling points within PSU 5 inaccessible by airboat - no sampling of this area has been conducted

[†] ENP = Everglades National Park, WCA1 = Loxahatchee National Wildlife Refuge (Water Conservation Area 1), WCA 2 = Water Conservation Area 2, WCA3AN, S = Water Conservation Area 3A North and South, WCA3B = Water Conservation Area 3B

Table 1: Contd.

PSU	Cycle (Year)	Dates Sampled	Classification		Location			
			Area [†]	Historic Ridge-Slough	Centroid Easting (UTM)	Centroid Northing (UTM)	Azimuth (Degrees)	UTM Zone
28	2	11/19, 12/2, 12/13/2010	WCA3B	Y	547035.43	2863766.37	350	17
29	2	9/16/2011	WCA3AN	Y	552008.07	2903701.35	349	17
30	2	9/6, 9/7, 11/16, 11/23, 11/28/2011	ENP	Y	525597.48	2882440.91	30	17
31	2	8/19, 8/22, 9/2/2011	WCA3AS	Y	535763.28	2882440.91	340	17
32	3	1/9, 1/13/2013	ENP	Y	534894.8	2838347.8	22	17
34	3	12/18/2012, 1/4/2013	WCA3AS	Y	530097.7	2852094.7	2.5	17
35	3	11/29/2012	WCA3AS	Y	523207.3	2905898.8	351	17
36	3	1/29, 2/1/2013	WCA3AS	Y	540859.6	2873130.6	345	17
37	3	12/11, 12/14/2013	WCA2	Y	563108.3	2909792.2	351	17
39	3	10/30,11/5/2012	WCA3AS	Y	520196.3	2890623	345	17
43	3	11/8,11/13/2012	WCA3AN	Y	539077.4	2897449.3	346	17
44	3	12/4,12/7//2012	WCA3B	Y	545823.9	2858632.9	353	17
45	3	11/27/2012, 1/21,1/22,1/24/2013	WCA3AN	N	550107.7	2883908.2	342	17
47	3	11/16, 11/19/2013	WCA3AN	Y	540134.9	2887740.3	340	17
513	3 [#]	1/15, 1/18/2013	ENP	Y	547619.4	2846243.2	7.8	17
DPM	3 [#]	1/1/2013	WCA3B	Y	538203	2858189.1	57.1	17
51	4	8/4/2014, 8/11/2014	WCA3AS	Y	522037.935	2900773.35	350	17
52	4	12/23/2013, 1/3, 1/10/2014	WCA3AS	Y	532107.596	2852288.62	351	17
53	4	1/31, 2/7, 2/12/2014	WCA2	Y	563079.229	2894981.98	340	17
55	4	1/13, 1/17/2014	WCA3AS	Y	521064.604	2876059.18	341	17
56	4	12/6, 12/10, 12/20/2013	ENP	Y	538819.489	2843183.13	16	17
58	4	1/6, 1/10, 1/15/2014	WCA3AS	Y	522023.675	2851319.84	0	17
61	4	2/14, 2/17/2014	WCA2	Y	556317.012	2914142.57	351	17
62	4	12/12, 12/13/2013	ENP	Y	522506.182	2825415.44	34	17
63	4	1/24, 1/27/2014	WCA3AS	Y	543511.682	2878334.19	343	17
220	4 [#]	10/8, 10/11/2013	ENP	Y	548070.764	2868866.39	0	17
BS1	4 [#]	9/24, 10/18/2013	ENP	Y	535434.721	2848146.91	0	17
BS2	4 [#]	10/22, 10/29/2013	ENP	Y	535135	2846113	0	17
BS3	4 [#]	11/1, 12/3, 12/5, 12/20,2103	ENP	Y	535354	2844092	0	17

* Extremely high cattail density made all sampling points within PSU 5 inaccessible by airboat - no sampling of this area has been conducted

[†] ENP = Everglades National Park, WCA1 = Loxahatchee National Wildlife Refuge (Water Conservation Area 1), WCA 2 = Water Conservation Area 2, WCA3AN, S = Water Conservation Area 3A North and South, WCA3B = Water Conservation Area 3B

[#] PSUs 220, 513, BS1-3, and DPM are not part of the GRTS sampling design, but were selected for the importance in support of major CERP projects.

Table 1: Contd.

PSU	Cycle (Year)	Dates Sampled	Classification		Location			
			Area [†]	Historic Ridge-Slough	Centroid Easting (UTM)	Centroid Northing (UTM)	Azimuth (Degrees)	UTM Zone
50	5	10/15, 10/22/2014	ENP	Y	528202.17	2833604.6	0	17
54	5	10/29, 10/31/2014	ENP	Y	517243.75	2825691.9	0	17
65	5	9/8, 10/1, 11/3/2014	LNWR	Y	565318.41	2930799.6	0	17
66	5	7/10, 9/18, 9/21, 9/23/2015	WCA3AS	Y	523983.12	2866499.2	0	17
67	5	8/25, 9/2/2014	WCA3AS	Y	525201.96	2906093.8	0	17
68	5	8/8, 8/15, 8/29/2014	WCA3AS	Y	535046.2	2862596.3	0	17
69	5	8/27, 9/5/2014	WCA2	Y	567098.46	2910181.7	0	17
71	5	8/13, 8/22/2104	WCA3AS	Y	525747.06	2886258.6	0	17
73	5	9/15, 9/17/2014	WCA2	Y	554872.25	2923975.3	0	17
79	5	8/18, 8/20/2014	WCA3AN	Y	542515.37	2892858.7	0	17

[†] ENP = Everglades National Park, WCA1 = Loxahatchee National Wildlife Refuge (Water Conservation Area 1), WCA 2 = Water Conservation Area 2, WCA3AN,S = Water Conservation Area 3A North and South, WCA3B = Water Conservation Area 3B

Site/Point Hydrology

Synoptic water depths can be useful for evaluating the distribution of soil elevations over a particular PSU, but it does not allow comparison across PSUs (because observations are done under different hydrologic conditions) and it does not provide a full hydrologic context for each PSU. To establish site hydrologic conditions, we coupled our synoptic measurements of water depths to the US Geological Survey's Everglades Depth Estimation Network (EDEN) based on the geographic location of each point. EDEN collects water stage data daily from 253 stations, and interpolates water levels across the entire Everglades landscape at daily time steps at a grid size of ca. 400 m². For each sampling point, we established a hydrologic history spanning from the day of sampling back to 1991, the earliest current hindcast date, by benchmarking measured water depth and EDEN-estimated water elevation at the center point of each PSU (Fig. 9). Because PSUs were not spatially situated to maximize proximity to sites where water level is directly recorded, we relied on spatially-interpolated EDEN water surfaces to estimate water depths on the day of sampling and to reconstruct point-scale hydrologic history. We evaluated the assumption of negligible water slope by examining relationships between UTM coordinates (easting, northing) and water elevation. For PSUs with significant relationships between water elevation and coordinates, we divided PSUs into 4 north-south bands and benchmarked points within each band to water elevations at the center point of that band.

To determine the particular conditions at a site requires first that soil elevation be determined from EDEN estimates of water elevation on the day of sampling and water depths (Fig. 10). From these hydrologic histories, we calculated mean water depth and inundation frequency at each point over the preceding 0.25, 0.5, 1, 2, 5, 10 and ca. 20 years (i.e. the complete hydrologic record). Because of strong correlation among these measures within PSUs, we use measures derived from the full hydrologic record as predictors of vegetative and microtopographic condition. Additional hydrologic metrics originally proposed include other attributes of point- and PSU- scale hydrologic regime: maximum annual water depth (point scale), water level variability (PSU scale); water level rates of change (PSU scales), and timing of water level maxima and minima (PSU scale). This full suite of hydrologic metrics for each PSU will ultimately be considered in point- and PSU scale analyses of hydrologic condition, but have not been incorporated into analyses presented here.

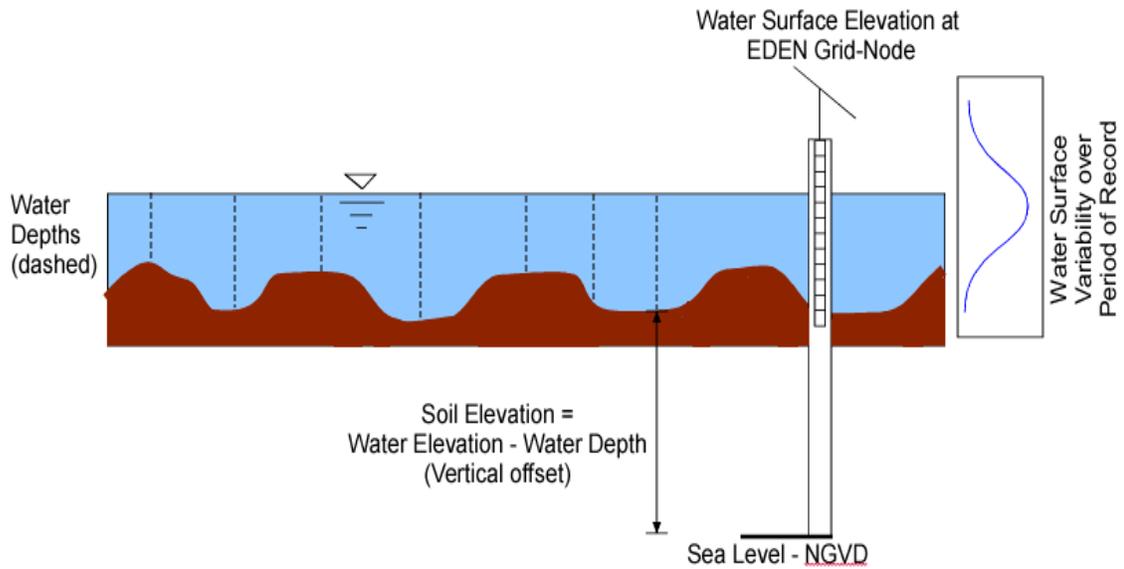


Figure 10: Determination of soil surface elevation from measurements of water depth (dashed lines) and water elevation (from EDEN) on the same day as water depths were measured. Hydrographs can be constructed from this soil elevation estimate and the time-series of water elevations (distribution at right). Time series of stage can be used to report hydroperiod, mean depth, water level variability, exposure frequency/duration/depth, etc.

Data Analysis – Microtopography

To assess microtopographic variation and hydrologic regime, we generated summary statistics of soil elevation and water level, including mean, standard deviation, skew and kurtosis (which describes the degree of shouldering in a distribution and can be used to diagnose bimodality). Standard deviation of water level describes the temporal variability of water level, while standard deviation of water depth (or soil elevation) describes the magnitude of spatial variation in microtopography. To test for bimodality in the peat elevation distributions, we used the R package 'mclust' to assess goodness-of-fit between the observed histogram of peat elevations and 1) a single normal and 2) a mixture of two normal distributions:

$$P_s = N(\mu_i, \sigma_i) \quad (1)$$

$$P_m = q \cdot N(\mu_1, \sigma_1) + (1 - q) \cdot N(\mu_2, \sigma_2) \quad (2)$$

where q represents the probability of falling within the first normal distribution, and N is a normal distribution with mean μ_i and standard deviation σ_i . We also determined whether models based on mixtures of larger number of normal distributions better fit the data; in the few cases where models with 3 or more modes had better goodness-of-fit, we report that finding but use the better of models 1 and 2 in subsequent data analysis. Model goodness of fit was compared using Bayes' information criterion (BIC). The best-fit model was considered to have the lowest BIC score. To evaluate how microtopographic structure responds to hydrologic regime, we examined the relationship between mean annual water depth and the elevation difference between modes of bimodal distributions, where present. To assess whether the persistence of microtopographic pattern might exhibit global bi-stability (Fig. 4), we generated histograms of PSU-scale elevation variance, and tested for bimodality across PSUs in the same manner as tests of elevation bi-modality within PSUs.

Data Analysis - Vegetation structure

In areas with relatively well maintained hydrologic regimes, vegetation communities are separated by clear topographic boundaries, and species preferentially inhabit distinct hydrologic niches. As the hydrologic regime degrades, this patterning is lost. If the topographic responses to changes in the hydrologic regime are the dominant environmental driver that maintains community distinctness, the similarity of communities within PSUs should be greater under either impounded or drained conditions than in relatively conserved landscapes.

To assess how the distinctiveness of vegetation communities changes in response to hydrologic and topographic change, we assessed the dissimilarity among vegetation community composition as the distance (in multivariate space) between artificially-imposed vegetation clusters. In this analysis, individual sampling points from all PSUs were ordinated using a Kruskal's non-metric multidimensional scaling (NMDS) ordination plot, in which more dissimilar sites align further apart in the NMDS plot with the objective of minimizing "stress" in the data. This single global NMDS ordination plot enabled us to 1) obtain a global estimate of the clustering of sampling points containing a set of species among all PSUs; and 2) standardize the among-PSU data. Four dimensions (axes) for the global NMDS ordination plot were decided on before further analysis, based on a scree plot of stress scores against the number of dimensions, where the appropriate number of dimensions balances simplicity and ecological relevance (a satisfactory

amount of total variation in the raw data explained). NMDS analysis in prior reports used five axes. Each individual PSU was then isolated from the global NMDS ordination plot, and coerced into two distinct clusters using k-means clustering. The sum of squares distance between the two cluster centers (BSS) based on their Voronoi sets was calculated for each PSU to obtain a test statistic that we used as a description of vegetation community distinctiveness. A higher BSS value (greater distance between the two clusters) means a more distinct vegetation community structure (Fig. 11a). Conversely, more overlapping clusters (smaller BSS) indicates less distinctiveness between sites, and a more degraded landscape structure (Fig. 11b). Because of the artificiality of segregating such data into two distinct clusters, rather than allowing for multiple clusters, we empirically assessed the extent to which this approach described the distinctiveness of ridge-slough communities, as described below.

We used several approaches to assess how well the clustering in the NMDS ordination plot described differentiation of ridge and slough communities. First, we analyzed the distribution of key indicator taxa (including sawgrass (*C. jamaicense*); *Eleocharis*; *Nymphaea*; *Utricularia*; in the two global clusters. Second, we compared the agreement between cluster assignments in the global analysis and within individual PSUs. While the former will best differentiate hydrologically distinct communities, the latter measures their separation within a given portion of the landscape. Third, we analyzed the covariation among characteristic species of each community in NMDS space. We plotted the 22 most abundant species in two-dimensional ordination space, and categorized them based on the a priori vegetation community in which they were most abundant. If our ordination and clustering approach captures ridge-slough community structure, then species within each a priori community should be closely associated in NMDS space. Last, we assessed distribution of sample points along individual axes from the global NMDS for an illustrative subset of PSUs that included three well-conserved landscape blocks from central WCA3AS, and degraded landscapes characterized by different hydrologic alterations. If our ordination approach and measurement of community distinctness effectively differentiates ridges and sloughs, then conserved landscapes should exhibit distinct modes along one or more NMDS axes, and these modes should correspond to k-means clusters. Overall, variation in cluster distance corresponded to the degree of clustering: vegetation in conserved landscapes was well-described as two distinct clusters.

We used regression analysis between long-term mean water depth and community distinctiveness for each PSU to assess how hydrologic regime influenced vegetation community distinctiveness. To assess whether the vegetation community distinctiveness might exhibit global bi-stability (Fig. 4), we generated histograms of cluster distances, and tested for bi-modality across PSUs in the same manner as tests of elevation bi-modality within PSUs.

In addition to separate measures of microtopographic structure and vegetation community distinctiveness, we also evaluated landscape structure based on three measures of the co-variation between elevation and vegetation community composition. First, we used bivariate regression analysis to assess the strength of the relationship between sawgrass abundance and elevation within each PSU. Second, again for each PSU, we used a Mantel test to determine the relationship between matrices of between-site dissimilarities in elevation and in community composition. The resulting test statistic r is a multivariate analog of Pearson's correlation coefficient. Finally, we evaluated the difference in elevation between points assigned to the two clusters in our k-means

analysis. This suite of measures provides a more integrated view of vegetative and microtopographic structure of ridge-slough landscapes, and differ in the effort required for data collection and analysis. To assess whether elevation-vegetation relationships within PSUs supported the occurrence of global bi-stability (Fig. 4), we generated histograms of all three measures of elevation PSU-scale elevation variance, and tested for bi-modality across PSUs in the same manner as tests of elevation bi-modality within PSUs.

We assessed the geographic variation in community distinctiveness (as measured by cluster distance, microtopographic heterogeneity (as measured by standard deviation of elevation), and elevation-vegetation association (as measured by sawgrass-elevation correlation, Mantel r , and elevation differences between vegetation clusters) to determine whether these characteristics co-varied across the greater Everglades. Maps of sampled PSUs were used to depict the condition of each PSU based on these measures. Because these measures have different units and different structures of variability across each PSU, scaling of condition is not uniform across different metrics, and we were not able to explicitly assess the relative degree of degradation by comparison of different metrics. However, spatial covariation among these measures provided some information about the extent of agreement among them.

We used Pearson's correlation coefficient to assess covariation among measures of ridge-slough landscape condition in a non-spatial context. If microtopography and vegetation structure (and their association within PSUs) covary strongly, then measures of these characteristics provide little independent information. However, weaker correlations between these measures would indicate that microtopography and vegetation structure vary somewhat independently. In that case, independent measures of these characteristics are important for assessment of ridge-slough condition. Moreover, the covariation of vegetation structure and microtopography across PSUs may provide some insight into the trajectories of landscape degradation. To assess the relative timescales of vegetation and topographic change in response to modification of the hydrologic regime, we compared the changes in the distinctness of the vegetation communities and loss of peat elevation structure within each PSU. We sorted PSUs into four quadrants delineated by the distinct modes observed in the distribution of each variable. In this design, if vegetation changes first, co-occurrence of intact topography (bimodal elevations) and reduced community distinctness should be observed. However, if topography changes first, then the reverse pattern should occur (Fig. 5). We tested this prediction by assigning for each PSU a single test statistic value for vegetation community distinctiveness (question 1, above) and another for microtopography distinctiveness, defined by the standard deviation of soil elevation. We then assigned each PSU to a quadrant, and compared these quadrants based on a variety of measures including hydrologic regime, vegetation community abundances, and vegetation-elevation correlations.

Software

All analyses and visualizations were performed in the open source statistical program R. The global NMDS plot was created using the metaMDS function in the vegan package (Oksanen et al. 2012). The default convergence criteria in monoMDS – the engine used by metaMDS which induces random starts – was too slack to find a convergent solution. The slack was tightened by using “ $sfgmin = 1e-7$ ”. The dissimilarity matrix for the NMDS was calculated using the vegdist function in vegan using the metric Jaccard index which was preferentially chosen over the popular

semi-metric Bray-Curtis index. k-means clusters were created using the R base package stats (R Core Team 2012).

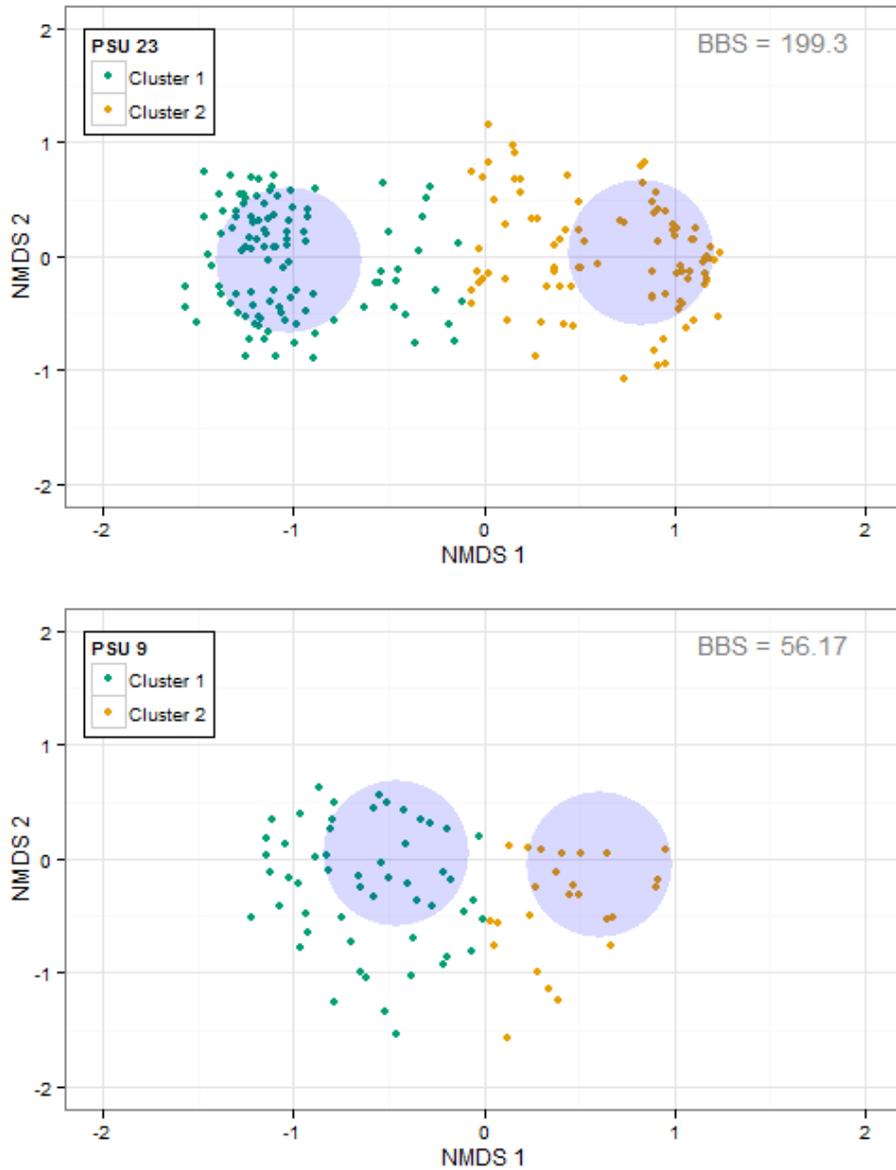


Figure 11: PSU 23 (a) and PSU 9 (b) species data ordinated by NMDS and clustered by k-means. In a), PSU 23 clusters are relatively far apart, indicating a significant separation of sites composed of species that occupy specific hydrologic niches; a relatively well preserved PSU. In b), PSU 9 clusters are closer, indicating a loss of distinctiveness in the vegetation community structure; a relatively degraded PSU.

RESULTS

Microtopographic and hydrologic patterns

Microtopographic patterns varied substantially across our broad landscape sample (Table 2; Fig. 12-16). Absolute mean elevations varied from 3 to 443 cm above sea level, and mean water level varied from 0.1 to 453 cm asl, both varying predominantly along the dominant north-south landscape slope from Lake Okeechobee to Florida Bay. Long-term mean water depths (spatially averaged over all points within each PSU) varied from -12 cm to 93 cm, with the lowest water depths found in units within the marl prairies of ENP. Temporal variability of water level also differed among PSUs, with the standard deviation of water elevation ranging from 14.3 to 34.7 cm.

Average water depth was highly but non-linearly correlated with average inundation across PSUs (Fig. 17). The non-linear effects of depth on inundation frequency means that inundation frequency does not capture effects of prolonged deep impoundment, and is therefore a less general and sensitive measure of hydrologic conditions for comparisons at the whole-PSU scale. Across PSUs, variation in water level over time was uncorrelated with mean water depth, and of much smaller magnitude (Fig. 18). Water level and hydroperiod calculated over different temporal windows covaried strongly; as a result, shorter windows provide little additional information over long-term averages.

The magnitude and structure of microtopographic relief also varied considerably among PSUs (Table 2). Standard deviations of elevation ranged from 2.3 to 26.3 cm, with most values falling between 6 and 11 cm. The magnitude of topographic relief was generally highest in central portions of WCA3AS, but was also high in individual PSUs within WCA1, WCA2, WCA3N, and ENP (Fig. 19). In PSU 73, extremely high soil elevation variance appears to be associated with cat-tail invasion into sloughs. Landscape-scale variation in elevation was bi-modally distributed, with modes centred on 6-7 cm and ca. 12 cm (Fig. 20). The skewness of elevations ranged from -1.35 to 1.47, with most values between -0.5 and 0.5. Kurtosis varied from -1.29 to 4.02, with most values slightly positive. One sampling area (PSU 61) exceeded these values considerably. Contrary to previous findings (Watts et al. 2010), kurtosis was not diagnostic of elevation bi-modality within PSUs.

Of the 62 PSUs sampled to date that fall within the historic distribution of the ridge and slough, 30 had elevation distributions that were better fit by a mixture of 2 normal distributions than by a single normal distribution (Table 2). These bimodal distributions were restricted to PSUs with long-term mean water depths of ca. 20-66 cm; differences in the elevation of these modes ranged from 9 to 23.4 cm (with PSU 73 as an outlier at 54 cm), and these differences increased with long-term mean water depth (Fig. 21). Differences between elevation modes were slightly lower than those measured by Watts et al. (2010) at comparable long-term mean water depths. One anomalous PSU (PSU 3 in WC3AN) exhibited bimodal elevation distributions but with minimal separation between means. Three other PSUs with high cattail abundance (Table 3) also exhibited high elevation variance and bi-modal distributions. PSUs outside of the historic ridge slough landscape, predominantly those within the marl prairie habitat of ENP, generally had unimodal elevation distributions with minimal variance; confidence in these distributions is lower

because data were collected during relatively dry periods when a large proportion of points were above the water surface.

Among PSUs with bi-modal elevation distributions, the difference in elevation between ridges and sloughs was closely correlated with the standard deviation of elevation; PSUs with unimodal elevation distributions generally occupied a lower and smaller range of elevation variance (Fig. 22).

Table 2. Hydrologic and microtopographic characteristics of year 1 and 2 PSUs. Additional hydrologic descriptors at the point scale are included in data reports for each PSU.

PSU	Water Elevation Statistics						Elevation Cluster Analysis						
	Water Elevation		Peat Surface		Kurtosis	Skew	Mode 1			Mode 2		*Best Model	
Mean (cm asl)	‡St. Dev. (cm)	Mean Water Depth (cm)	†St. Dev. (cm)	Depth (cm)			†St. Dev. (cm)	‡‡Mode Weight (q)	Depth (cm asl)	†St. Dev. (cm)	‡‡Mode Weight (q)		
0	180.0	24.8	30.80	7.31	-0.32	0.80	25.45	2.35	0.52	39.38	6.28	0.48	2V
1	448.4	15.3	8.18	5.98	-0.25	-0.16	8.18	5.97	1.00	-	-	-	1X
2	254.4	24.5	50.11	10.65	-0.27	-0.48	36.12	7.42	0.23	54.23	7.42	0.77	2E
3	305.0	25.9	-4.61	3.63	-1.07	-0.19	-6.28	2.88	0.71	-0.50	0.95	0.29	2V
4	261.9	26.3	40.77	11.89	-1.03	0.16	32.64	6.63	0.59	52.66	6.63	0.41	2E
5	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
6	33.5	22.2	27.31	6.49	3.56	0.65	27.31	6.47	1.00	-	-	-	1X (7V)
7	287.6	22.2	33.05	6.46	0.97	0.05	33.05	6.44	1.00	-	-	-	1X
8	113.0	34.4	-8.24	12.64	0.08	0.87	-16.37	5.25	0.55	1.64	11.83	0.45	2V
9	357.7	26.1	28.34	4.94	0.85	0.35	28.34	4.93	1.00	-	-	-	1X
10	195.1	28.2	23.97	2.28	0.13	0.80	23.97	2.24	1.00	-	-	-	1X (3V)
11	271.6	32.1	53.36	6.89	1.20	-0.07	53.36	6.87	1.00	-	-	-	1X
81	177.8	21.5	31.85	5.73	1.04	-0.44	31.85	5.72	1.00	-	-	-	1X
13	190.9	15.5	52.73	8.32	0.02	-0.40	52.73	8.30	1.00	-	-	-	1X
14	0.1	19.8	-3.36	5.26	0.65	0.59	-3.36	5.25	1.00	-	-	-	1X (5V)
15	272.1	31.2	71.57	8.76	0.09	0.02	71.57	8.74	1.00	-	-	-	1X
16	112.9	34.7	-12.41	7.48	1.56	1.40	-18.45	0.50	0.38	-8.65	7.22	0.62	2V (3V)
17	448.2	19.6	27.65	13.09	3.01	1.11	20.68	5.33	0.50	34.48	14.68	0.50	2V
18	152.9	24.5	29.74	7.09	-1.25	-0.01	23.97	3.48	0.53	36.26	3.48	0.47	2E
19	289.1	22.7	20.70	8.34	-0.45	0.30	15.26	4.98	0.60	28.86	4.98	0.40	2E
20	184.7	15.8	31.17	5.10	-0.19	-0.64	23.90	3.41	0.21	33.14	3.41	0.79	2E (9V)
21	329.2	28.8	39.52	11.56	-0.11	0.85	31.18	4.17	0.48	47.35	10.72	0.52	2V
22	31.5	17.8	20.02	7.02	-0.25	0.08	20.02	6.99	1.00	-	-	-	1X (13V)
23	265.3	21.8	30.89	10.34	-1.20	0.33	23.99	5.27	0.62	42.30	5.27	0.38	2E (3E)
24	157.5	20.4	34.18	6.22	-0.06	-0.51	34.18	6.20	1.00	-	-	-	1X
25	449.9	15.4	6.53	6.63	2.27	0.57	6.36	4.31	0.66	6.86	9.66	0.34	2V (8V)
26	261.7	23.2	41.47	10.85	-0.79	0.01	33.53	6.00	0.56	51.70	6.00	0.44	2E
27	283.1	29.6	18.95	13.73	-1.27	0.12	3.09	2.66	0.33	26.75	9.55	0.67	2V (4E)
28	187.3	17.3	32.15	5.31	0.10	-0.35	32.15	5.29	1.00	-	-	-	1X
29	302.0	30.9	-8.30	3.02	1.24	-0.88	-8.30	3.00	1.00	-	-	-	1X
30	123.8	21.1	23.11	8.85	-0.35	0.03	23.11	8.83	1.00	-	-	-	1X
31	268.6	26.9	38.12	6.69	-0.25	0.27	38.12	6.67	1.00	-	-	-	1X

‡Standard Deviation of water elevation describes the temporal variability of water level at the center point of each PSU.

†Standard Deviation of water depth describes the spatial variability of soil elevation across all points sampled within each PSU.

‡‡ Mode weight describes the proportion of data that occur within each mode, allowing for imbalance in mode prevalence

* Best fit model selected based on Bayes' Information Criterion; number refers to the number of modes, E and V denote whether variances of the two modes are equal (E) or unequal (V). Where the best fit model included more than 2 modes, data presented are from the best fit model among 1 and 2 mode models.

Table 2: Continued.....

PSU	Water Elevation Statistics						Elevation Cluster Analysis						
	Water Elevation			Peat Surface			Mode 1			Mode 2			*Best Model
	Mean (cm asl)	§St. Dev. (cm)	Mean Water Depth (cm)	†St. Dev. (cm)	Kurtosis	Skew	Depth (cm)	†St. Dev. (cm)	††Mode Weight (q)	Depth (cm asl)	†St. Dev. (cm)	††Mode Weight (q)	
32	160.02	19.04	37.08	6.34	0.1	-0.18	37.08	6.32	1	-	-	-	
34	247.67	22.92	55.34	12.97	3.16	-0.71	55.34	12.92	1	-	-	-	1X
35	311.31	25.56	13.74	4.39	-0.95	0.24	13.74	3.63	1	-	-	-	1X
36	256.76	30.52	83.08	7.62	0.41	-0.93	70.11	4.4	0.09	84.39	4.4	0.91	2E
37	335.98	24.9	31.33	10.05	-0.71	0.45	25.31	5.91	0.64	42.17	5.91	0.36	2E
39	289.66	24.34	23.76	6.68	-0.66	0.09	17.86	3.72	0.4	27.77	3.72	0.6	2E
43	275.9	25.67	27.61	5.32	-0.36	0.33	27.61	3.99	1	-	-	-	1X
44	177	19.04	34.23	4.67	4.02	-1.35	17.33	3.86	0.02	34.63	3.86	0.98	2E
45	265.73	33.94	92.5	9.28	0.37	-0.5	71.77	6.87	0.05	93.65	6.87	0.95	2E
47	269.42	33.02	51.93	11.26	-0.4	0.39	51.93	7.31	1	-	-	-	1X
513	154.55	22.82	29.77	5.19	0.53	-0.41	29.77	5.08	1	-	-	-	1X
DPM	188.96	14.27	65.95	14.31	-0.84	-0.21	53.09	8.23	0.45	76.53	8.23	0.55	2E
51	304.11	23.59	20.06	6.84	-0.64	0.00	20.06	6.81	1	-	-	-	1X
52	242.0	25.9	54.6	15.7	2.7	-0.6	51.59	27.15	0.19	55.29	11.09	0.81	2V
53	254.5	30.0	60.7	13.8	-1.1	0.1	50.28	7.65	0.54	73.16	7.65	0.46	2E
55	267.6	20.7	38.0	9.4	-0.7	-0.3	22.66	2.19	0.18	41.28	6.64	0.82	2V
56	163.0	16.8	32.5	8.7	1.5	0.2	32.51	8.68	1	-	-	-	1X
58	244.7	23.1	59.7	11.0	-0.4	-0.3	59.71	10.93	1	-	-	-	1X
61	342.5	21.5	29.0	10.6	58.3	6.3	28.40	5.33	0.98	55.67	55.92	0.02	2V
62	99.3	20.3	30.4	7.3	-0.1	-0.1	30.42	7.23	1	-	-	-	1X
63	259.7	31.8	81.1	7.3	1.5	-0.6	64.61	5.64	0.07	82.39	5.64	0.93	2E
220	186.1	17.6	36.6	4.1	1.3	-0.2	36.56	4.10	1	-	-	-	1X
BS1	169.1	18.0	18.1	7.9	1.3	0.6	18.65	2.07	0.32	17.88	9.38	0.68	2V
BS2	169.6	18.3	24.3	9.4	0.6	0.2	24.34	9.32	1	-	-	-	1X
BS3	167.1	17.7	28.8	9.3	5.7	0.9	28.78	9.24	1	-	-	-	1X

§Standard Deviation of water elevation describes the temporal variability of water level at the center point of each PSU.

†Standard Deviation of water depth describes the spatial variability of soil elevation across all points sampled within each PSU.

†† Mode weight describes the proportion of data that occur within each mode, allowing for imbalance in mode prevalence. Modes

* Best fit model selected based on Bayes' Information Criterion; number refers to the number of modes, E and V denote whether variances of the two modes are equal (E) or unequal (V). Where the best fit model included more than 2 modes, data presented are from the best fit model among 1 and 2 mode models.

Table 2: Continued.....

PSU	Water Elevation Statistics						Elevation Cluster Analysis						
	Water Elevation			Peat Surface			Mode 1			Mode 2			
	Mean (cm asl)	§St. Dev. (cm)	Mean Water Depth (cm)	†St. Dev. (cm)	Kurtosis	Skew	Depth (cm)	†St. Dev. (cm)	††Mode Weight (q)	Depth (cm asl)	†St. Dev. (cm)	††Mode Weight (q)	*Best Model
50	148.3	18.9	36.2	9.64	0.00	-0.28	36.25	9.61	1	-	-	-	1X
54	78.6	19.2	27.1	8.31	0.06	-1.00	20.58	4.66	0.52	34.27	4.66	0.48	2E
65	452.6	15.6	33.4	14.88	0.28	0.08	33.36	14.82	1	-	-	-	1X
66	257.0	23.9	45.0	11.0	-1.1	0.0	36.60	5.60	0.55	55.43	5.60	0.45	2E
67	305.7	23.3	11.2	6.62	-0.12	-0.26	11.23	6.60	1	-	-	-	1X
68	251.1	27.1	46.9	16.38	-0.15	-0.41	46.90	16.32	1	-	-	-	1X
69	337.9	25.8	44.6	9.21	0.10	-0.19	44.54	9.13	1	-	-	-	1X
71	278.5	22.2	35.9	7.32	-0.33	-1.09	29.23	4.35	0.48	41.96	2.68	0.52	2V
73	367.1	27.1	49.2	26.29	0.94	-0.71	33.51	7.07	0.71	87.67	12.66	0.29	2E
79	273.6	28.3	48.4	6.38	-0.17	-0.25	48.36	6.36	1	-	-	-	1X

§Standard Deviation of water elevation describes the temporal variability of water level at the center point of each PSU.

†Standard Deviation of water depth describes the spatial variability of soil elevation across all points sampled within each PSU.

†† Mode weight describes the proportion of data that occur within each mode, allowing for imbalance in mode prevalence. Modes

* Best fit model selected based on Bayes' Information Criterion; number refers to the number of modes, E and V denote whether variances of the two modes are equal (E) or unequal (V). Where the best fit model included more than 2 modes, data presented are from the best fit model among 1 and 2 mode models.

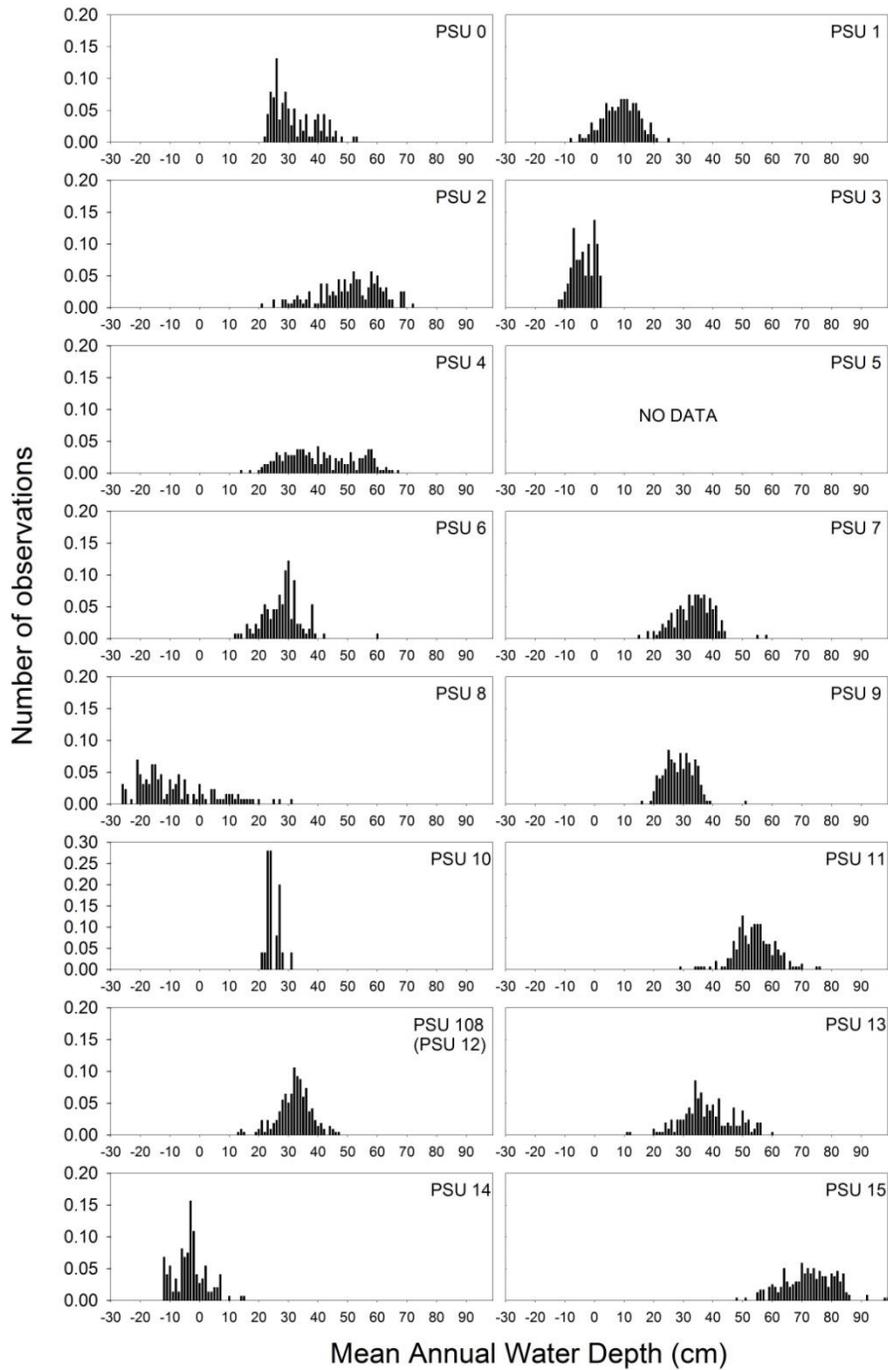


Figure 12: Elevation distributions of Year 1 PSUs. Bimodality and high variability in elevation (e.g. PSU 4) are characteristics of conserved conditions, while low variability and unimodality (e.g. PSU 11) are characteristic of degraded conditions. Mean annual water depth (our measure of relative elevation) is calculated from water depth on the day of sampling, and benchmarked to long-term average water level at the center point of each PSU. Summary statistics and bimodality analysis for each PSU are presented in Table 2.

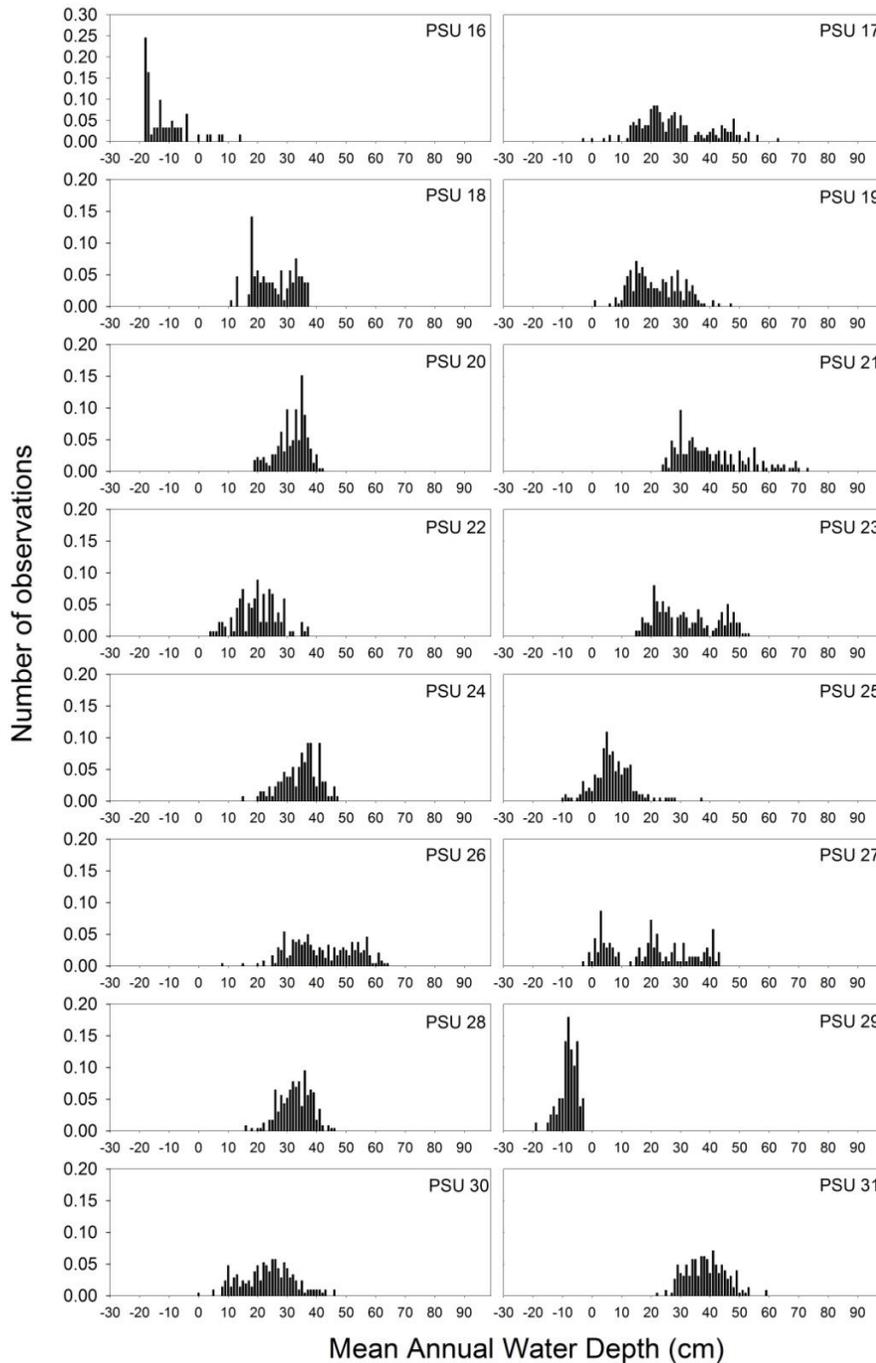


Figure 13: Elevation distributions of Year 2 PSUs. Bimodality and high variability in elevation (e.g. PSU 26) are characteristics of conserved conditions, while low variability and unimodality (e.g. PSU 28) are characteristic of degraded conditions. Mean annual water depth (our measure of relative elevation) is calculated from water depth on the day of sampling, and benchmarked to long-term average water level at the center point of each PSU. Summary statistics and bimodality analysis for each PSU are presented in Table 2.

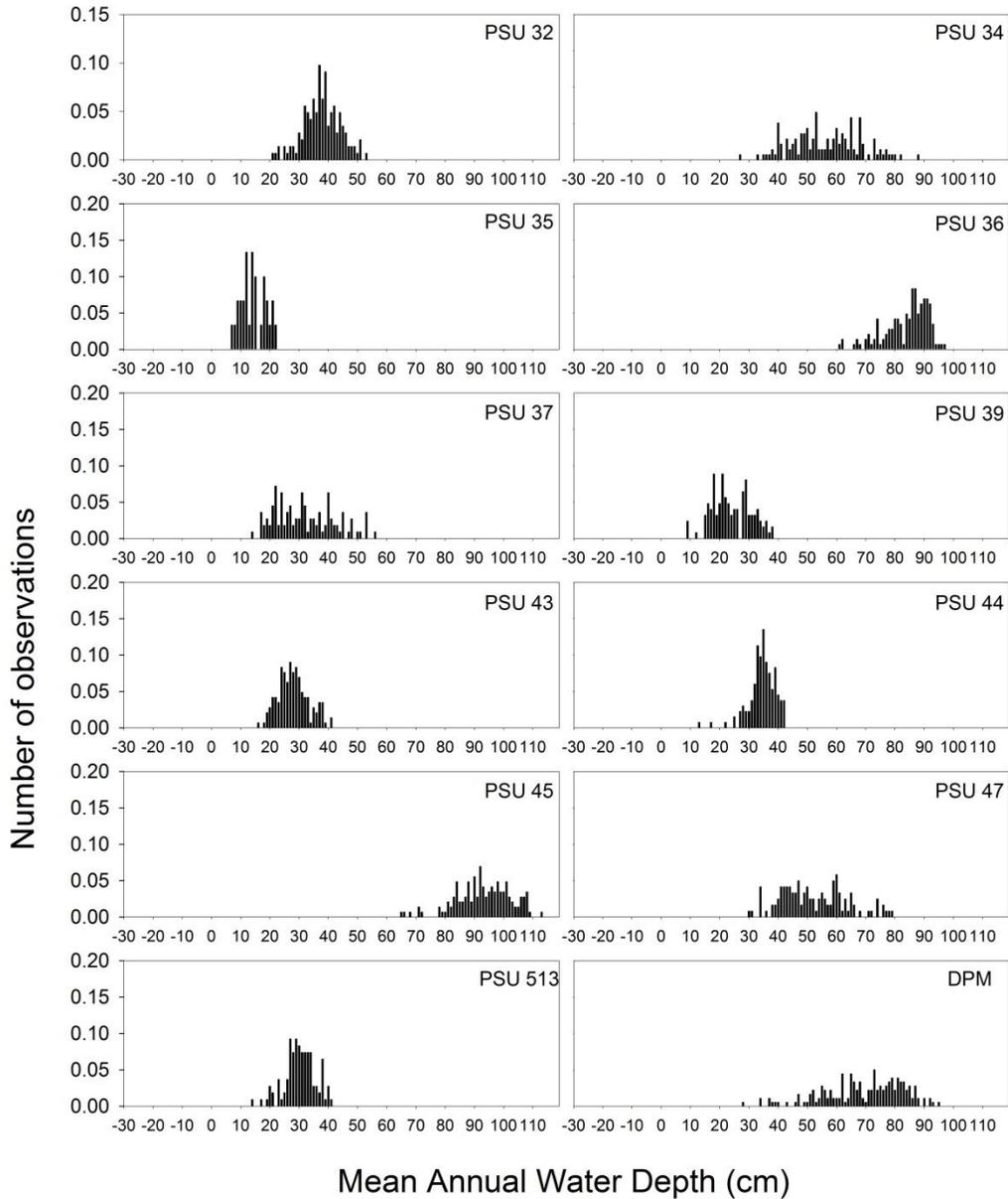


Figure 14: Elevation distributions of Year 3 PSUs. Bimodality and high variability in elevation (e.g. PSU 34) are characteristics of conserved conditions, while low variability and unimodality (e.g. PSU 43) are characteristic of degraded conditions. Mean annual water depth (our measure of relative elevation) is calculated from water depth on the day of sampling, and benchmarked to long-term average water level at the center point of each PSU. Summary statistics and bimodality analysis for each PSU are presented in Table 2.

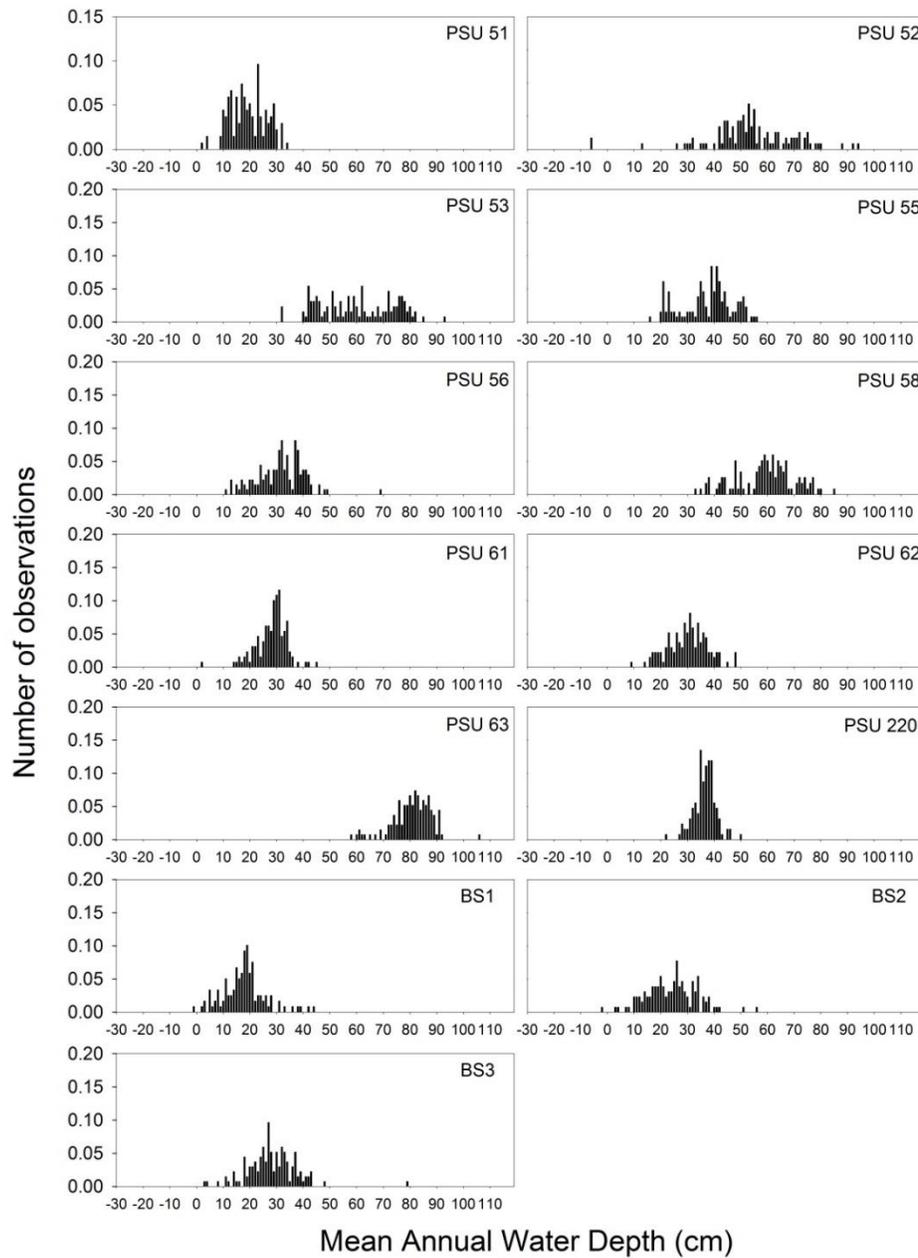


Figure 15: Elevation distributions of Year 4 PSUs. Bimodality and high variability in elevation (e.g. PSU 53) are characteristics of conserved conditions, while low variability and unimodality (e.g. PSU 62) are characteristic of degraded conditions. Mean annual water depth (our measure of relative elevation) is calculated from water depth on the day of sampling, and benchmarked to long-term average water level at the center point of each PSU. Summary statistics and bimodality analysis for each PSU are presented in Table 2.

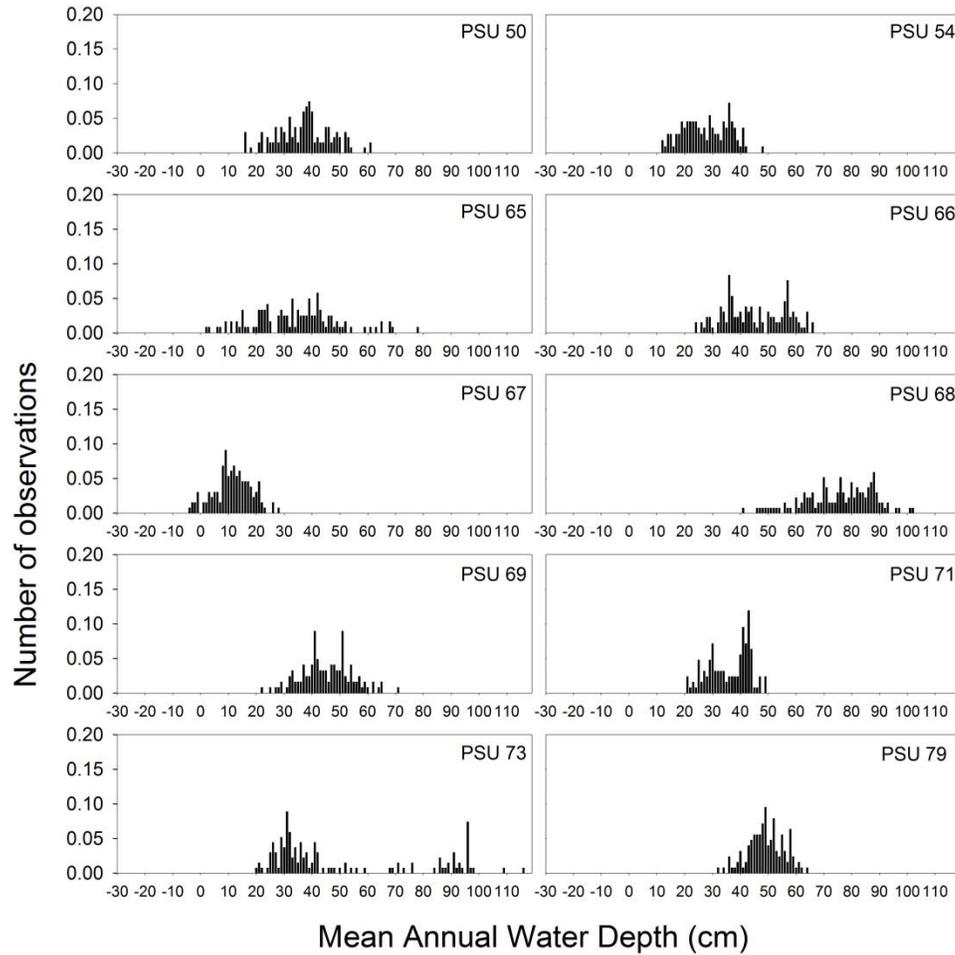


Figure 16: Elevation distributions of Year 5 PSUs. Bimodality and high variability in elevation (e.g. PSU 66) are characteristics of conserved conditions, while low variability and unimodality (e.g. PSU 67) are characteristic of degraded conditions. Mean annual water depth (our measure of relative elevation) is calculated from water depth on the day of sampling, and benchmarked to long-term average water level at the center point of each PSU. Summary statistics and bimodality analysis for each PSU are presented in Table 2.

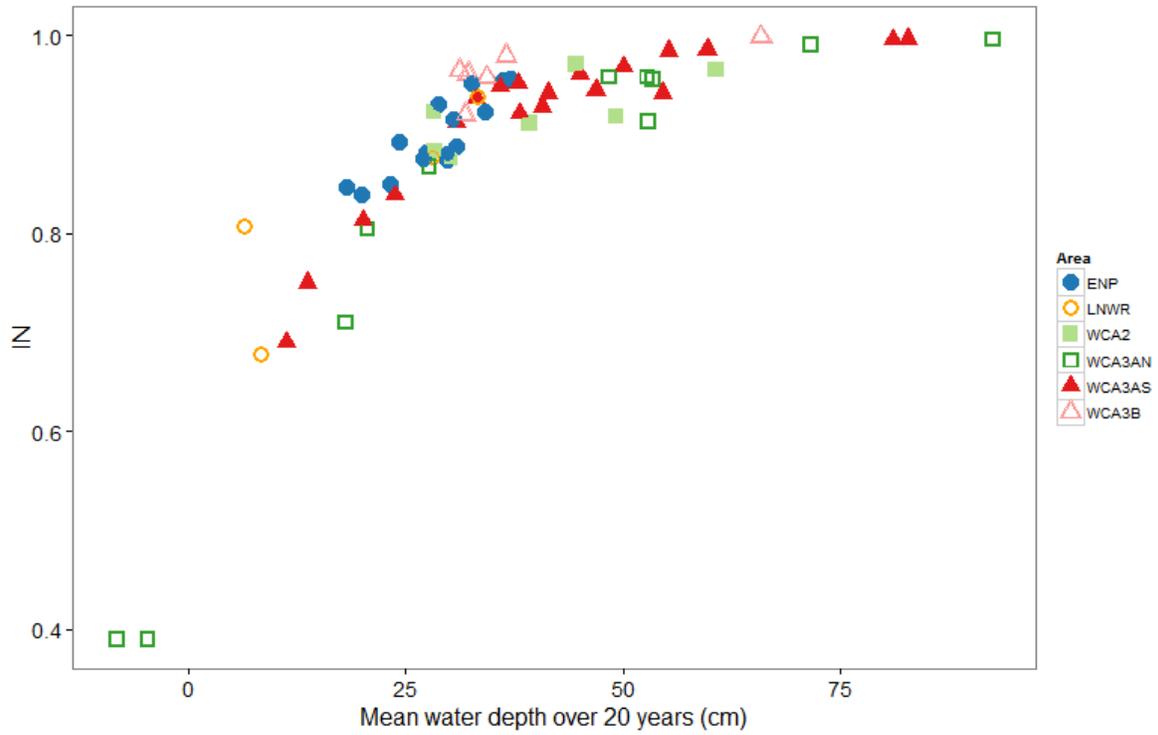


Figure 17: Relationship between average long-term mean water depth and average inundation frequency across PSUs. Symbols indicate PSUs within different hydrologic basins.

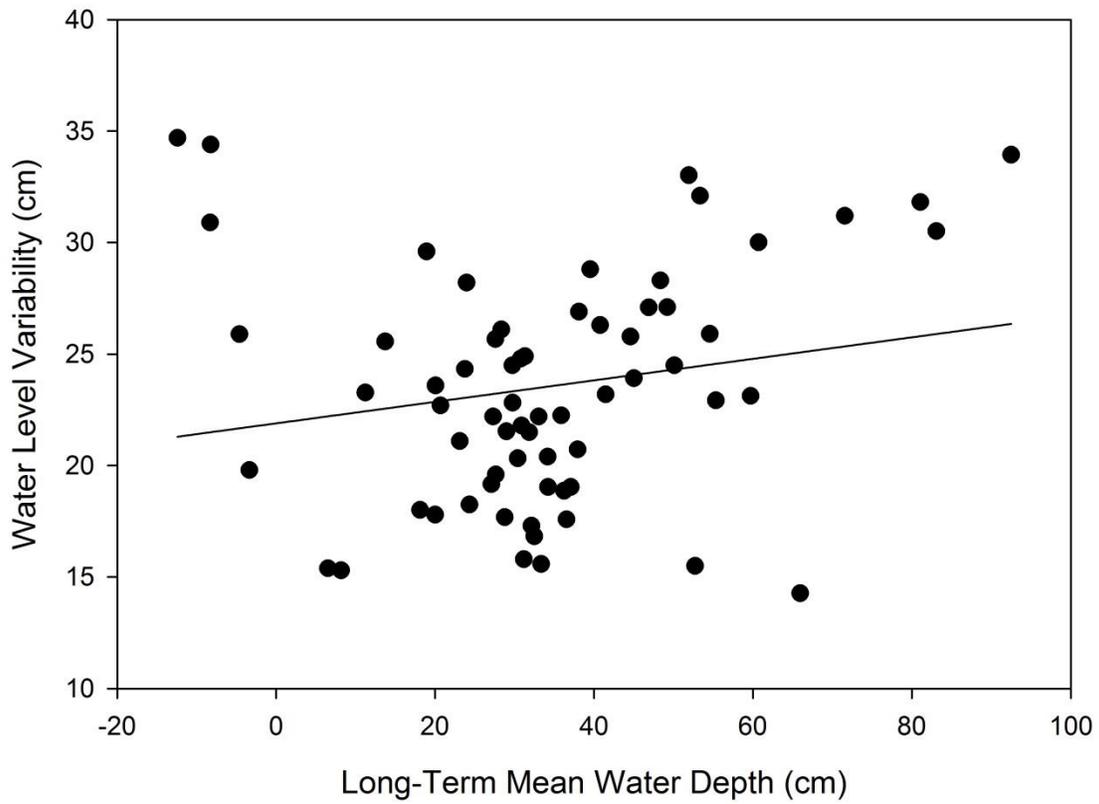


Figure 18: Relationship between average long-term mean water depth and temporal standard deviation of water depth. The weak correlation between these variables indicates that regions with dryer or wetter conditions do not necessarily have more stable or variable conditions.

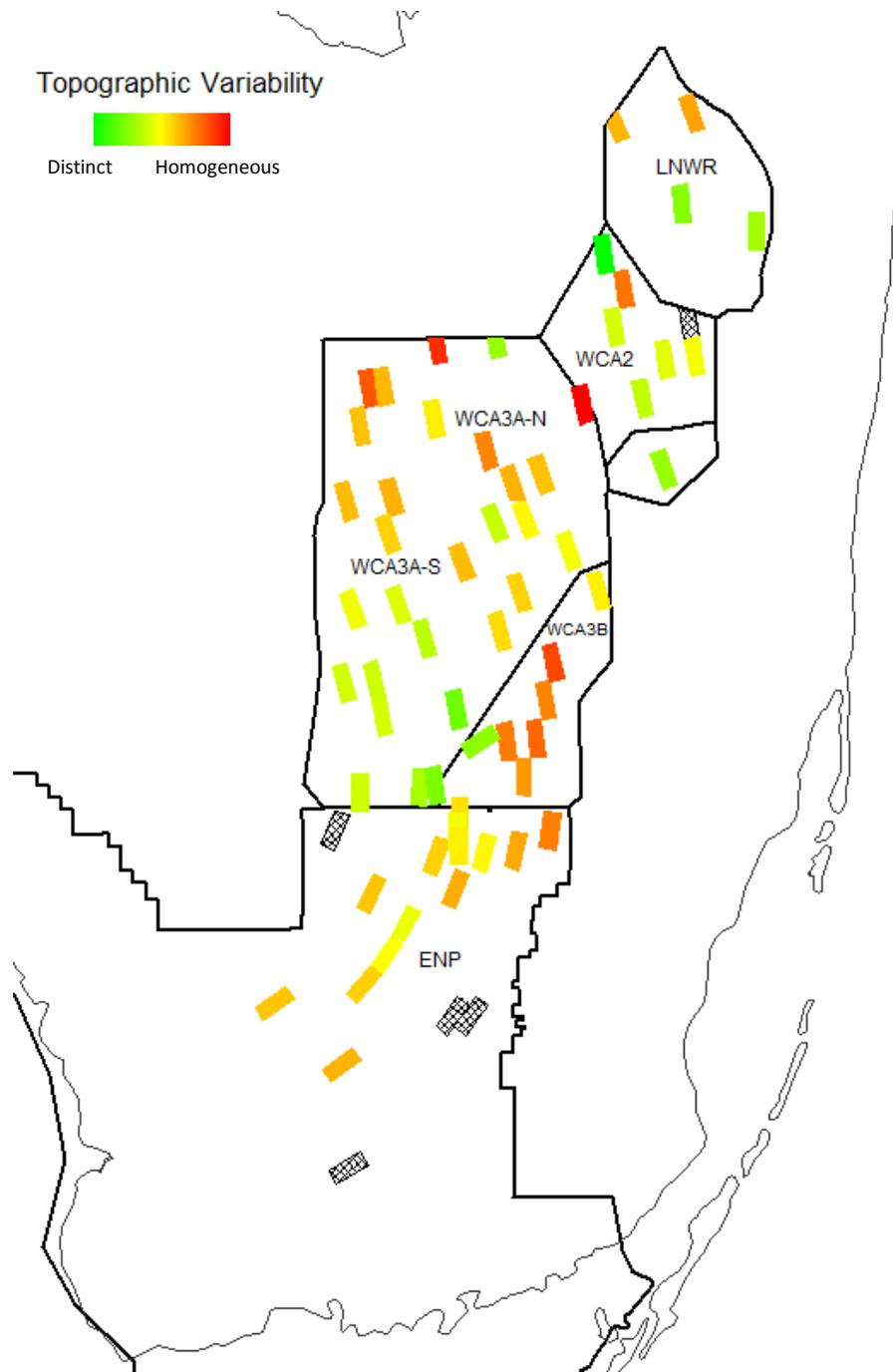


Figure 19: Spatial patterns of elevation variance across the historic ridge-slough landscape. Colors indicate the amount of microtopographic relief (measured as the standard deviation of elevation within each PSU). Color scale is log-transformed due to high variability in one outlier PSU (PSU 73- northern most site in WCA2)

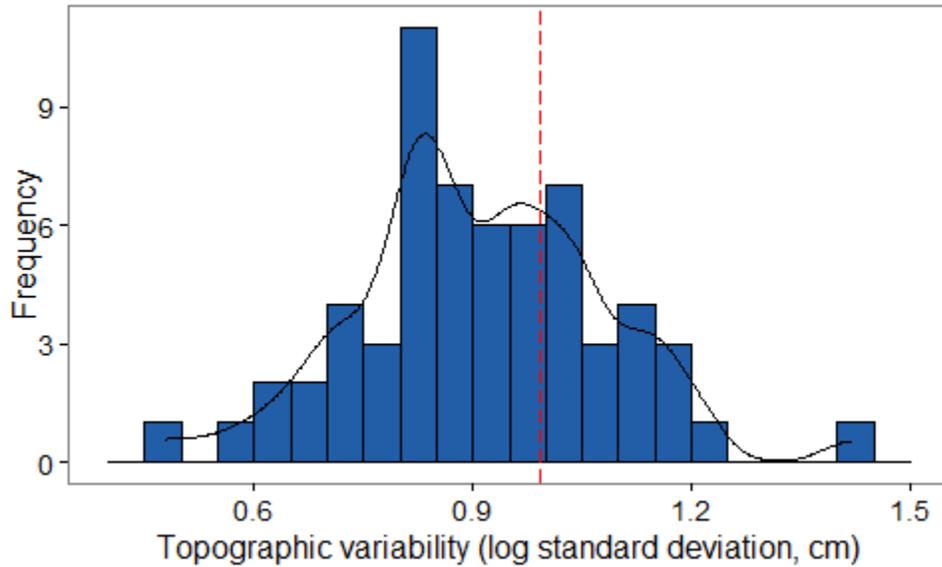


Figure 20: Frequency of elevation variation across PSUs. The bi-modal pattern observed in these data is consistent with current hypotheses about bistability of homogeneous and patterned configurations of the ridge-slough landscape. Data from PSUs outside the historic ridge and slough are not included.

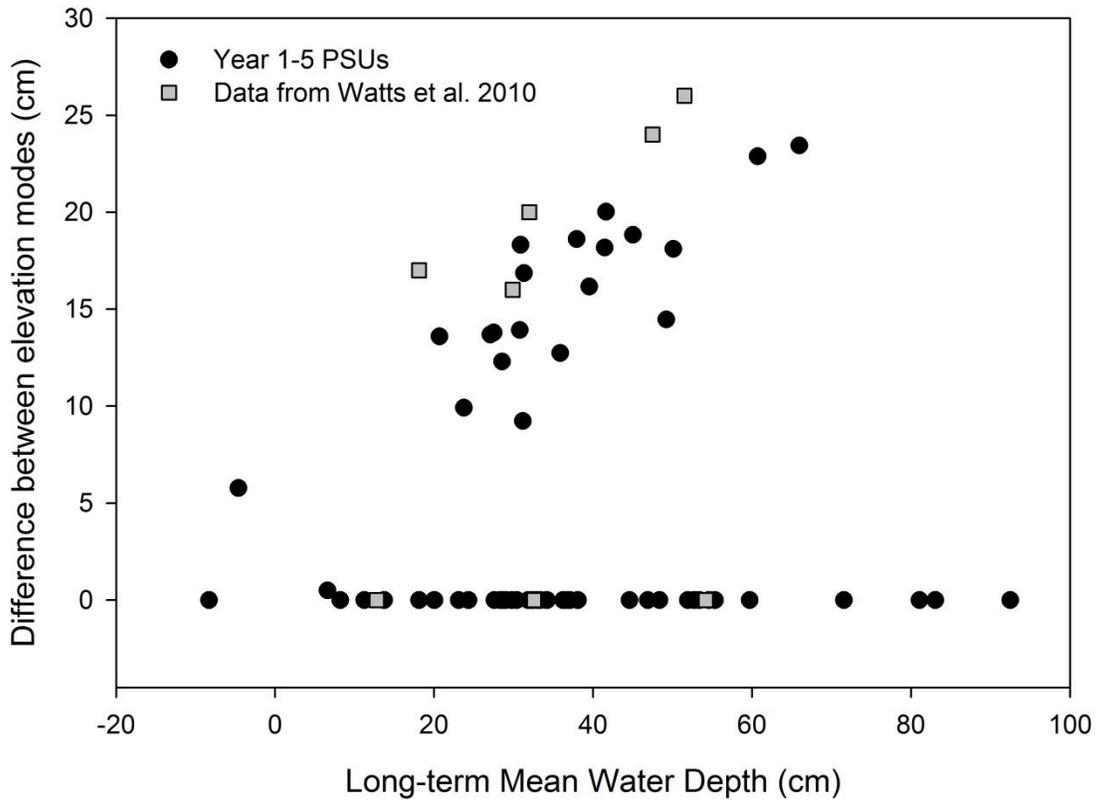


Figure 21: Relationship between hydrologic conditions and elevation mode separation across PSUs. Mean annual water depth is calculated as the difference between the temporal mean of water level since 1991 and the mean elevation of sampled points within each PSU. Elevation mode differences are based on cluster analysis results presented in Table 2. Points with zero difference between elevation modes are those whose elevation distributions were best fit by a single normal distribution. Data include PSUs sampled by Watts et al. (2010) but not any duplicate measurements of individual PSUs. Data from PSUs outside the historic ridge and slough are not included. The anomalous observation from PSU 27 was excluded based on severe recent fire history and incursion by cattail.

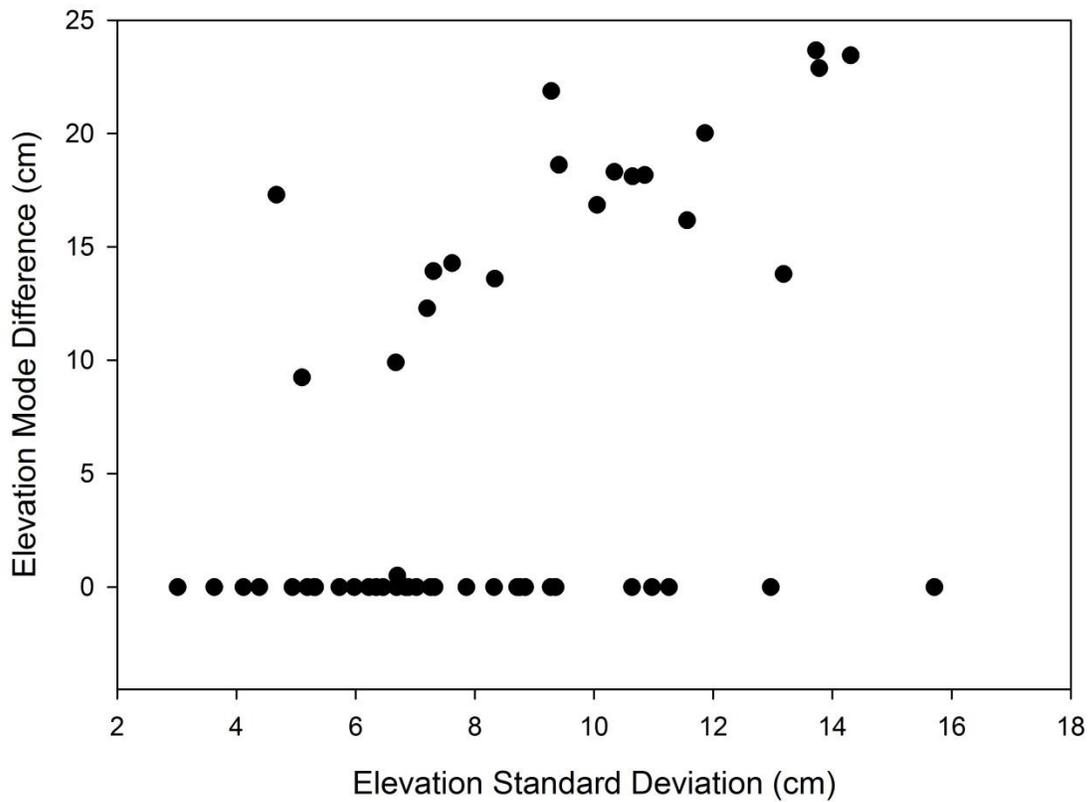


Figure 22: Covariation between two measures of landscape-scale microtopographic structure: the standard deviation of soil surface elevation within a PSU, and the difference between the means of elevation modes in the same PSU, as estimated from mixed distribution modeling. Points with zero difference between elevation modes represent PSUs whose elevation distributions were best described by a single normal distribution.

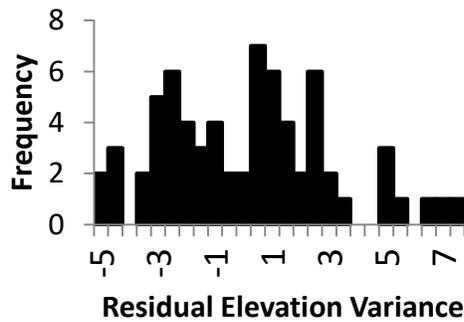
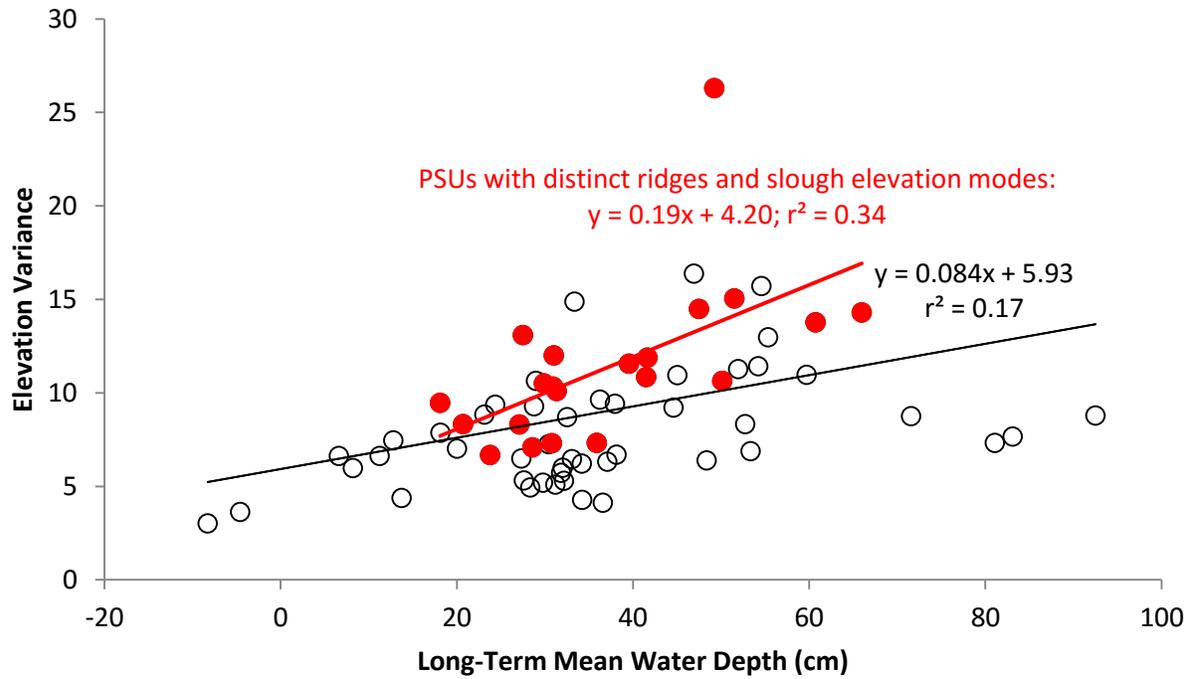


Figure 23: Correlation between long-term mean water depth and elevation variance. In the top panel, red symbols indicate PSUs with bi-modal soil elevation distributions. Elevation variance in these PSUs is higher and more strongly correlated with long-term mean water depth. Residual variation in soil elevation standard deviation exhibits clear bi-modality (bottom).

Vegetation community composition and structure

The composition of vegetation communities varied considerably within and across sampled PSUs (Table 3). Within a sample of 9239 points across sixty-two 2x5 km PSUs, species generally exhibited expected trends with water depth; however, these relationships varied in their strength across PSUs, and most species occurred across a wide range of local hydrologic conditions. For example, sawgrass (*Cladium jamaicense*) biomass (as measured by sawgrass biomass index) was most abundant in locations where long-term mean water depth were relatively shallow (15-35 cm long-term mean water depth), but dense tall sawgrass was observed across virtually the entire range of sampled hydrologic conditions (Fig. 24). Despite this variation in local hydrologic condition, the abundance of taxa followed expected trends with depth at the scale of whole PSUs. The average % cover of sawgrass was highest in PSUs with lower long-term mean water depth, while *Utricularia* spp. and *Nymphaea* spp. were most abundant in PSUs with high long-term mean water depths (Fig. 25). *Eleocharis* spp. were most abundant in PSUs with intermediate water depths (Fig. 25). In years 1 and 2, periphyton was more evenly distributed, occupying all depths except the driest (tree island and mixed tree island) and likely most eutrophic (cattail) communities. Periphyton abundance was not recorded in years 3-5.

Non-metric multidimensional scaling ordinated species in a manner consistent with previous studies of vegetation communities in the Everglades (Fig. 26). Sawgrass and other species common on ridges and tree islands were clearly separated from slough species along axis 1, while wet prairie species were intermediate along this axis, and somewhat differentiated along axis 2. The global k-means clustering analysis identified ridges dominated by sawgrass as one dominant cluster, and communities including both wet prairies and sloughs as a second dominant cluster (Fig. 27). Because those taxa separated strongly along NMDS axis 1, cluster membership sorted strongly along that axis (Fig. 28). Across all sample points, clusters also separated clearly along the hydrologic gradient of long-term mean water depth (Fig. 29). The dominant water depth threshold separating these vegetation clusters occurred around 40 cm.

K-means clustering within individual PSUs, from which we derived cluster distance as a measure of community distinctness, generally corresponded to the global K-means clustering. In 32 of the 58 historic ridge and slough PSUs, local and global clustering had greater than 90% agreement (i.e., less than 10% of points were differently assigned in the local and global cluster analysis). Agreement in 17 other PSUs was greater than 75%, and greater than 50% in all but 3 PSUs. PSUs with less agreement between local and global clusters tended to be those with low cluster distances.

Distributions of observations along NMDS axes from selected PSUs indicated that cluster distance was an effective proxy for the distinctness of ridge and slough communities distinctness (Fig. 30). In PSUs within central WCA3AS, local plant assemblages were strongly separated, with most observations occurring at the extremes of NMDS axis 1, in particular, which is the axis that most clearly differentiates ridge and slough communities. In PSUs with more extreme hydrologic regimes, observations tended toward intermediate values along the same axis, suggesting a blending of communities.

Table 3: Vegetation characteristics of PSUs sampled to date.

PSU	Field-Assigned Communities				Species Mean Relative Cover (%)					
	Ridge (%)	Wet Prairie (%)	Slough (%)	Mixed/Edge (%)	<i>Typha spp.</i>	<i>Cladium jamaicense</i>	<i>Nymphaea odorata</i>	<i>Eleocharis spp.</i>	<i>Periphyton</i>	<i>Utricularia Spp.</i>
0	57.9%	18.4%	7.9%	2.6%	0.0	53.2	0.0	14.0	20.8	4.4
1	48.1%	0.6%	39.5%	1.2%	0.3	38.1	7.2	26.2	0.0	13.5
2	34.6%	0.0%	57.2%	4.4%	0.8	27.1	10.7	0.4	30.3	22.7
3	100.0%	0.0%	0.0%	0.0%	0.0	37.2	0.0	1.5	1.9	1.3
4	44.4%	7.0%	47.7%	0.0%	0.4	37.3	28.4	4.4	1.7	15.5
5	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
6	38.2%	35.9%	0.0%	15.3%	0.2	39.5	0.0	13.2	40.4	3.7
7	55.7%	0.0%	44.3%	0.0%	0.3	37.5	7.9	10.6	21.2	6.7
8	16.28%	28.68%	1.55%		0.0	33.9	0.0	1.4	47.0	1.4
9	92.0%	0.0%	8.0%	0.0%	0.5	52.5	1.4	0.3	0.0	0.1
10	85.93%	8.89%	0.00%		0.0	42.9	0.0	0.3	46.4	0.0
11	31.6%	33.3%	35.1%	0.0%	7.2	22.1	31.5	1.6	9.6	18.9
81	49.8%	9.7%	5.1%	25.3%	3.4	52.6	1.5	4.0	16.4	13.2
13	70.0%	24.8%	3.3%	1.9%	2.0	72.1	3.0	0.2	16.4	3.6
14	6.12%	34.01%	0.00%		0.0	30.1	0.0	0.5	64.8	0.1
15	38.7%	3.4%	56.3%	1.3%	2.5	17.1	11.5	0.0	39.5	29.1
16	11.11%	45.19%	0.00%		0.0	52.5	0.0	1.6	41.7	0.8
17	44.6%	0.0%	38.3%	3.1%	1.6	40.8	9.0	13.4	0.4	18.9
18	32.1%	50.0%	0.0%	0.0%	0.0	31.4	0.0	13.0	39.0	0.2
19	31.4%	15.2%	7.6%	22.9%	22.6	35.7	2.0	3.8	17.8	2.6
20	92.0%	3.6%	0.9%	3.6%	0.0	29.7	1.0	3.1	61.1	0.4
21	54.3%	0.0%	42.5%	3.2%	1.0	45.7	0.1	15.7	35.2	2.0
22	39.3%	45.2%	0.0%	3.0%	0.0	41.2	0.0	15.5	26.4	1.7
23	44.7%	10.1%	28.7%	9.3%	0.3	42.8	10.3	9.3	13.7	8.3
24	71.0%	13.7%	3.1%	2.3%	0.0	58.0	0.0	2.0	31.9	2.9
25	86.9%	0.0%	9.9%	2.1%	2.6	78.4	0.1	1.1	0.5	2.4
26	25.8%	9.6%	45.0%	15.4%	0.7	27.6	9.5	3.3	20.0	25.0
27	34.1%	0.0%	0.7%	37.0%	18.6	56.6	0.2	0.0	3.1	0.3
28	68.0%	6.1%	11.3%	12.6%	0.0	58.4	2.2	3.1	18.3	14.5
29	66.7%	3.8%	6.4%	19.2%	0.0	82.9	0.0	2.0	8.2	0.8
30	56.9%	34.4%	1.0%	4.3%	0.0	55.0	0.5	13.8	14.7	2.7
31	43.6%	20.0%	18.7%	13.8%	0.0	50.0	10.7	6.1	14.9	5.2

Table 3: Continued.....

PSU	Field-Assigned Communities				Species Mean Relative Cover (%)					
	Ridge (%)	Wet Prairie (%)	Slough (%)	Mixed/Edge (%)	<i>Typha spp.</i>	<i>Cladium jamaicense</i>	<i>Nymphaea odorata</i>	<i>Eleocharis spp.</i>	<i>Periphyton*</i>	<i>Utricularia Spp.</i>
32	79.70%	10.50%	2.10%	10.55	0.00%	40.30%	1.50%	7.30%	-	19.50%
34	81.90%	10.10%	0%	8.00%	0.00%	23.90%	20.90%	2.80%	-	16.10%
35	73.30%	13.30%	0%	13.30%	0.00%	23.00%	0.00%	4.90%	-	0.30%
36	13.90%	0%	70.10%	16.00%	0.00%	10.60%	18.90%	0.00%	-	9.40%
37	27.90%	0%	11.70%	60.30%	0.00%	44.20%	6.60%	4.80%	-	8.10%
39	33.90%	43.60%	0%	22.60%	0.00%	27.20%	1.40%	7.00%	-	4.70%
43	84.70%	11.10%	0%	4.20%	0.00%	37.20%	0.90%	4.80%	-	8.10%
44	69.20%	18.00%	1.50%	11.30%	0.00%	20.20%	1.50%	4.80%	-	17.50%
45	8.30%	2.10%	43.80%	45.80%	0.00%	9.70%	9.40%	0.10%	-	25.00%
47	59.20%	0	9.20%	31.70%	0.00%	27.30%	7.30%	1.20%	-	16.70%
513	93.50%	0%	0%	6.55	0.00%	40.40%	0.00%	0.60%	-	11.80%
DPM	59.30%	5.56%	11.10%	24.10%	0.00%	26.20%	5.30%	5.50%	-	25.30%
51	60.74%	19.26%	0.00%	4.44%	0.00%	38.19%	4.85%	0.00%	-	0.11%
52	26.50%	2.56%	32.48%	15.38%	0.00%	28.43%	5.04%	14.69%	-	10.13%
53	29.46%	0.00%	57.36%	6.98%	0.00%	20.97%	3.79%	19.09%	-	19.81%
55	37.40%	9.92%	51.15%	1.53%	0.00%	21.56%	4.94%	9.13%	-	19.35%
56	64.44%	15.56%	0.00%	18.52%	0.00%	35.39%	5.52%	1.15%	-	8.56%
58	29.91%	66.67%	0.00%	0.00%	0.00%	21.54%	20.81%	1.54%	-	7.78%
61	44.96%	0.00%	13.18%	41.09%	0.00%	32.87%	0.84%	5.91%	-	9.69%
62	71.11%	22.96%	0.00%	4.44%	0.00%	35.44%	13.07%	0.78%	-	9.70%
63	7.41%	0.00%	87.41%	0.00%	0.00%	11.00%	0.44%	27.43%	-	46.48%
220	84.92%	0.00%	2.38%	8.73%	0.00%	50.82%	1.39%	3.69%	-	9.56%
BS1	58.33%	10.83%	0.00%	20.00%	0.00%	30.85%	4.71%	0.29%	-	14.98%
BS2	67.44%	6.98%	0.00%	18.60%	0.00%	32.33%	6.94%	0.05%	-	18.68%
BS3	72.59%	13.33%	0.00%	13.33%	0.00%	28.31%	5.81%	0.39%	-	27.68%

*Periphyton was not recorded during years 3 and 4

Table 3: Continued.....

PSU	Field-Assigned Communities				Species Mean Relative Cover (%)					
	Ridge (%)	Wet Prairie (%)	Slough (%)	Mixed/Edge (%)	<i>Typha spp.</i>	<i>Cladium jamaicense</i>	<i>Nymphaea odorata</i>	<i>Eleocharis spp.</i>	<i>Periphyton</i> *	<i>Utricularia Spp.</i>
50	51.11%	2.96%	11.85%	19.26%	0.74%	42.25%	5.84%	14.60%	-	28.66%
54	45.95%	0.00%	25.23%	25.23%	0.47%	58.38%	0.00%	14.28%	-	20.24%
65	13.22%	29.75%	1.65%	1.65%	0.00%	17.60%	16.61%	7.53%	-	20.12%
66	53.79%	0.00%	29.55%	16.67%	0.03%	43.07%	17.95%	5.06%	-	23.41%
67	56.06%	0.00%	14.39%	12.88%	1.99%	50.66%	0.29%	7.96%	-	0.03%
68	30.15%	43.38%	0.74%	2.21%	0.29%	24.66%	25.71%	8.76%	-	34.52%
69	52.46%	10.66%	1.64%	3.28%	11.23%	62.08%	9.22%	2.52%	-	10.35%
71	42.86%	38.10%	5.56%	0.00%	0.30%	43.70%	17.32%	9.35%	-	7.45%
73	56.30%	0.00%	0.00%	0.00%	29.93%	66.35%	0.15%	0.00%	-	1.62%
79	46.03%	3.17%	15.87%	23.81%	4.05%	51.48%	16.26%	4.77%	-	0.97%

*Periphyton was not recorded during year 5

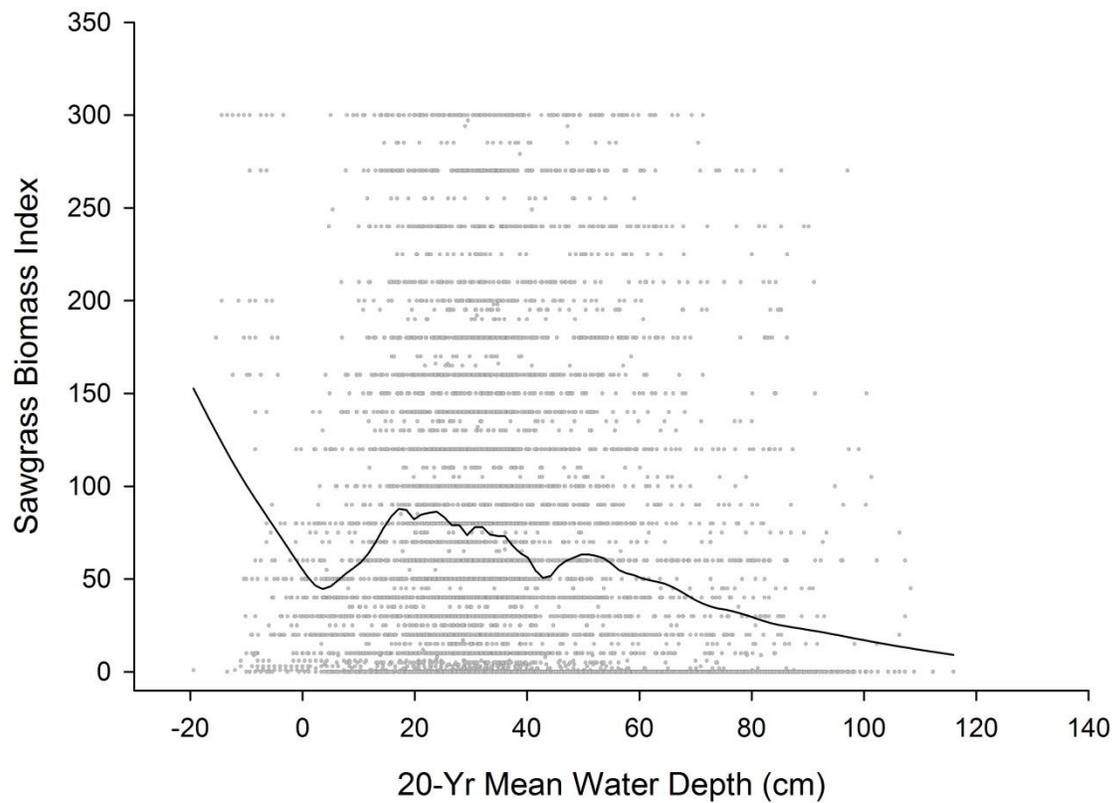


Figure 24: Relationship between sawgrass biomass index (% cover x height category) and 20-Year mean water depth. Line is the LOESS estimate of mean Sawgrass Biomass Index (SBI) as a function of water depth, and indicates peak sawgrass biomass occurs at an elevation of ca. 18 cm.

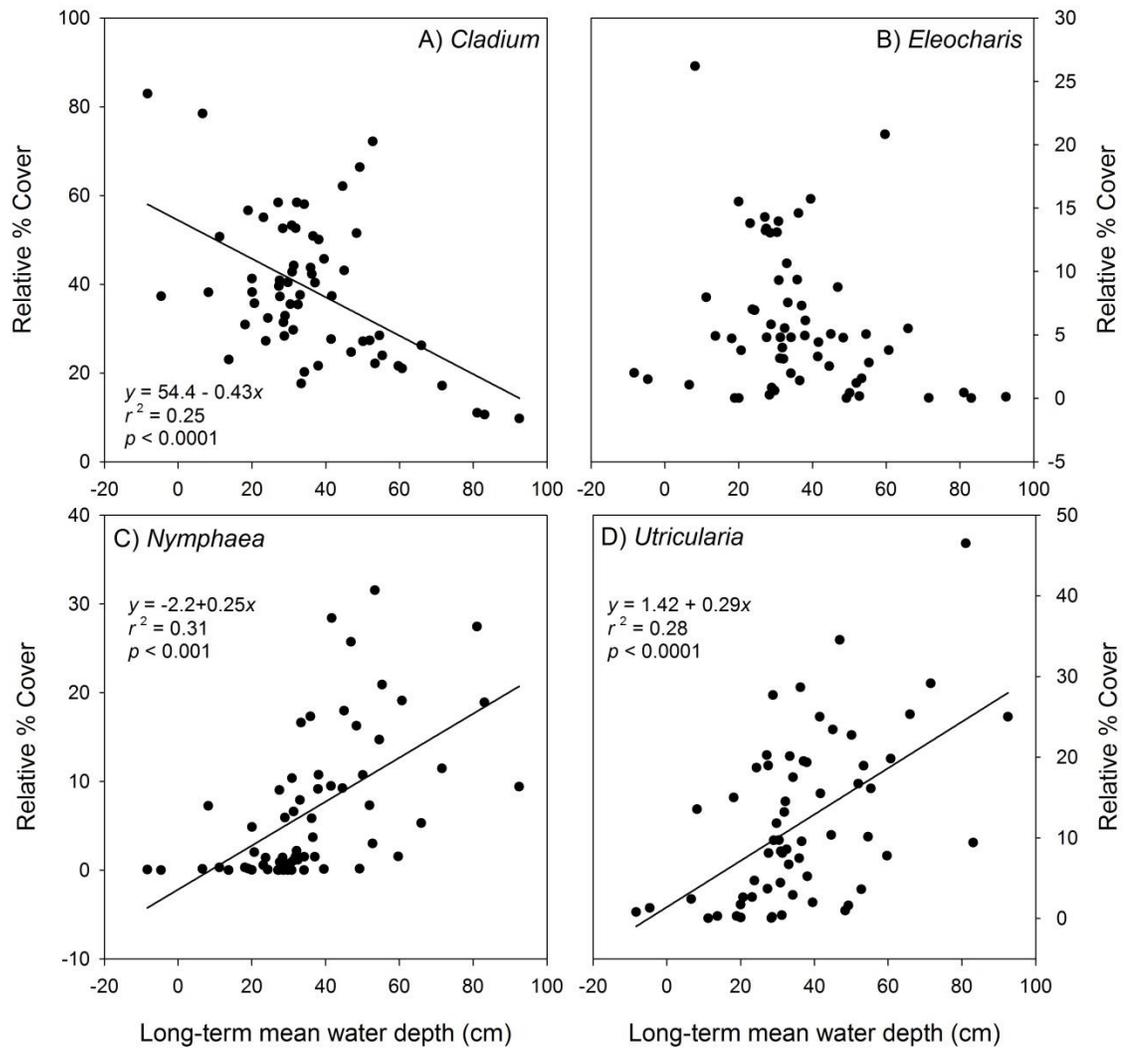


Figure 25: Mean abundance within PSUs of common ridge, slough, and wet prairie species as a function of long-term mean water depth at PSU scale. These data include PSUs outside the historic range of the ridge-slough landscape (i.e. marl prairie areas).

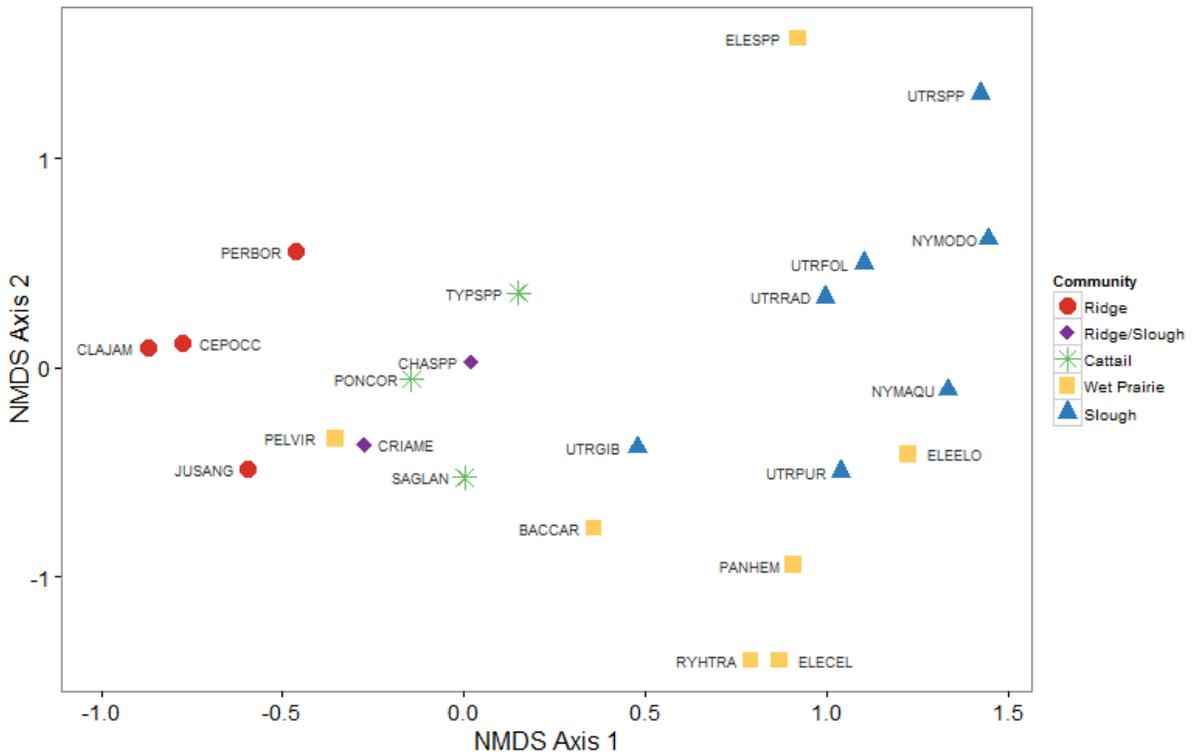


Figure 26: Distribution of the 22 most abundant ridge-slough plant species in ordination space. Note coherent clustering of species by community type, which indicates relatively strong fidelity of species to their associated communities across the landscape.

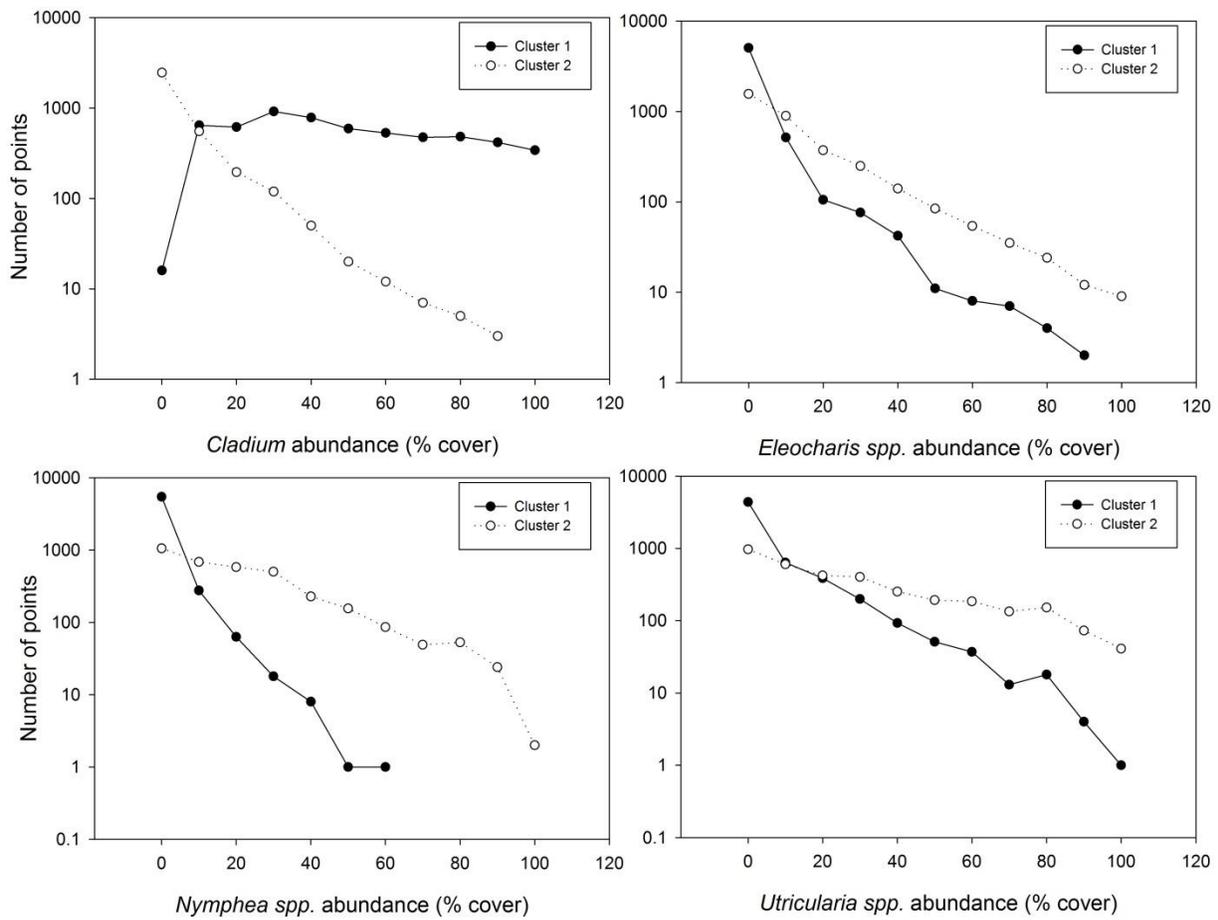


Figure 27: Assignment of sampling points to cluster 1 (Ridge) or cluster 2 (wet prairie/slough) as a function of the abundance (as % cover) of common taxa. Sawgrass (top left) occurs primarily in cluster 1, and the likelihood of a point being assigned to cluster 2 decreases rapidly as sawgrass cover increases. *Eleocharis*, *Nymphaea*, and *Utricularia* show the opposite pattern.

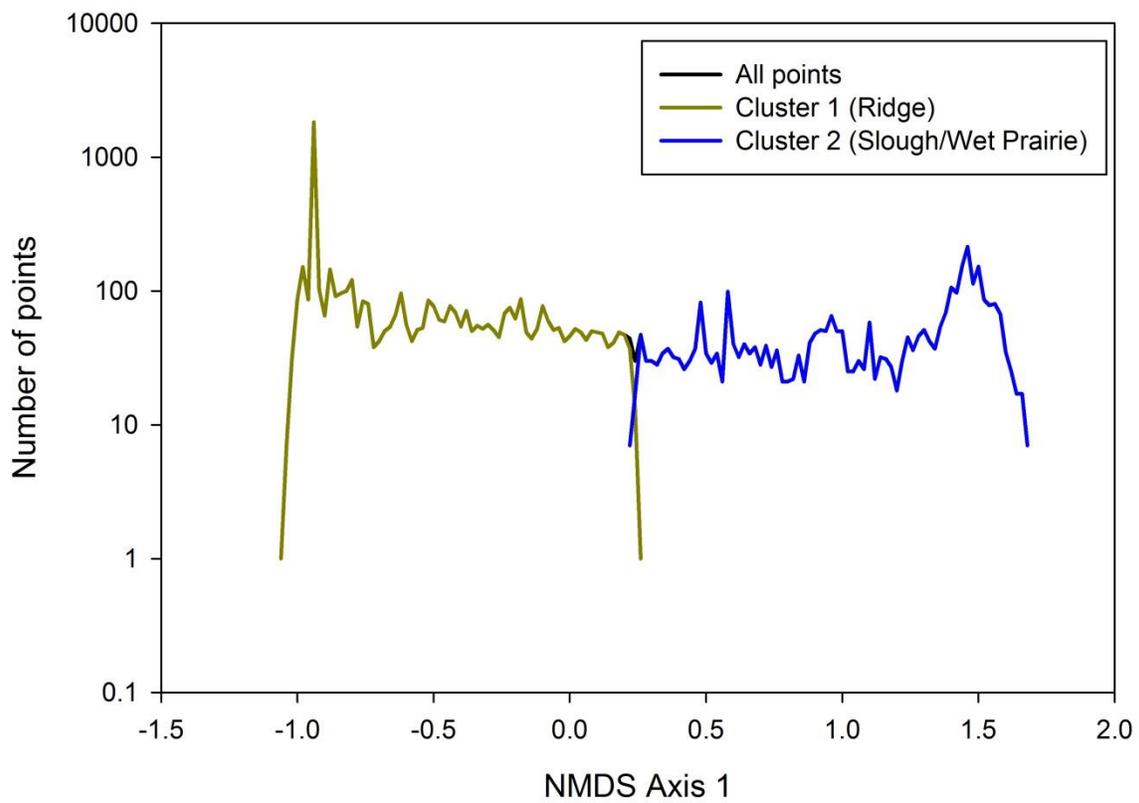


Figure 28: Distribution of all sampling points along NMDS axis 1 (rotated to the dominant hydrologic vector). NMDS axis 1 exerts the strongest influence on cluster membership.

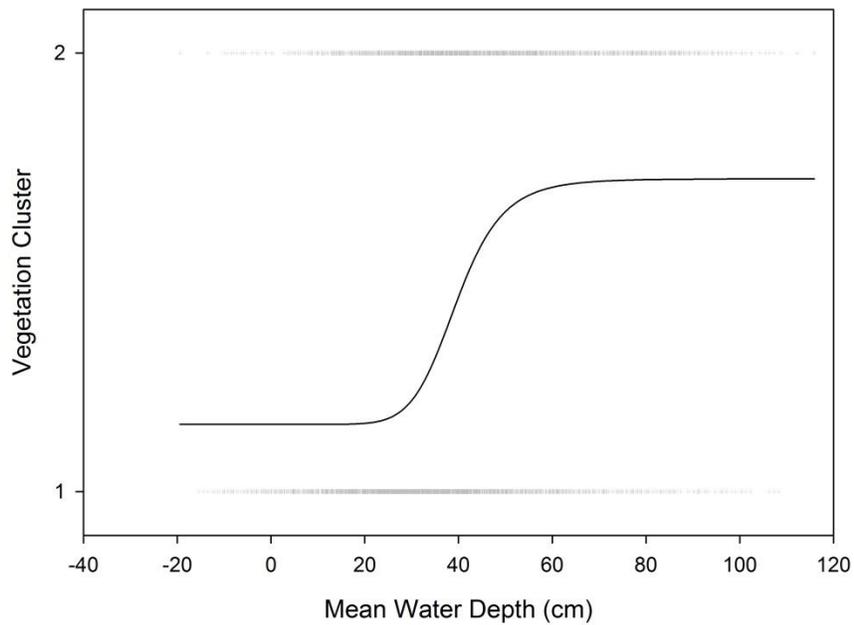
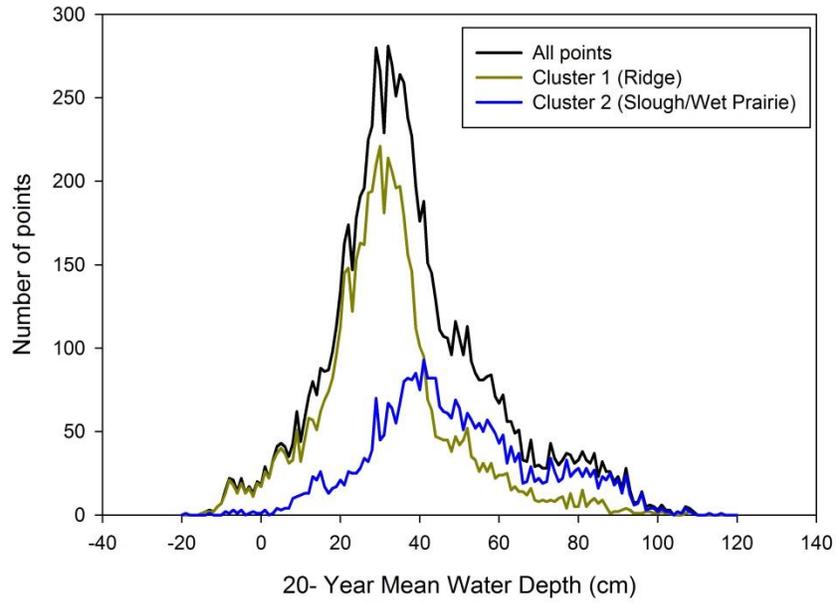


Figure 29: Distribution of all sampling points and points within the 2 clusters (top), and probability of a point being assigned to cluster 1 or cluster 2 (bottom), as a function of mean water depth.

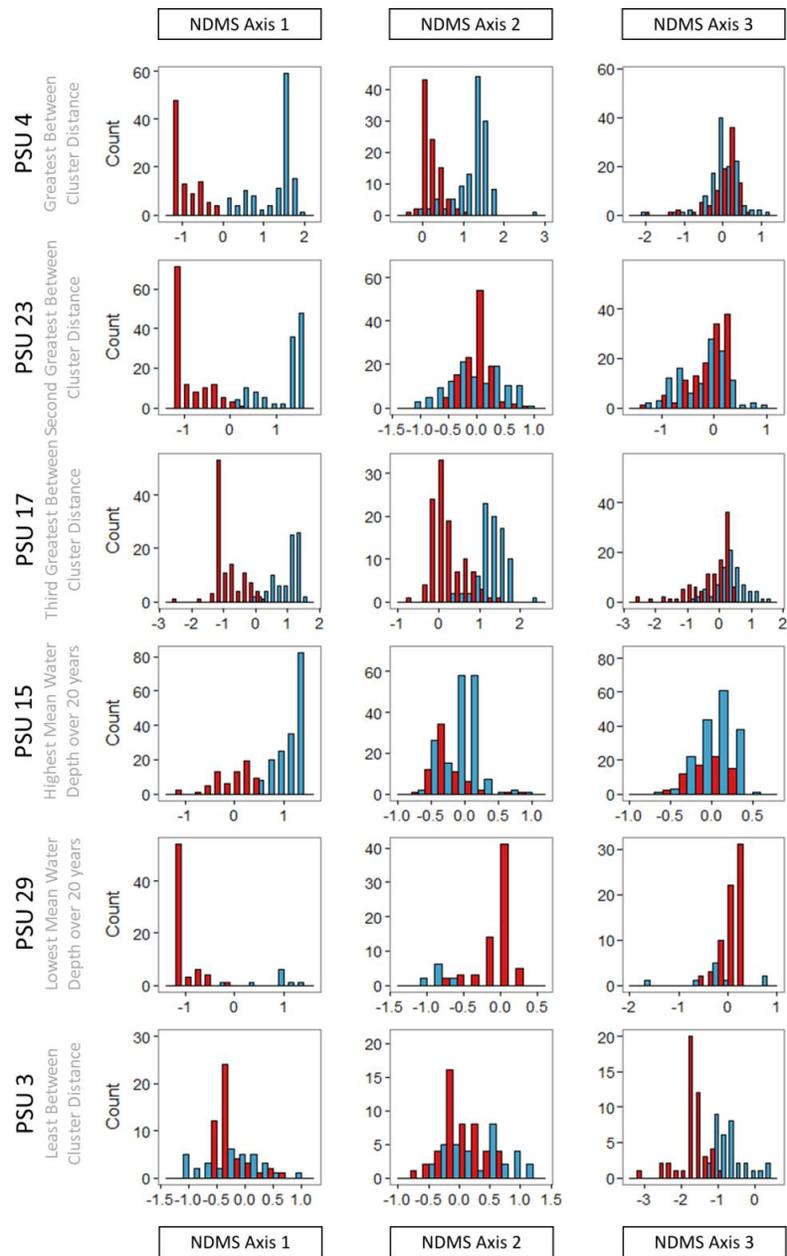


Figure 30: Examples of distribution of sample points in ordination space. In conserved PSUs (4, 17, and 23), vegetation clusters (red and blue bars) are clearly differentiated, and most sample points occur at the extremes of ordination axes (particularly axis 1). In degraded landscapes, clusters are less clearly differentiated, less equally represented, and more commonly occur at intermediate values along ordination axes. These patterns support the inference that shifts in the mean value, rather than changes in variance, account for differences in community cluster separation, and that decreases in community cluster distance capture decreasing distinctiveness of vegetation communities.

Community distinctiveness followed similar geographic patterns to those observed for microtopographic variability (Fig. 31). Only a small fraction of the historic ridge-slough landscape – namely central WCA 3A S – is in a relatively conserved condition reminiscent of the pre-drainage conditions as measured by community distinctiveness. Large sections of ENP, WCA 3B, WCA 3A N, and to a lesser extent WCA 2 and LNWR, are in a more degraded state as defined by this community distinctiveness metric. We note that the scaling of these measures of landscape integrity do not yet allow for direct comparisons of condition as measured by elevation variance and community distinctiveness.

Distance between clusters representing distinctiveness of communities within a landscape was not correlated with long-term mean water depth (Fig. 32), although maximal community distinctness generally occurred within PSUs with long-term mean water depths between 20 and 50 cm. PSUs in ENP, WCA 3A S and WCA 3B clustered relatively closely on both the community distinctiveness and the mean water depth axes; WCA3AN PSUs were notably all indistinct but had very high 20 year mean water depth variability. Variation among PSUs in community distinctiveness followed a skewed shape (Fig. 33) with a long tail represented by the most conserved PSUs.

Measures of local relationships between elevation and community composition corresponded to PSUs with strong community distinctiveness (Table 4). PSUs with high distinctiveness also had higher separation of those communities in water depth; these PSUs also exhibited stronger correlations between point-scale water depth and sawgrass abundance and between local water depth and vegetation community composition (as measured by Mantel's r). The strongest vegetation-elevation relationships were observed at intermediate water depths, generally between 30 and 50 cm (Fig. 34). Spatial distribution of the vegetation-elevation association followed similar patterns to those observed for microtopographic variability and vegetation community distinctness (Fig. 35).

Across PSUs, community distinctiveness increased with microtopographic variation (as measured by water depth standard deviation; Fig. 36[top]). Everglades areas generally clustered in discrete ranges of microtopographic variability and community distinctiveness. PSUs within WCA 3A S generally had both high topographic variability and distinct communities, whereas PSUs within ENP, LNWR, WCA 3B, and WCA 2 largely (but not exclusively) exhibited less topographic variability and reduced community distinctiveness. PSUs with higher long-term mean water depth had greater topographic variability and lower vegetation community cluster distances than those with low (<25 cm) long-term mean water depth (Fig. 36[bottom]).

In prior analyses, bimodal distributions of both community distinctiveness (Fig. 30) and topographic standard deviation (Fig 17) provided one way to partition PSUs into quadrants corresponding to positions along trajectories of degradation (Figs. 5, 36). With the current cluster analysis of the complete data set, neither variable exhibited as clear separation between conserved and degraded conditions.

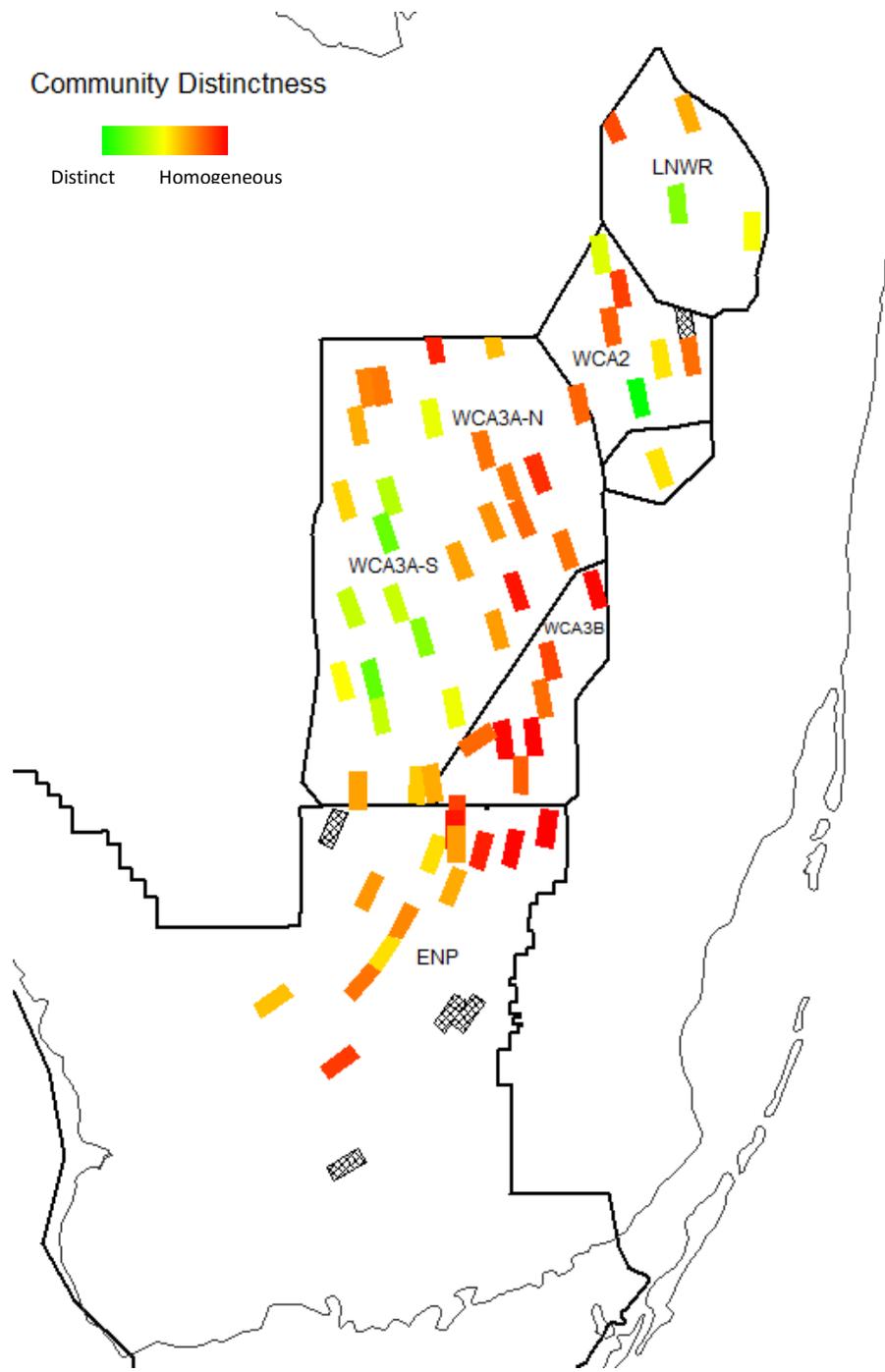


Figure 31: Spatial patterns of vegetation community distinctness across the historic ridge-slough landscape. Colors indicate the separation of vegetation communities (as measured by the distance between midpoints of clusters in NMDS ordination space).

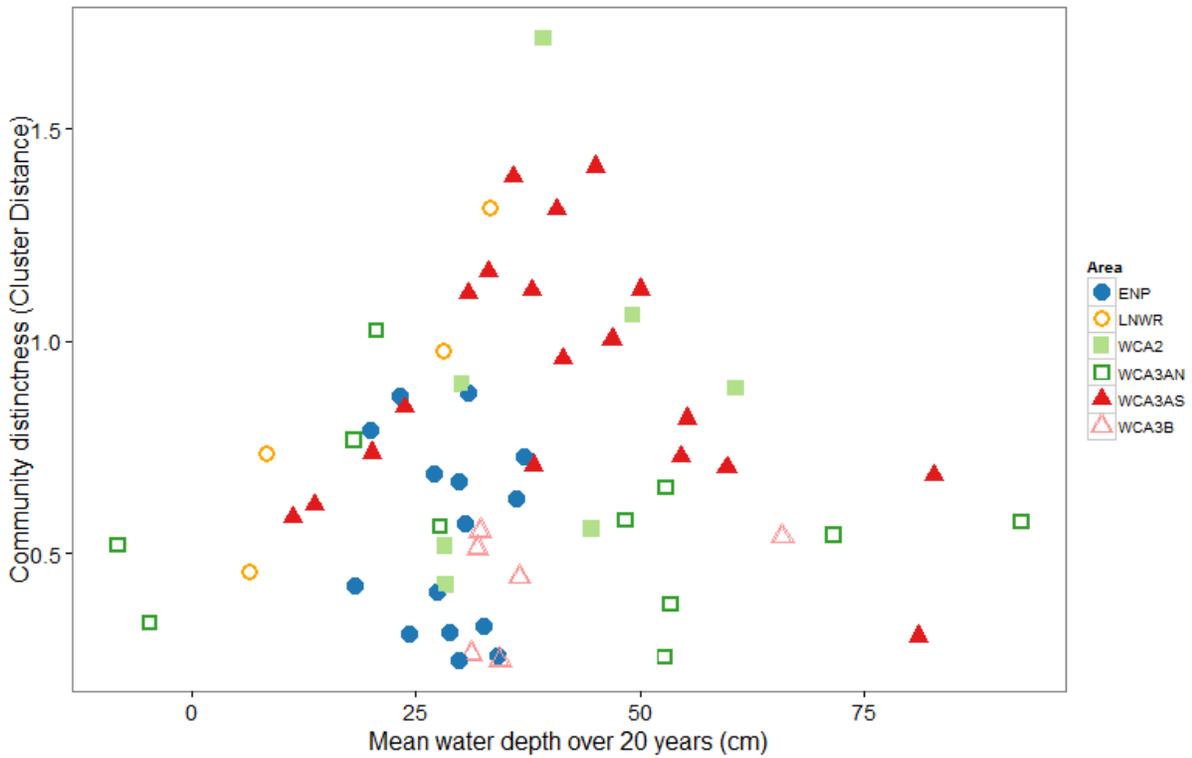


Figure 32: Relationship between hydrologic regime and vegetation community distinctiveness across PSUs within the historic ridge and slough. Symbols represent PSUs in different hydrologic management basins. Differences in community distinctness among PSUs with similar contemporary hydrologic regimes (e.g., ENP and WCA3B vs. WCA3S) may indicate that antecedent hydrologic regime and other past disturbances may continue to shape existing landscape structure.

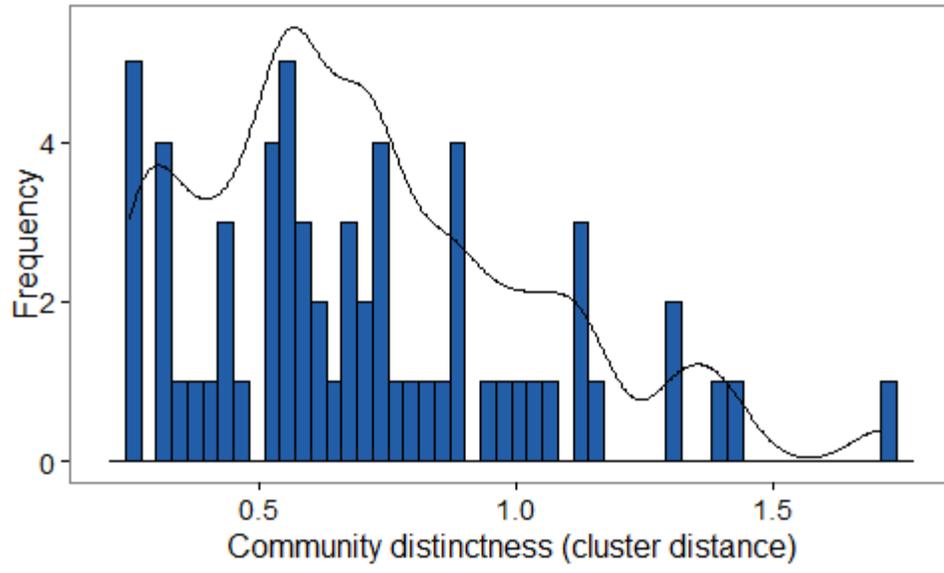


Figure 33: Frequency of community distinctiveness variation across PSUs. Black line indicates the best-fit model of frequency distribution.

Table 4. Measures of co-variation among elevation, hydrologic regime, and vegetation community structure.

PSU	Community Elevation Distributions										Community Distinctness	Elevation-Composition Relationships			
	Cluster 1					Cluster 2						Cluster Distance	k-means WD difference (cm)	Mantel's r	r^2 Cladium-WD
	MWD	StDev	Kurtosis	Skew	%	MWD	StDev	Kurtosis	Skew	%					
0	27.47	4.95	1.66	3.11	71.1%	38.98	5.36	0.27	0.51	28.9%	0.88	11.5	0.52	0.52	
1	4.84	4.54	-0.31	0.42	55.6%	12.51	4.50	-0.58	1.62	44.4%	0.73	7.7	0.33	0.27	
2	42.42	9.72	-0.20	-0.75	62.3%	54.78	8.13	-0.46	0.30	37.7%	1.12	12.4	0.3	0.29	
3	-4.23	3.36	-0.26	-0.89	71.2%	-6.03	3.47	0.63	-0.79	28.8%	0.34	1.8	0.15	0.05	
4	33.83	9.01	0.64	0.31	53.5%	46.84	10.64	-0.37	-0.81	46.5%	1.31	13.0	0.29	0.25	
5	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	
6	27.60	5.32	0.00	-0.08	53.4%	26.97	7.56	0.98	4.46	46.6%	0.41	0.6	0.09	0.01	
7	29.38	6.08	0.88	4.21	50.6%	36.81	4.28	0.46	2.43	49.4%	1.16	7.4	0.33	0.44	
8*	-	-	-	-	-	-	-	-	-	-	-	-	0.24	0.01	
9	28.48	4.61	0.07	-0.85	55.5%	28.23	5.17	0.53	1.88	44.5%	0.425	0.3	0.01	0	
10*	-	-	-	-	-	-	-	-	-	-	-	-	-0.09	0.02	
11	51.79	7.77	0.43	0.87	62.7%	54.12	6.25	-0.32	2.16	37.3%	0.38	2.3	0.08	0.05	
81	31.50	4.99	-1.20	2.52	64.5%	32.47	6.79	-0.02	-0.27	35.5%	0.51	1.0	0.22	0.03	
13	52.48	8.52	-0.40	0.00	83.8%	54.03	6.94	-0.11	0.10	16.2%	0.26	1.5	0.01	0	
14*	-	-	-	-	-	-	-	-	-	-	-	-	0.13	0.04	
15	71.40	10.32	0.03	-0.11	65.5%	71.67	7.78	0.03	0.01	34.5%	0.54	0.3	0.07	0.03	
16*	-	-	-	-	-	-	-	-	-	-	-	-	0.08	0	
17	20.76	5.85	0.10	1.34	56.9%	37.58	13.38	0.62	3.90	43.1%	0.98	16.8	0.51	0.25	
18	28.42	6.79	0.18	-1.04	52.4%	31.06	7.09	-0.19	-1.28	47.6%	0.67	2.6	0.08	0.06	
19	15.77	6.04	0.74	1.78	52.9%	25.68	6.97	-0.18	0.14	47.1%	1.03	9.9	0.33	0.3	
20	30.64	4.89	-0.59	-0.29	81.8%	33.58	5.27	-1.40	2.09	18.2%	0.26	2.9	0.35	0.3	
21	32.26	5.83	1.72	6.17	58.0%	49.36	9.56	0.37	-0.31	42.0%	1.71	17.1	0.44	0.45	
22	21.23	7.79	-0.24	-0.22	51.9%	18.60	5.75	0.32	0.20	48.1%	0.79	2.6	0.11	0.03	
23	23.00	4.76	0.59	0.12	54.4%	40.31	6.60	-0.38	-0.95	45.6%	1.11	17.3	0.68	0.61	
24	34.91	5.88	-0.72	0.77	78.3%	31.72	6.84	0.13	-0.82	21.7%	0.26	3.2	0.1	0.03	
25	6.40	5.37	-0.38	0.60	85.0%	6.77	10.87	1.04	0.74	15.0%	0.46	0.4	0.21	0	
26	33.37	6.55	0.27	3.17	55.0%	48.11	8.97	-0.89	0.96	45.0%	0.96	14.7	0.45	0.32	
27	13.84	10.98	0.40	-0.67	79.5%	34.23	7.54	-2.22	6.67	20.5%	0.77	20.4	0.45	0.37	
28	31.70	5.26	-0.40	0.33	79.7%	33.93	5.04	-0.18	-0.90	20.3%	0.55	2.2	0.02	0.03	
29	-8.13	2.78	-0.48	-0.04	89.6%	-9.64	4.32	-1.63	2.41	10.4%	0.52	1.5	-0.01	0.01	
30	20.25	7.38	-0.27	-0.49	64.9%	28.54	8.73	-0.33	-0.23	35.1%	0.87	8.3	0.26	0.05	
31	38.99	6.31	0.09	-0.74	57.3%	36.75	7.00	0.63	0.78	42.7%	0.71	2.2	0.15	0.07	

* Elevation mode analysis was not performed on PSUs outside the historic distribution of the ridge-slough landscape.

Table 4: continued

PSU	Community Elevation Distributions										Community Distinctness	Elevation-Composition Relationships			
	Cluster 1					Cluster 2						Cluster Distance	k-means WD difference (cm)	Mantel's r	r^2 Cladium-WD
	MWD	StDev	Kurtosis	Skew	%	MWD	StDev	Kurtosis	Skew	%					
32	34.96	5.68	-0.42	0.00	67.6%	41.31	5.29	0.18	-0.82	32.4%	0.73	6.4	0.174	0.156	
34	48.16	12.01	-1.60	7.27	58.0%	60.55	10.91	-0.11	-0.33	42.0%	0.82	12.4	0.252	0.085	
35	14.52	4.54	-0.06	-1.31	56.7%	17.91	19.03	3.51	12.76	43.3%	0.61	3.4	-0.003	0.029	
36	78.05	9.41	-0.26	-0.89	73.2%	84.57	6.00	-0.65	-0.25	26.8%	0.68	6.5	0.114	0.063	
37	26.39	7.76	0.74	-0.34	73.5%	40.68	8.13	-0.61	0.00	26.5%	0.90	14.3	0.382	0.21	
39	19.66	5.05	0.95	2.95	57.3%	26.82	5.99	-0.55	0.01	42.7%	0.84	7.2	0.29	0.299	
43	27.25	5.26	0.32	-0.26	66.7%	28.34	5.32	0.34	-0.50	33.3%	0.56	1.1	0.08	0.04	
44	34.22	4.42	-0.94	2.17	72.7%	34.85	3.82	-0.52	0.40	27.3%	0.25	0.6	-0.037	0	
45	94.06	8.80	0.04	-1.00	63.0%	89.97	7.97	-1.31	3.20	37.0%	0.58	4.1	0.123	0.04	
47	48.57	8.67	1.25	1.67	51.8%	57.34	10.93	-0.24	0.04	48.2%	0.66	8.8	0.228	0.105	
513	29.95	5.64	-0.50	0.71	54.3%	29.78	4.48	-0.15	-0.10	45.7%	0.25	0.2	0.082	0.029	
DPM	63.21	13.87	-0.09	-0.72	74.3%	73.76	12.54	-0.64	-0.86	25.7%	0.54	10.5	0.067	0.118	
51	21.34	6.45	-0.11	-0.72	76.3%	16.41	6.50	0.54	0.81	23.7%	0.74	4.9	0.0993	0.024	
52	48.04	14.32	-1.19	5.58	59.0%	63.96	12.34	-0.09	-0.20	41.0%	0.73	15.9	0.276	0.110	
53	51.55	11.94	0.92	0.23	61.2%	66.54	11.42	-0.16	-0.76	38.8%	0.89	15.0	0.254	0.196	
55	30.25	7.69	0.23	-1.21	57.3%	43.71	5.65	0.24	-0.66	42.7%	1.12	13.5	0.451	0.481	
56	33.65	6.67	-0.17	-0.21	59.3%	31.07	10.50	0.57	1.70	40.7%	0.33	2.6	0.197	0.003	
58	51.60	9.08	-0.11	-1.05	56.4%	65.98	7.60	0.15	0.00	43.6%	0.70	14.4	0.349	0.254	
61	26.20	5.66	-1.13	2.70	65.1%	31.99	3.99	1.06	2.62	34.9%	0.52	5.8	0.215	0.069	
62	29.02	5.62	-0.66	-0.21	58.5%	32.39	8.66	-0.29	-0.45	41.5%	0.57	3.4	0.272	0.111	
63	76.24	8.01	-0.86	-0.34	70.4%	83.13	5.89	0.30	1.26	29.6%	0.30	6.9	0.149	0.095	
220	36.35	4.43	-0.15	1.20	70.6%	37.07	3.12	0.27	0.91	29.4%	0.44	0.7	-0.074	0.001	
BS1	16.63	6.35	0.07	1.23	70.0%	21.36	9.55	0.42	0.39	30.0%	0.42	4.7	0.284	0.036	
BS2	22.93	8.81	-0.12	-0.36	68.2%	28.82	9.49	0.83	2.09	31.8%	0.31	5.9	0.140	0.037	
BS3	27.08	6.86	-0.61	1.56	65.9%	29.68	10.17	1.01	5.63	34.1%	0.31	2.6	0.341	0.057	

Table 4: continued.

PSU	Community Elevation Distributions										Community Distinctness	Elevation-Composition Relationships			
	Cluster 1					Cluster 2						Cluster Distance	k-means WD difference (cm)	Mantel's r	r^2 Cladium-WD
	MWD	StDev	Kurtosis	Skew	%	MWD	StDev	Kurtosis	Skew	%					
50	32.79	7.81	-0.34	-0.27	74.1%	46.1	7.1	-0.2	-0.1	25.9%	0.63	13.3	0.330	0.238	
54	25.22	7.24	0.36	0.24	50.5%	29.0	8.8	-0.3	-1.3	49.5%	0.69	3.8	0.147	0.020	
65	25.05	12.78	0.47	-0.24	57.0%	39.6	13.1	0.4	1.1	43.0%	1.31	14.6	0.247	0.121	
66	36.44	5.96	-0.09	-0.47	51.5%	54.2	6.8	-0.8	0.4	48.5%	1.41	17.7	0.611	0.444	
67	14.63	4.44	-0.80	1.87	54.5%	8.4	6.8	0.7	0.8	45.5%	0.58	6.2	0.170	0.060	
68	39.11	16.74	-0.06	-0.65	64.7%	51.1	14.4	0.1	-0.6	35.3%	1.00	12.0	0.080	0.216	
69	43.84	9.32	0.14	-0.04	77.9%	47.3	8.1	0.2	-0.4	22.1%	0.56	3.4	0.088	0.028	
71	29.59	5.13	0.70	1.11	53.2%	41.4	3.3	-0.6	1.8	46.8%	1.39	11.8	0.609	0.660	
73	34.59	10.30	2.48	9.11	70.4%	85.4	16.8	-1.3	1.4	29.6%	1.06	50.8	0.710	0.612	
79	49.37	6.07	-0.23	0.06	69.0%	46.1	6.4	0.0	-0.2	31.0%	0.58	3.3	0.180	0.086	

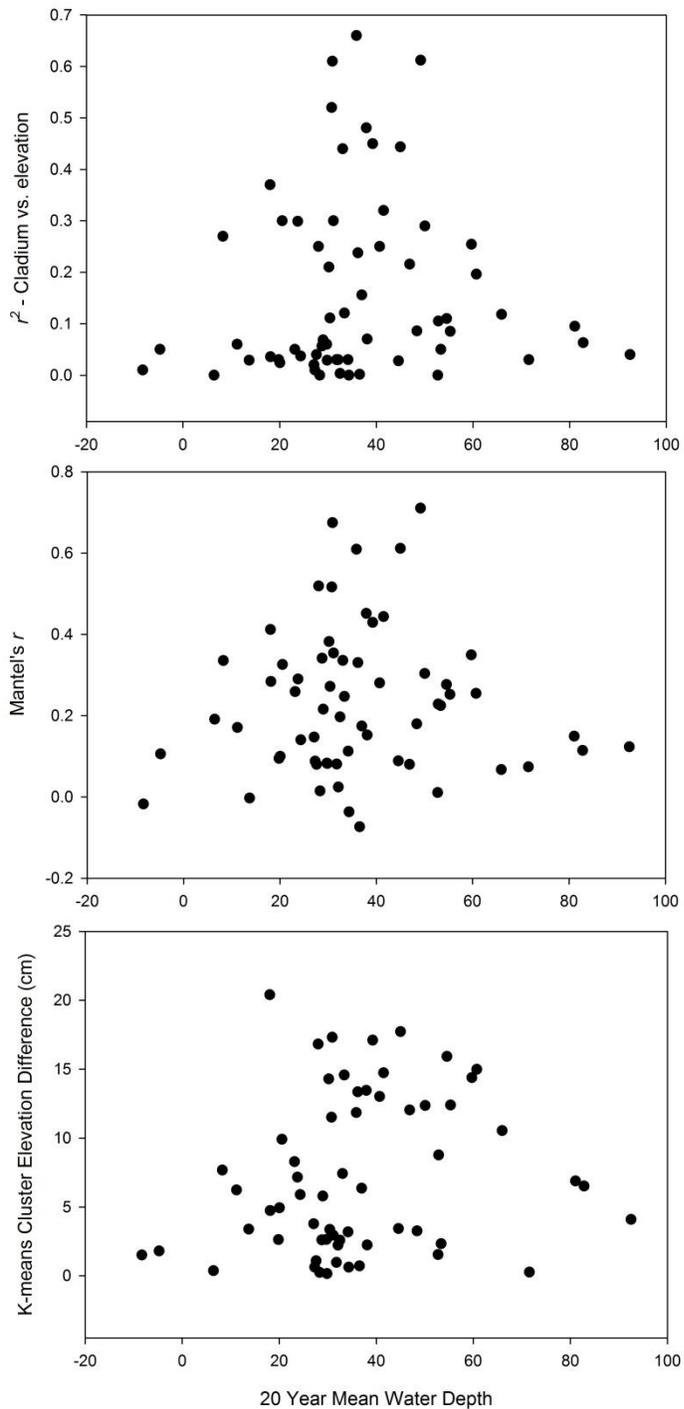


Figure 34: Relationships between long-term mean water depth (calculated at PSU scale) and 3 measures of the strength of elevation-vegetation associations (correlation between sawgrass abundance and elevation (top); Mantel's correlation coefficient [r], which is the multivariate equivalent of Pearson's correlation coefficient (middle); and the elevation difference between the two ordination clusters).

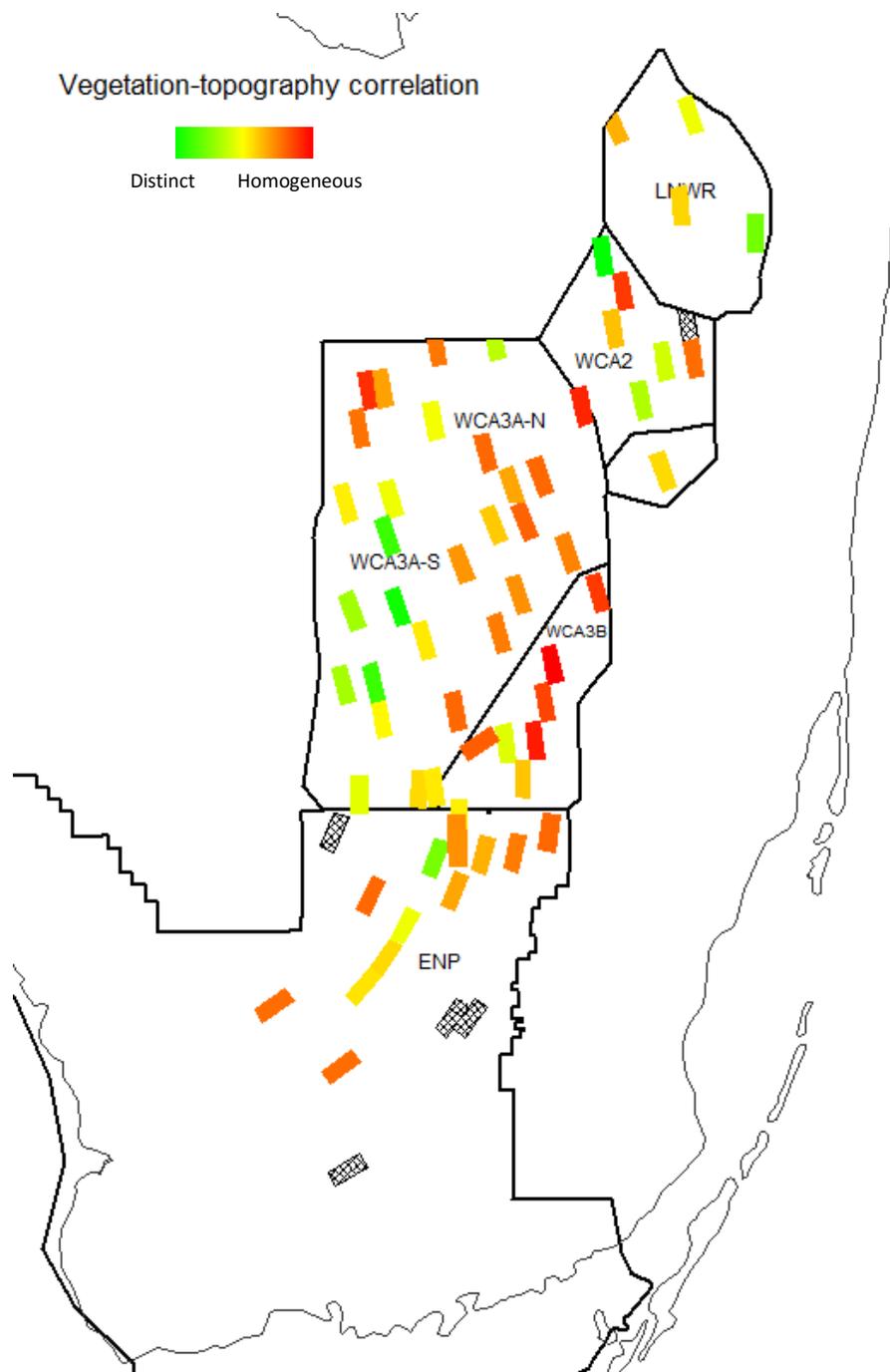


Figure 35: Spatial patterns of elevation-vegetation associations (as measured by Mantel's correlation coefficient [r], which is the multivariate equivalent of Pearson's correlation coefficient). Colors indicate the strength of association between elevation and vegetation community structure. Hatching indicates PSUs that were not sampled (PSU 5 in WCA 2) or that were excluded from the analysis because they lie outside the historic distribution of the ridge-slough landscape (PSUs 8, 10, 14, 16 in ENP).

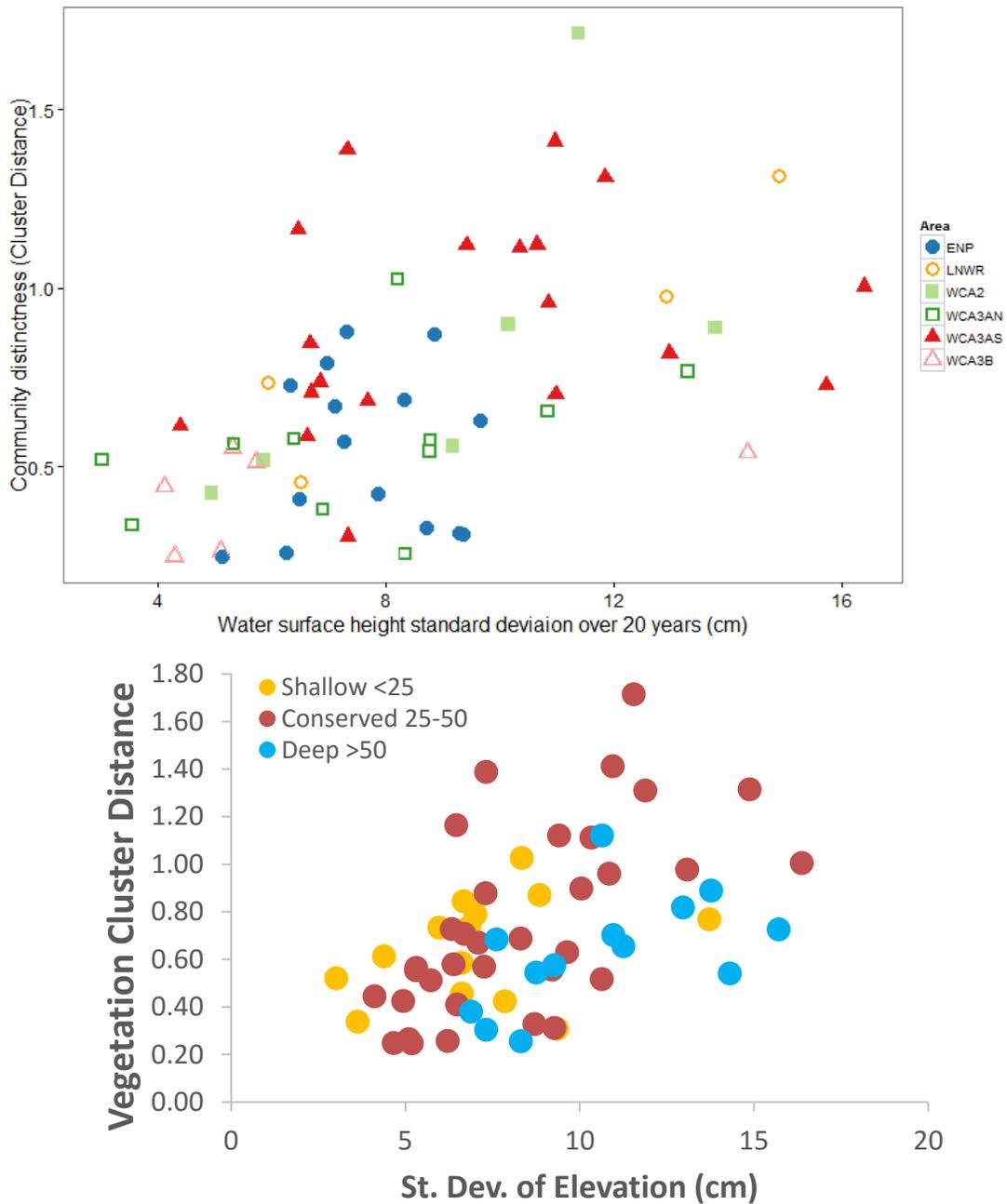


Figure 36: Relationship between microtopographic variability (as measured by standard deviation of elevation within each PSU) and vegetation community distinctiveness (as measured by distance separating clusters in ordinations space). Top: symbols indicate Everglades Basin. Bottom: Symbols indicate long-term mean water depth.

DISCUSSION

The results of this monitoring study support and expand on current understanding of landscape pattern in the ridge-slough-tree island mosaic, its responses to hydrologic regime, and the spatial distribution of its degradation and persistence. Microtopographic structure and vegetation structure exhibited broadly similar geographic patterns (Fig. 19, 31); landscapes in southern and central WCA3AS had greater topographic variability, and bi-modally distributed soil elevations that maintained distinct ridge and slough communities. These features were less prevalent in areas that have or are currently experiencing hydrologic alteration. Distributions of soil elevation variance (Fig. 20), community distinctiveness (Fig. 33), and the strength of elevation-vegetation associations provide tentative support for the hypothesis that the historic patterned structure and degraded landscape patterns represent alternative stable states (Scheffer and Carpenter 2003).

Topographic measures of ridge-slough landscape condition

The microtopographic differentiation of ridges and sloughs is a fundamental feature of the historic ridge-slough landscape (SCT 2003, McVoy et al. 2011, USACE and SFWMD 2000). This differentiation is widely hypothesized to reflect local feedbacks among elevation, hydrologic regime plant community productivity and composition, and peat production and decomposition (Larsen et al. 2007, Cohen et al. 2011), although a variety of mechanisms have been proposed to account for the spatial patterning of these patch types. To date, our monitoring efforts have focused on two measures of microtopographic differentiation: elevation variance and the occurrence of distinct elevation modes. These measures are related in that the feedbacks hypothesized to create distinct elevation modes should also promote greater overall variation in elevation. The empirical relationship between the standard deviation of elevation and the difference between elevation modes (Fig. 22) supports this hypothesis, as the greatest elevation variance occurred only in PSUs with distinct elevation modes, and specifically among the PSUs with the greatest differences between elevation modes. However, it is also clear that the mere occurrence of bi-modality is insufficient to ensure strongly differentiated elevation modes, as some PSUs with bi-modal elevation distributions had relatively low overall elevation variance (Fig. 22).

The differences between elevation modes we observed in conserved areas are consistent with previous measurements in central and southern WCA3AS (Watts et al. 2010). In this study, elevation mode differences ranged from as low as 9 to 24 cm, with one cattail-dominated PSU exhibiting a mode difference of 54 cm. Most elevation mode differences were similar to the range observed by Watts et al (2010) in a smaller number of landscape blocks. Historical estimates of elevation differences between ridges and sloughs are generally higher, ranging from 30 to 60 cm, and in some cases as great as 90 cm (McVoy et al. 2011). A direct comparison of these historic and contemporary estimates suggests that considerable deflation of ridge-slough topography has occurred even in conserved areas. However, such a direct comparison should be undertaken with caution. Historic ridge-slough elevation differences and our estimates of elevation mode differences is that these measure subtly different things. Historic ridge-slough measurements such as those reported in McVoy et al. (2011) measure elevation differences between the highest point in a ridge and the lowest point in an adjacent slough. Such isolated observations may be biased toward the greatest elevation differences, and certainly do not represent a random sampling of

locations. In contrast, the elevation mode differences reported here are based on spatially-integrated means of ridges and sloughs over entire landscapes, and would not be biased toward higher or lower elevations within ridges and sloughs, respectively, or toward locations with particularly distinct elevations. As such, comparison of historic measurements of elevation differences may somewhat overestimate the incongruity of these measurements. However, the relationship between elevation mode difference and long-term mean water depth across PSUs (Fig. 21) supports the underlying hypothesis of McVoy and co-authors that ridge-slough elevation differences are sensitive to hydrologic regime, and that elevation differences may have decreased even in relatively conserved areas in response to recent anthropogenic hydrologic alteration.

Vegetation structure as a measure of ridge-slough condition

The historic ridge-slough mosaic was characterized by the distinct zonation of plant species whose distributions were shaped by abrupt differences in elevation between ridges and sloughs (Ogden 2005, McVoy et al. 2011). The distinct sawgrass-dominated ridges and *Nymphaea*- and *Utricularia*-dominated sloughs observed in conserved landscapes of WCA3AS are consistent with these previous findings. While a number of studies have documented shifts in community composition in response to changing hydrologic regimes at relatively broad scales (Givnish et al. 2008, Zweig and Kitchens 2008, Zweig and Kitchens 2009, Todd et al. 2010, Foti et al. 2012, Todd et al. 2012), this study is the first to systematically document those differences at the scale of the entire Greater Everglades ecosystem, and to do so based on field observations. We found that high mean long-term water levels were associated with relatively low abundance of sawgrass, and more strongly with high abundances of *Nymphaea* and *Utricularia* (Fig. 25). *Eleocharis cellulosa*, a characteristic wet prairie species, was most abundant in PSUs with intermediate long-term water depths. These patterns support the conclusions of a number of studies that hydrologic regime shapes plant species composition at broad spatial scales across the Greater Everglades ecosystem. However, considerable variation within and among PSUs in the abundance of individual taxa suggest that the simple prevalence of indicator species, at least as measured in this study, may be a relatively poor measure of landscape condition.

In addition to these compositional shifts, we observed declines in the distinctness of local plant communities. Previous studies have indicated that hydrologic alteration promotes the blending of ridge and slough communities, as well as increased prevalence of species characteristic of intermediate elevations (e.g., *Eleocharis* spp.). Our landscape-scale assessment confirms that in conserved landscape such as those found in central and southern WCA3AS, local vegetation communities are highly distinct (Fig. 31, 35). In areas subject to increased or decreased water levels by hydrologic alteration, this distinctness is reduced.

Our approach to measuring community distinctness is a newly developed measure based on measurements of distances between two artificially imposed clusters of plant communities in ordination space. To ensure that this measure is a reasonable proxy for the distinctness of ridge and slough communities, we assessed the prevalence of sawgrass and other taxa in each of the two clusters (Fig 27). The close agreement of global and local cluster assignments indicates that cluster distance within PSUs is in fact a reasonable proxy for plant community distinctness. More sophisticated descriptions of plant communities similarly supported use of cluster distance as a measure of community distinctness within PSUs (Fig. 28-30). In conserved landscapes, most

sampled plant communities occupied one or the other extreme of NMDS axis 1, whereas in degraded landscapes many local plant assemblages had intermediate values. These patterns are consistent with the positions of sawgrass and characteristic deepwater slough species at opposite ends of NMDS axis 1 in our ordination.

Vegetation-elevation relationships as a measure of ridge-slough condition

Conceptual models and empirical observations of the ridge-slough landscape suggest that strong relationships between microtopography and vegetation community structure were characteristic of the historic and conserved ridge-slough landscape (Larsen et al. 2011, McVoy et al. 2011). We used three approaches to characterize how the strength of this association varied across the historic ridge-slough landscape (Table 4). The first and simplest of these measures is the correlation coefficient between elevation and sawgrass abundance within a PSU. The second of these uses Mantel's r to measure the association between elevation and overall community composition. The third of these measures the difference in mean elevation between points assigned to each of the vegetation community clusters that we used to measure community distinctness. As for isolated measures of microtopographic and vegetation community structure, vegetation-elevation relationships were strongest in areas of conserved condition, principally in central and southern WCA3AS (Fig. 35). Individual PSUs within other Everglades basins also exhibited strong associations between elevation and community composition.

Geostatistical and geospatial measures of ridge-slough landscape structure

The historic ridge-slough landscape and current conserved portions are characterized not simply by topographic heterogeneity and distinct vegetation communities, but by a characteristic flow-parallel patterning of ridges and sloughs. A variety of metrics have been proposed to measure that spatial structure from aerial photographs and other remotely sensed measurements (Wu et al. 2006, Foti et al. 2012, Larsen et al. 2012; Casey 2015). These include measurements of ridge geometry, slough connectivity, and the relative abundance of these patch types. The maps generated as part of this study could provide a more spatially and taxonomically resolved data set on which to base such measurements, but current approaches largely assume 2-3 patch types in their assessments of spatial pattern (Wu et al. 1997, Wu et al. 2006, Nungesser 2011). A strength of these approaches is that historic aerial imagery is available for some portions of the Everglades, allowing for more robust analysis of long-term trends (Nungesser 2011). However, because they use vegetation structure to assess overall landscape condition, these geospatial measures cannot evaluate changes to microtopographic structure that may in some cases precede degradation of vegetation structure.

Watts et al (2010) present several geostatistical measures of microtopographic structure based on field measurements of soil elevation similar to those reported for this study. One such measure is the spatial autocorrelation of elevation, which in patterned landscapes is predicted not simply to decrease with distance but to become negative at some distance. Such patterns are thought to reflect the operation of negative feedbacks at distance, which are necessary for the formation of regular spatial pattern (Rietkerk and Van de Koppel 2008). A second measure is the degree of spatial structure of elevation based on sill and nugget variance measured by semi-variance analysis. This measure describes the extent to which heterogeneity is structured in space,

as opposed to randomly-distributed. Conserved ridge-slough landscapes would be expected to exhibit strong spatial structure (Watts et al. 2010). A third measure, anisotropy, measures the directionality of elevation based on the semi-variance observed for points oriented perpendicular or parallel to some axis (in this case, the orientation of historic flow). Conserved ridge-slough landscapes would be expected to exhibit strong anisotropy.

Hydrologic conditions supporting persistence of conserved conditions

The current distribution of conserved microtopographic and vegetative conditions provides some information about the hydrologic regimes that support the persistence of relatively intact ridge-slough landscapes, and the conditions that promote degradation of landscape structure. Conserved microtopographic structure, as inferred from bi-modal elevation distributions, was observed in PSUs with long-term mean water depths between 25 and 51 cm. The difference between elevation modes was relatively small in most PSUs with long term mean water depths less than 35 cm, and a number of PSUs with long-term mean water depths less than 40 cm did not exhibit bi-modal distributions. At long-term mean water depths less than 25 and greater than 51 cm, we did not observe PSUs with distinct elevation modes (with a single exception in PSU 3, where the difference between elevation modes was <5 cm, and given the extremely dry conditions, probably represents a statistical artefact rather than a truly bi-modal distribution). In a previous study, Watts et al (2010) found bi-modal elevation distributions in PSUs with long-term mean water depths that ranged from 18 to 65 cm. In all, our data suggest that the preservation of microtopographic differentiation of ridges and sloughs is best maintained by long-term mean water depths between 40 and 50 cm, but that microtopographic structure sometimes resists degradation at water levels as low as 25 cm.

Using vegetation community distinctness yields similar estimates of hydrologic regimes that support relatively conserved landscape structure. Communities with distinct clusters in ordination space occurred in PSUs with long-term mean water depths ranging from ca. 15 cm to ca. 51 cm. The dry extreme of this range represents a single PSU within Loxahatchee National Wildlife Refuge, whose vegetation communities are not necessarily comparable to those of other areas. Within the other water conservation areas, we observed conserved vegetation structure at long-term mean water levels greater than 20 cm. As was observed for microtopographic structure, the drier end of this range (between 20 and 35 cm) also included PSUs with indistinct communities indicative of degraded conditions. All PSUs with long-term mean water depths between 35 and 50 cm had well-differentiated ridge and slough communities.

Previous studies have demonstrated that Everglades plant species respond to diverse characteristics of hydrologic regime that are not captured by the relative coarse metric of long-term mean water depth (Givnish et al. 2008, Zweig and Kitchens 2009). For example, the relative abundance of slough and wet prairie species can shift in response to seasonal and interannual variation in water level. However, given the topographic variation within PSUs, measures of local hydrologic regime cannot easily be scaled to broader landscapes. Moreover, the frequency of hydrologic extremes, especially drying, are likely to covary strongly with long-term mean water level. Finally, long-term mean water depth provides a relatively simple measure of hydrologic regime that may serve as a tractable management target. While future analyses will assess whether other measures of hydrologic regime help explain variation in landscape condition, we contend

that long-term mean water depth provides the most useful measure of hydrologic conditions. Based on the general agreement between microtopographic and vegetation community structure as measures of ridge-slough landscape condition, **we recommend a range of long-term mean water depths between 35 and 50 cm as supportive of relatively intact ridge-slough landscapes.** Even in the fifth year of this monitoring effort, new data has continuously reinforced, rather than undermined, this fundamental conclusion.

Spatial distribution of conserved and degraded ridge-slough landscape conditions

The results of this study support previous conclusions (Wu et al. 2006, Nungesser 2011, McVoy et al. 2011) that historic ridge-slough landscape conditions are best conserved in central WCA3AS. PSUs in that area (specifically, PSUs 2, 4, 23, 26, 66) are characterized by distinct ridge and slough communities that are well sorted along widely separated elevation modes. Microtopographic and vegetation structure and their covariation also indicated moderately conserved conditions in isolated portions of other water management basins, including southeastern WCA1 (17) and southern WCA2 (21). Two PSUs (27 in northern WCA3AN, 73 in northern WCA2) superficially exhibit conserved microtopographic structure and community distinctiveness; however, these characteristics appear to reflect a severe recent fire regime and significant invasion (and perhaps creation) of deep water areas by dense stands of *Typha*, rather than the persistence of historic ridge slough structure.

Microtopographic and vegetation structure and their covariation were also generally in agreement in their identification of highly degraded landscapes. PSUs throughout WCA 3B exhibited degraded conditions by most measures, as did PSUs in northern portions of WCA1, WCA2, and most peripheral areas of WCA3AN, WCA3AS and ENP.

Microtopographic and vegetative measures of landscape condition were generally in agreement, with the largest number of PSUs exhibiting degradation of both characteristics, and a smaller group exhibiting conservation of both characteristics. However, in many PSUs, measures of topographic and vegetation structure diverged in their assessment of conserved and degraded conditions. In PSUs with low long-term mean water depth (<25 cm), we observed relative distinct vegetation communities occupying landscapes with reduced elevation variance; in PSUs with high long-term mean water depths (>50 cm) we observed relatively indistinct vegetation communities despite persistent high microtopographic relief. One important conclusion to be drawn from this pattern is that concurrent monitoring of both vegetation and microtopographic structure is essential for a comprehensive assessment of ridge-slough condition. Remote assessment of changes in patch structure and landscape pattern provides important information about landscape change, but our data support and strengthen the inference of Watts et al. (2010) that **microtopographic changes are leading indicators of change in vegetation composition and structure for drained landscapes. In contrast, vegetation distinctness appears to serve as a leading indicator of degradation associated with impounded conditions.** We conclude that PSUs with degraded topography but intact vegetation and vice versa are likely to be undergoing transitions to fully degraded states.

Special sampling regions of specific relevance to Everglades Restoration

The landscape block associated with the DECOMP Physical Model has a number of noteworthy characteristics. In terms of hydrologic regime, the compartment in which the DPM has been implemented has experienced a relatively high mean water level over the past two decades. Despite, or perhaps because of, these hydrologic conditions, the DPM footprint exhibits a remarkable degree of topographic heterogeneity. In terms of absolute relief, it is more varied than any other measured area, and so may serve as a useful model of hydrodynamics under elevated flow velocities. What distinguishes the DPM footprint from other portions of the Everglades with persistent ridge-slough elevation differences is the relative indistinctness of its vegetation community. In PSUs with comparable relief, ridge and slough communities are easily distinguishable, but our measures suggest that intermingling of these communities occurs to an unusual degree within the DPM footprint. Understanding the characteristics of the DPM footprint relative to other portions of the ridge-slough landscape may help guide interpretation of data collected in response to flow re-establishment, and apply data appropriately to other parts of the landscape.

The landscape block located downstream of the 1-mile bridge (PSU 513) shares many characteristics of other sampled sections of Everglades National Park. This area has experienced moderate water levels over the past 20 years, though anecdotal evidence indicates that much drier regimes may have preceded the establishment of ENP. Topographic relief is relatively muted, and, by our measures, vegetation communities are relatively indistinct. Sloughs have largely been invaded by Wet Prairie species such as *Eleocharis*. Because of these characteristics, PSU 513 can serve as an appropriate tool for assessment of the ecological and topographic response to sheetflow restoration. Unfortunately, budget constraints and logistical and permitting challenges have limited data collection from ENP during the Year 3 period (and since). As a result, assessment of the responses of the central and lower portions of Shark River Slough to flow restoration will be of less value than originally anticipated.

Like other parts of northern Everglades National Park upstream of the Shark River Slough narrowing, the Blue Shanty region exhibits clear characteristics of degradation, including low variance and unimodality of soil elevations, homogenous vegetation communities, and poor sorting of species along hydrologic gradients. This is particularly the case for the northernmost section of the sampled area, near Tamiami Trail. Southern portions of the Blue Shanty were similar, but exhibited characteristic degradation of soil and vegetation less clearly.

Trajectories and mechanisms of degradation

The spatial feedbacks that create regularly patterned landscapes are also thought to produce global bi-stability, meaning that both homogenous and heterogeneous states can occur and persist under the same external or environmental conditions. In the ridge slough landscape, a variety of local positive and distal negative feedbacks have been proposed as explanations for the regularity of ridge-slough patterns (Larsen and Harvey 2007, 2010; Ross et al. 2006, Cheng et al. 2011, Cohen et al. 2011, Heffernan et al. 2013). Models of both the sediment re-distribution hypothesis (Larsen and Harvey 2010) and the discharge competence hypothesis (Heffernan et al. 2013) suggest the potential for global bi-stability of ridge-slough pattern, but in response to different

hydrologic variables. The sediment redistribution hypothesis proposes that bistability of conserved and degraded (sawgrass-dominated) conditions should occur in response to variation in velocities during periods of high flow, which are controlled by landscape slope and by the density of sawgrass ridges. In contrast, the discharge competence hypothesis (Cohen et al. 2011) predicts global bi-stability of conserved and degraded (spatially homogenous elevations) conditions in response to variation in water level (Heffernan et al. 2013). Both models suggest that restoration of conserved conditions will be difficult to reverse once landscape degradation has occurred, but point to dramatically different management approaches (maintenance of hydroperiods vs. re-establishment of peak flows) to preserve extant conserved landscapes and to restore degraded ones. Despite this integrated theoretical and applied rationale, no empirical studies have directly assessed the potential for global bi-stability in general nor rigorously discriminated between alternative patterning mechanisms.

We propose that the inter-relationships among hydrologic regime, microtopographic structure, and vegetation composition and structure presented as part of this study means to directly assess the potential for global bi-stability in the Everglades landscape, and to at least indirectly evaluate alternative patterning mechanisms. The bi-modal distribution of soil elevation variance (Fig. 23) is consistent with the existence of two landscape equilibria: a patterned landscape characterized by well-differentiated ridges and sloughs, and a homogenous landscape with reduced topographic variation and poorly-differentiated vegetation communities. That conserved conditions were observed only within a restricted range of hydrologic conditions provides additional support for the general global bi-stability hypothesis. Finally, the occurrence of degraded landscape structure under hydrologic conditions that also support relatively conserved conditions is consistent with the existence of multiple landscape equilibria. Degraded landscapes whose contemporary hydrologic regime supports bi-modal soil elevations and distinct ridge and slough communities, but where those features are not actually observed, are concentrated in Everglades National Park and Water Conservation Area 3B; one PSU in northern WCA2 also fits this description. Current hydrologic regimes in these areas are much wetter than relatively recent historic conditions, owing to changes in hydrologic management during the last decade of the 20th century (McVoy et al. 2011). The drier conditions that predominated during most of the 20th century may have caused degradation of landscape structure in these areas, and the absence of recovery of microtopographic structure and distinct vegetation communities may reflect the existence of an alternative landscape equilibrium. However, this failure to re-establish historic pattern could also reflect other aspects of the disturbance regime that could be inhibiting re-establishment of historic landscape, or more simply the slow rate of peat formation and spatial feedbacks. Overall, however, our results tentatively support the hypothesis that patterned and homogenous landscapes represent alternative equilibria.

The divergence of microtopographic and vegetative measures of landscape heterogeneity suggest that peat degradation, rather than vegetation change, is the initial phase of landscape degradation, under dry conditions, while the reverse is the case under impounded conditions (Fig. 36). In addition to its implications for monitoring of these landscape characteristics, this observation provides some indirect support for the discharge competence hypothesis. Equilibrium solutions of a model of the discharge competence hypothesis lead to several predictions that can be assessed using data from this monitoring study (Heffernan et al. 2013). First, the model predicts that elevation differences between ridges and sloughs should increase with increasing landscape-

scale water levels, as we have observed in this study. Second, the model predicts that under increasingly dry conditions, ridge-slough elevation differences should decline smoothly, but that under increasingly wet conditions, elevation differences should collapse catastrophically when water depths exceed a threshold of tolerance for sawgrass. In this study, we observed a number of PSUs that exhibited relatively conserved vegetation structure, but dramatically reduced topographic variability under moderate to dry conditions; however, we observed high microtopographic variation and degraded vegetation in PSUs with high water depths (Fig. 36). This observation is consistent with the differential trajectories of change in response to drainage and inundation that are predicted by the model of Heffernan et al.; relatively gradual deflation of topography allows vegetation patterning and associated distinctness of communities to persist, while the rapid collapse of microtopography in response to impoundment disrupts vegetation communities and microtopographic spatial structure. In further support of this interpretation, we note that the small number of PSUs with reduced microtopography but distinct vegetation communities experience drier conditions than PSUs with degraded microtopographic and vegetative structure. Degraded landscapes occur under both extremes of hydrologic conditions, but the transient configuration of intact vegetation patterning and deflated microtopographic structure occurs only under relatively dry conditions. While more direct assessments of alternative patterning mechanisms are clearly needed, the results of this study appear to be consistent with several predictions of the discharge competence hypothesis.

Change Detection

With the completion of the full cycle of PSUs, continued monitoring of ridge-slough topography provides an opportunity to directly assess trajectories of change over 5 year time periods. Future monitoring will return to re-sample PSUs over the next 5 years, and points will be re-located as closely as possible to initial sampling points. This resampling will allow comparisons of elevation and vegetation change at individual points and across whole PSUs. Such changes may not be uniform, but may depend on local and regional hydrologic, topographic, and vegetative status. However, efforts at change detection will encounter several challenges. First, precise resampling of locations (i.e. to within the 1 m² radius of sampling plots) is virtually impossible under the field conditions and access available to us. Second, reductions in funding mean that many early PSUs included larger numbers of sampling points (up to 240), while more recent PSUs include a smaller number of sampling points (~ 120-135). This reduction in sampling density means that some locations will not be resampled. More importantly, the original design of this study was deliberately over-powered to detect the patterns and relationships analyzed in this report (i.e., bi-modality of elevation, hydrology-vegetation relationships, etc.). This over-design was established because of the value of a large amount of statistical power for change detection. Reductions in sampling density will reduce our ability to detect changes in topography, vegetation, and their relationships at the landscape scale.

Summary of Recommendations for management and restoration

The systematic sampling enabled by the GRTS design provides a rigorous grounding for a comprehensive assessment of landscape condition using a suite of established and newly developed measures. This study confirms previous findings that substantial portions of the ridge-slough landscape are severely degraded. Moreover, because historic microtopographic structure appears to be even more geographically restricted than vegetation pattern, our results suggest that the extent of degraded or degrading conditions may be greater, and the extent of historic conditions lesser, than indicated by previous studies. Continued monitoring and data analysis will enable better spatial resolution of these patterns, and the incorporation of additional metrics of landscape structure into assessment.

Our data suggest that a relatively restricted range of hydrologic conditions are best suited to the persistence of existing areas of conserved pattern. Specifically, we observed conserved microtopographic and vegetation structure under long-term mean water depths of 35-50 cm. This range of spatially-averaged hydrologic regimes would be equivalent to maintaining mean slough water depths between 50 and 65 cm, if ridge-slough elevation differences are ca. 30 cm. This empirically-derived range of hydrologic conditions is comparable to those suggested by the RASCAL model (Larsen et al. 2010). Maintaining this relatively narrow range of hydrologic conditions across extensive portions of the Water Conservation Areas and Everglades National Park will be challenging under current hydrologic management, but restoration of sheet flow would enable relatively even distribution of water depths across larger areas. Whether other hydrologic conditions (e.g., magnitude of hydrologic variability, etc.) are also necessary for the persistence of conserved conditions remains unclear.

The results of this study suggest that restoration of degraded landscapes within the historic ridge and slough region may require active intervention. Large areas within Everglades National Park, and Water Conservation Areas 3A and 3B exhibit degraded conditions, but currently experience long-term mean water depths that support relatively conserved landscape structure in central WCA3AS and elsewhere. While this observation has several possible explanations, it is consistent with, and provides tentative support for, the hypothesis that degraded landscapes represent an alternative equilibrium that will resist restoration that relies strictly on re-establishment of historic hydrologic regimes.

SPATIAL STRUCTURE IN SPECIES COMPOSITION WITHIN RIDGE AND SLOUGH LANDSCAPE

Introduction

Spatial structure in plant community composition is an expression of response to several environmental and ecological processes that determine the aggregation of individuals and species. In a self-organizing system such as the Everglades ridge-and-slough landscape (R&S), both plant communities and environmental drivers are spatially structured. In the R&S landscape, several studies have addressed the spatial structure of the landscape features, plant community patches, and environmental factors (Wu et al. 2006; Larsen et al. 2012; Yuan et al. 2015). These studies establish hydrologic processes as the basis for landscape patterning, and provide useful metrics to characterize landscape condition. In this study we describe spatial structure within the R&S landscape based on species composition, and relate the derived structural measures to landscape condition.

Traditional methods for analysis of spatial structure in plant community composition include: block-size variance analysis, particularly for data along a transect or in the form of a grid of contiguous plots (Dale 1999); multi-species spatial pattern analysis performed on the ordination axis scores (Galiano 1983); and multiscale ordination followed by partitioning of variance of ordination axes by block size (Noy-Meir and Anderson 1971). Recently, geostatistical approach in which analysis is based on distance, rather than blocks, has become increasingly popular. In this approach, the spatial structure in data is represented by an empirical variogram, which is a plot of half of squared variance, or by differences among pairs of sampling units as a function of geographical distance (Isaaks and Srivastava 1989). However, many of the spatial analyses using the semi-variogram in plant communities have focused on single community measures, e.g., species richness, plant height, biomass, etc. When the analysis includes multivariate species data, the approach is most often to summarize species composition by ordination, and perform the spatial analysis on site scores along a major ordination axis (Jonsson and Moel 1998; Kienel and Kumke 2002). However, a disadvantage of this method is the smoothing effect of ordination analyses, which are primarily used to explore community composition pattern along the underlying environmental gradients. In such secondary data, small scale spatial information is possibly diluted.

Species composition-based ordination is usually the representation of a measure of multispecies resemblance (similarity or dissimilarity) between sites. Thus, there has also been a tendency among researchers to model the distance decay of similarity by plotting resemblance values against geographic distance to show that similarity in species composition between sampling units declines with distance (Nekola and White 1999; Soininen et al. 2007). Conversely, a plot of dissimilarity against distance is the representation of an increase in dissimilarity in species composition between sites with distance, also called “dissimmogram”, the term first used by Mistral et al. (2000) to describe the multivariate spatial structure in species composition. While using a regression approach, Mistral et al. (2000) showed the variation in dissimilarity with distance, and determined the break-points and the half-distance where the increase in dissimilarity was half way to the maximum dissimilarity observed among sites. But, they did not determine the parameters comparable to those in the semi-variograms. However, Ross et al. (1997) used a similar

approach, but also quantified the spatial structure with the three parameters, *nugget*, *sill* and *range* that are commonly used in semi-variance analysis in geostatistics (Isaaks and Srivastava 1989). The *nugget* represents the portion of variance that is not spatially auto-correlated or the presence of autocorrelation at a scale finer than that present in the data, *sill* is the value of variance at which the variograms levels off, and *range* is the distance at which the variance levels off. A typical dissimogram (variogram) represents the variation in all possible directions, hence the model is considered ‘omnidirectional’. However, Ross et al. (1997) modeled the data from the sites arranged in an array of transect, and thus their model represented unidirectional variance.

In a healthy R&S landscape, ridge and slough are oriented along the water flow direction. Consequently, the spatial structure in this landscape is not the same in all directions (Watts et al. 2010). Typically, a spatial process within a landscape is termed ‘anisotropic’ if the resulting covariance differs with direction. Hence, in the R&S landscape, the *range* value of a dissimogram is expected to change with the direction, i.e. the spatial pattern will show *anisotropy*, represented by the ratio of major (along the longest axis) and minor (perpendicular to the major axis) ranges, as a quantitative expression of the degree of patterned directionality. In the healthy R&S, the presence of strong anisotropy is expected. Moreover, in a spatial model of the landscape, the nugget to sill ratio (nugget:sill) of variances represents the strength of spatial structure. In general, less than 25 percent value of the ratio suggests the presence strong spatial structure, between 25% and 75% moderately, and >75% weak spatial structure (Cambardella et al. 1994; Cohen et al. 2008). In a patterned landscape like well-conserved areas of R&S in the Everglades, the value of the nugget:sill ratio in the direction of flow is expected to be relatively low. Therefore, these both metrics are useful to address the hypothesis related to ridge-slough micro-topography associated with landscape pattern and ecological processes.

Methods

Field Sampling

Plant species data were collected in a sub-set of the Primary Sampling Units (PSU) outlined in 2007 (Philippi 2007) and monitored since 2009 (Heffernan et al. 2009). Additional areas within the footprint of the DECOMP Physical Model (PSU DPM, between WCA3AS and WCA3B), downstream of the raised section (accomplished and planned) of Tamiami Trail (PSU 513 and Blue Shanty Area in northern ENP) were also sampled (Fig. 9 in Section 1). Prior to beginning vegetation sampling, we generated locations for a number of sampling site clusters dispersed randomly within each PSU. A cluster is the group of 1 m x 1 m sampling plots placed together at certain distance. Each cluster consisted of a pivot plot and two plots randomly placed at 3 to 35 meters from the central plot in two cardinal directions, North and East. For each PSU, a total of 80 clusters were generated. In the first two years, an effort to sample the sites in each cluster was made. However, in next three years, out of 80 clusters, 45 clusters were randomly selected for field sampling. After selecting the clusters, the plot coordinates were exported to Google Earth for visual analysis of their position in the landscape. Specifically, the location of sampling plots in relation to the landscape features and their accessibility by air boat were visually assessed on the map. When a cluster was found positioned on tree islands, levees, outside PSU boundary, or was otherwise inaccessible via airboat, the clusters were dropped from the sampling cohort, and a

replacement cluster was randomly selected from the remaining clusters. Later, a tentative airboat route was created on the map. Preexisting airboat trails were included in the route wherever possible to reach the study sites.

In the field, the predetermined positions of sampling plots were located using Garmin GPS-62S with an accuracy of 2-3 m. Within each sampling plot, all submerged, floating and emergent vascular plants were identified, and their respective percent cover recorded to the nearest 5% (e.g. 5, 10, 15, 20, 25, etc.). In some PSUs, the species cover was estimated using a Braun-Blanquet scale (1 = 1-5%, 2 = 5-25%, 3 = 25-50%, 4 = 50-75%, 5 = 75-95%, and 6 = 95-100%). Three water depth measurements were also recorded within each plot. In addition, if sawgrass was present within the plot, its height above the soil was measured. Finally, within the central plot of each cluster, soil depth was recorded. To measure the soil depth within the plot, a probe was driven into the soil until bedrock was reached and the length of the probe remaining above the soil measured. Later, the measured probe length was subtracted from the known length of the probe in order to calculate soil depth by difference.

Data Analysis

Vegetation composition

For the PSUs sampled over five years (2010-2015), we summarized the species cover data by individual 1 m² plots, and calculated Bray-Curtis (BC) dissimilarity between each pair of plots within each PSU. Prior to calculating BC dissimilarity, species cover was relativized by species maxima. This standardization method in combination with the BC dissimilarity index is considered very robust in community analysis (Faith et al. 1987). The BC dissimilarity was calculated using the software DECODA (Kantvilas and Minchin 1989; Minchin 1998), and the output of the dissimilarity matrix was exported and then unfolded to use in spatial analysis.

Spatial Analysis

To examine spatial structure in species composition in relation to general flow direction in ridge-slough landscape, we first determined the direction of water flow, expressed as an angle, within each PSU. We drew lines along the major axis of identifiable landscape linear features, such as tree islands and sawgrass ridges, and used the Linear Directional Mean function in ArcMap (ArcGIS 10.2) to calculate the angle of orientation of those features. Later, the average angle value ('PSU angle') was calculated to represent the flow direction specific for each PSU. Four PSUs (8, 10, 14 & 16) sampled in first two years are within the marl prairie landscape and did not have any linear feature, thus the flow direction was not determined for these PSUs.

The geographic distance and the angle between each pair of sites were calculated using the 'Generate Near-Table' function in ArcMAP. We then extracted pairs of sites that were located at an angle of PSU angle \pm 22.5°, i.e. 45° wide range along the flow direction (Group 1), as well as perpendicular to the flow (Group 2). For instance, if the flow direction within a PSU was 174°, a pair of sites that were at an angle of 122.5° - 167.5° (or 302.5° - 347.5°) from one another were linked in the direction of flow, i.e., in Group 1 (Figure 2.1). Similarly, a pair of plots that were

located at an angle between 32.5° and 77.5° (or between 212.5° and 257.5°) from one another were joined in the direction perpendicular to the flow, i.e. in Group 2.

For pairs of sites within the targeted zone of a PSU, mean BC-dissimilarity was plotted against mean distances for pairs within 10 m intervals (lag distances), and an exponential model was used to define the spatial parameters as given below (Isaaks and Srivastava 1989):

$$BC(h) = C_0 + C * [1 - \exp\left(-3 * \frac{h}{A0}\right)]$$

where, C_0 is nugget, C = structural variance, and $A0$ = range. Since in the exponential model the sill never reaches the asymptote, the effective range is the distance at which the sill ($C_0 + C$) is within 5% of the asymptote.

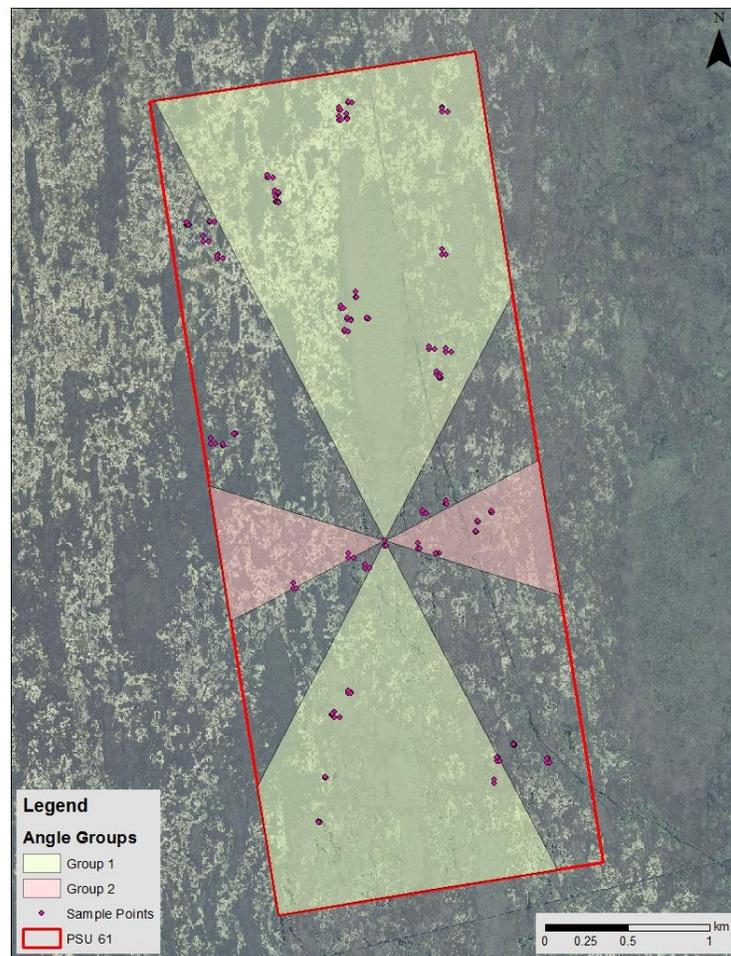


Figure 2.1: A PSU showing the angle ($\pm 22.5^\circ$) in both directions, along the flow and perpendicular to it.

With the assumption that most of spatial structure in plant community composition exists within relatively short distances, in developing our dissimogram we considered only pairs that

were 400 m apart or less. Moreover, in some of the PSUs, a solution for the exponential model did not converge, probably because of the relationship was linear or there was no spatial relationship at all. In such PSUs, we fit the linear model. In the linear model, there is no effective range, as spatial correlation occurs across the entire data range, so in these cases we considered *slope* the measure of the spatial structure, where steeper slopes represent PSUs with more spatial structure. For two PSUs (DPM and 79) that are distinctly divided by a canal/levee or road, the analysis was done separately for each part; in DPM area, IDPM (between L67A and L67C) and ODPM (SE of L67C), and for PSU-079, that were N79 and S79 for north and south of Alligator Alley, respectively.

The linear and non-linear models were fitted using `lm` and `nls` functions, respectively, in R v. 64x 3.0.1 (R Development Core Team 2013). The figures were made using the program STATISTICA 7.0 (StatSoft, Inc., Tulsa, OK).

Results

In several PSUs located in the R&S landscape of different regions in the Everglades system, plant species composition showed apparent spatial structure. In many of those PSUs, BC dissimilarity first increased with distance, and then showed no relationship (Figure 2.2). In 24 PSUs, distributed in different regions the spatial structure was apparent in both directions, and they showed some extent of anisotropy, as the range values were higher in one direction than the other (Table 2.1). However, in several other PSUs, there was no indication of spatial autocorrelation in either direction (e.g. PSU 24, 25, 37, 44, 67 & 73) or in one direction, along the flow or perpendicular to it (Table 2.1). In PSU 13, 26, 53 and 220, the relationship was linear within the range of distances considered in both directions. A weak linear relation was also observed along the direction perpendicular to flow in PSUs 29, 62 and 63.

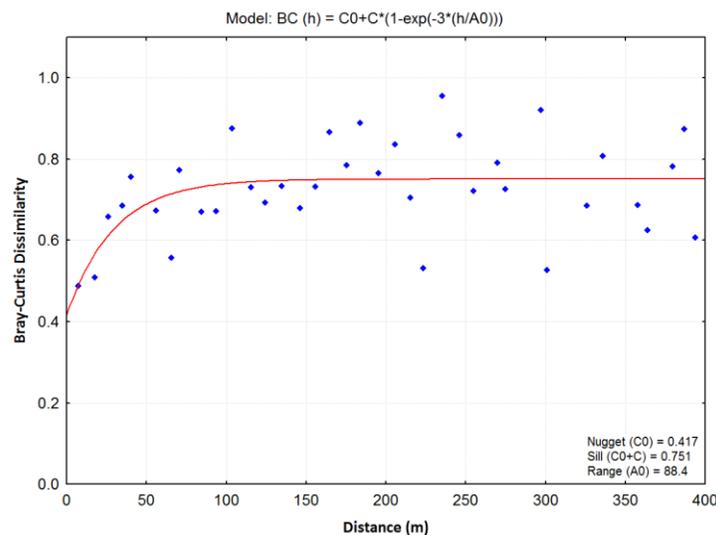


Figure 2.2: Plant species composition-based dissimogram for PSU-58. Exponential model was fitted to the distance (lag distance = 10 m) and Bray-Curtis dissimilarity data.

Table 2.1: Dissimogram models (Exponential or Linear) and model parameters for Bray-Curtis dissimilarity calculated from species cover data recorded in some of PSUs in 2013 (Year-4). Angle group-1 = flow direction, and Angle group 2 = perpendicular to the flow direction.

Year	PSU	Angle Group	Model	C0 or Intercept	Slope	C	A0	C0/(C0+C)	Anisotropy (R1/R2)
1	P001	1	Exponential	0.261		0.565	35.7	0.316	1.8
1		2	Exponential	0.449		0.391	20.3	0.534	
1	P002	1	Exponential	0		0.647	23.1	0.000	
1		2	<i>No Pattern</i>	-	-	-	-		
1	P003	1	Linear	0.654	0.0008				
1		2	Exponential	0.577		0.187	74.7	0.755	
1	P004	1	Exponential	0.606		0.251	213.8	0.707	0.8
1		2	Exponential	0.556		0.277	274.2	0.668	
1	P006	1	Exponential	0.369		0.234	372.2	0.612	
1		2	<i>No Pattern</i>	-	-	-	-		
1	P007	1	Exponential	0.449		0.368	126.4	0.550	
1		2	<i>No Pattern</i>	-	-	-	-		
1	P011	1	Exponential	0.102		0.557	43.3	0.155	0.4
1		2	Exponential	0.018		0.622	115.3	0.028	
1	P013	1	Linear	0.274	0.0004				
1		2	<i>No Pattern</i>	-	-	-	-		
1	P015	1	Exponential	0.281		0.329	38.5	0.461	
1		2	<i>No Pattern</i>	-	-	-	-		
2	P017	1	Exponential	0.488		0.222	74.7	0.688	
2		2	<i>No Pattern</i>	-	-	-	-		
2	P018	1	Exponential	0.194		0.563	172.0	0.256	2.0
2		2	Exponential	0.432		0.307	84.6	0.584	
2	P019	1	Exponential	0.242		0.515	75.0	0.320	
2		2	<i>No Pattern</i>	-	-	-	-		
2	P020	1	Exponential	0.107		0.340	28.1	0.239	
2		2	<i>No Pattern</i>	-	-	-	-		
2	P022	1	<i>No Pattern</i>	-	-	-	-		
2		2	Exponential	0.232		0.445	62.4	0.343	
2	P026	1	Linear	0.616	0.0004				0.5
2		2	Linear	0.511	0.0008				
2	P028	1	Exponential	0.312		0.207	216.2	0.602	1.2
2		2	Exponential	0.115		0.316	179.8	0.267	
2	P029	1	<i>No Pattern</i>	-	-	-	-		
2		2	Linear	0.354	0.0006				
2	P030	1	<i>No Pattern</i>	-	-	-	-		
2		2	Exponential	0.438		0.394	149.6	0.527	

Year	PSU	Angle Group	Model	C0 or Intercept	Slope	C	A0	C0/(C0+C)	Anisotropy (R1/R2)
2	P031	1	Linear	0.522	0.0003				
2		2	<i>No Pattern</i>	-	-	-	-		
3	P032	1	<i>No Pattern</i>	-	-	-	-		
3		2	Exponential	0.345		0.310	169.7	0.526	
3	P034	1	Exponential	0.425		0.378	83.2	0.529	0.4
3		2	Exponential	0.552		0.282	185.9	0.662	
3	P035	1	<i>No Pattern</i>	-	-	-	-		
3		2	Exponential	0.044		0.772	45.7	0.054	
3	P036	1	Exponential	0.312		0.446	246.0	0.412	
3		2	<i>No Pattern</i>	-	-	-	-		
3	P039	1	Exponential	0.390		0.461	71.7	0.458	0.8
3		2	Exponential	0.301		0.591	93.9	0.338	
3	P043	1	Exponential	0		0.773	48.3	0	0.3
3		2	Exponential	0.435		0.234	150.9	0.650	
3	P045	1	Exponential	0.254		0.502	62.2	0.336	0.6
3		2	Exponential	0.486		0.283	110.9	0.632	
3	P047	1	Exponential	0.388		0.225	164.5	0.633	2.6
3		2	Exponential	0.470		0.254	63.6	0.649	
3	P513	1	<i>No Pattern</i>	-	-	-	-		
3		2	Exponential	0.304		0.312	119.4	0.494	
3	IDPM	1	Exponential	0.453		0.262	266.5	0.634	1.8
3		2	Exponential	0.463		0.248	145.3	0.652	
3	ODPM	3	<i>No Pattern</i>	-	-	-	-		
3		4	<i>No Pattern</i>	-	-	-	-		
4	P051	1	Exponential	0.003		0.696	50.6	0.004	0.5
4		2	Exponential	0		0.709	98.5	0.000	
4	P052	1	Exponential	0.379		0.253	14.8	0.600	0.1
4		2	Exponential	0.578		0.147	103.3	0.797	
4	P053	1	Linear	0.517	0.0008				
4		2	Linear	0.359	0.0011				
4	P055	1	Exponential	0.089		0.705	52.5	0.112	1.3
4		2	Exponential	0.368		0.328	41.7	0.529	
4	P056	1	Exponential	0		0.641	36.3	0.000	
4		2	<i>No pattern</i>	-	-	-	-		
4	P058	1	<i>No pattern</i>	-	-	-	-		
4		2	Exponential	0.417		0.334	88.4	0.555	
4	P061	1	<i>No pattern</i>	-	-	-	-		
4		2	Exponential	0.363		0.217	28.4	0.625	
4	P062	1	<i>No pattern</i>	-	-	-	-		

Year	PSU	Angle Group	Model	C0 or Intercept	Slope	C	A0	C0/(C0+C)	Anisotropy (R1/R2)
4		2	Linear	0.553	0.0008				
4	P063	1	Exponential	0.405		0.107	214.9	0.791	
4		2	Linear	0.544	-0.0004				
4	P220	1	Linear	0.396	0.0008				
4		2	Linear	0.349	0.0005				
4	P401	1	Exponential	0.333		0.447	282.9	0.427	7.4
4		2	Exponential	0.445		0.244	38.1	0.646	
4	P402	1	Exponential	0.538		0.137	95.6	0.797	2.6
4		2	Exponential	0.394		0.294	36.1	0.573	
4	P403	1	Exponential	0.523		0.060	54.2	0.897	0.3
4		2	Exponential	0.401		0.280	210.5	0.589	
5	P050	1	Exponential	0.582		0.164	322.8	0.781	8.3
5		2	Exponential	0.114		0.618	39.0	0.156	
5	P054	1	Exponential	0.443		0.264	427.3	0.627	4.8
5		2	Exponential	0.131		0.587	89.4	0.182	
5	P065	1	Exponential	0.297		0.479	29.2	0.010	1.5
5		2	Exponential	0.273		0.496	19.3	0.355	
5	P066	1	Exponential	0.486		0.339	81.5	0.589	1.9
5		2	Exponential	0.200		0.562	41.9	0.263	
5	P068	1	Exponential	0.004		0.707	42.6	0.006	0.8
5		2	Exponential	0.296		0.407	54.2	0.421	
5	P069	1	Exponential	0.078		0.570	47.3	0.120	
5		2	<i>No Pattern</i>	-	-	-	-		
5	P071	1	Exponential	0.529		0.374	140.2	0.586	
5		2	<i>No Pattern</i>	-	-	-	-		
5	PN79	1	Exponential	0.451		0.241	112.4	0.652	2.5
5		2	Exponential	0.272		0.432	45.1	0.387	
5	PS79	1	<i>No Pattern</i>	-	-	-	-		
5		2	Exponential	0.000		0.619	29.4	0.000	

In the PSUs, in which the dissimogram computed along the flow direction and perpendicular to it revealed a similar structural behavior, the effective range, beyond which there was no further increase in dissimilarity with distance, varied from 14.8 m to 427.3.8 m. In those PSUs, the values of anisotropy ranged between 0.1 and 8.3. The value of anisotropy >1 in 14 PSUs, mostly in WCA3A (for example: PSU 66, Figure 2.3) and some areas within ENP indicated relatively healthy condition of R&S landscape (Table 2.1). In contrast, the values were <1 in 11 PSUs (for example: PSU11, Figure 2.4), indicating the deteriorated condition of R&S patterning. Those PSUs are mostly in northern (usually both sides of Alligator Alley) and southern (near Tamiami Trail) WCA3A. The study site in northern part of Blue Shanty area (P403) also showed deteriorated pattern (the value of anisotropy = 0.3).

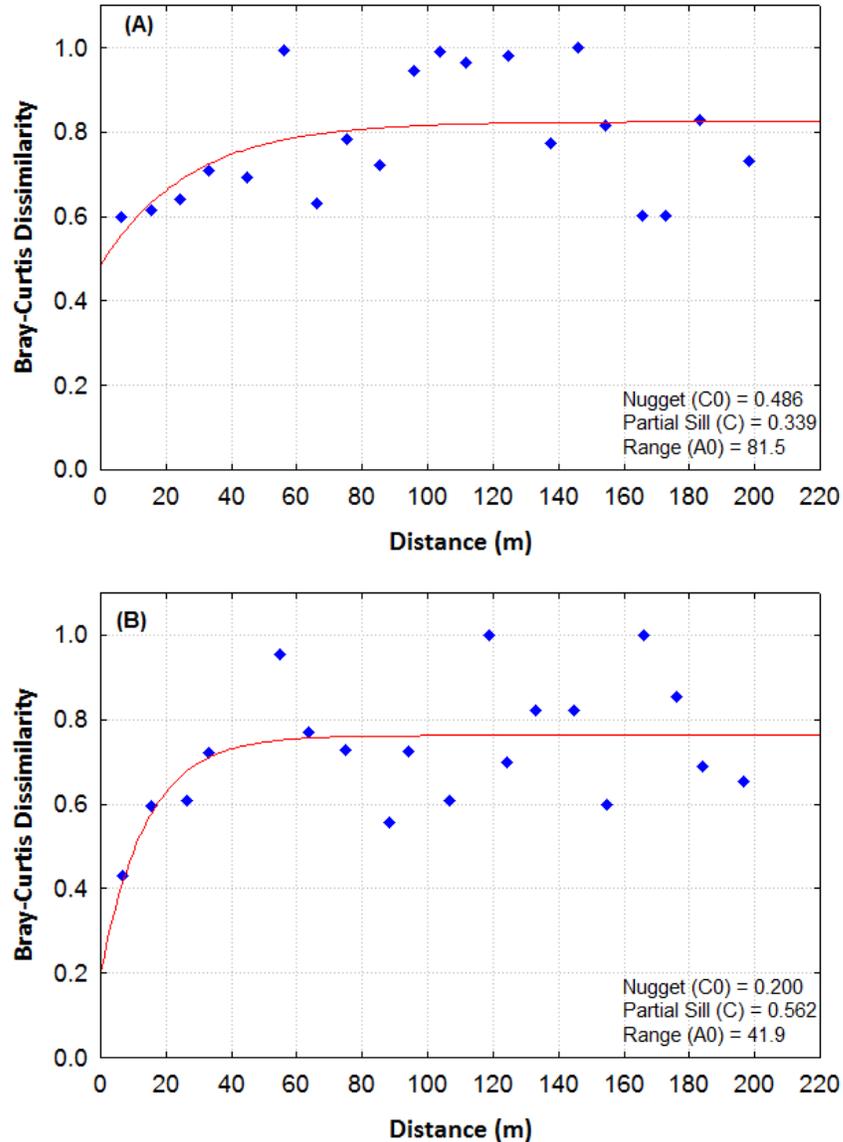


Figure 2.3: Plant species composition-based dissimogram for PSU-66. Exponential model was fitted to the distance (lag distance = 10 m) and Bray-Curtis dissimilarity data. (A) Along the flow direction. (B) Perpendicular to flow direction.

The strength of spatial structure in vegetation composition, when expressed as nugget to sill ratio (nugget:sill), also greatly varied throughout the landscape. In the PSUs that show some degree of spatial structure in both directions, along the flow as well as perpendicular to it, the values of nugget:sill ranged between 0% and 89.7% (Table 2.1). In two thirds of those PSUs, the spatial structure was moderately strong (nugget:sill values between 25 and 75%), and in one-fifth of the PSUs it was highly strong (nugget:sill <25%). In contrast, in several PSUs, the spatial structure was relatively weak (the value of nugget:sill >75%) or lacking, either in both directions or in at least one direction, along the flow or perpendicular to it (Table 2.1).

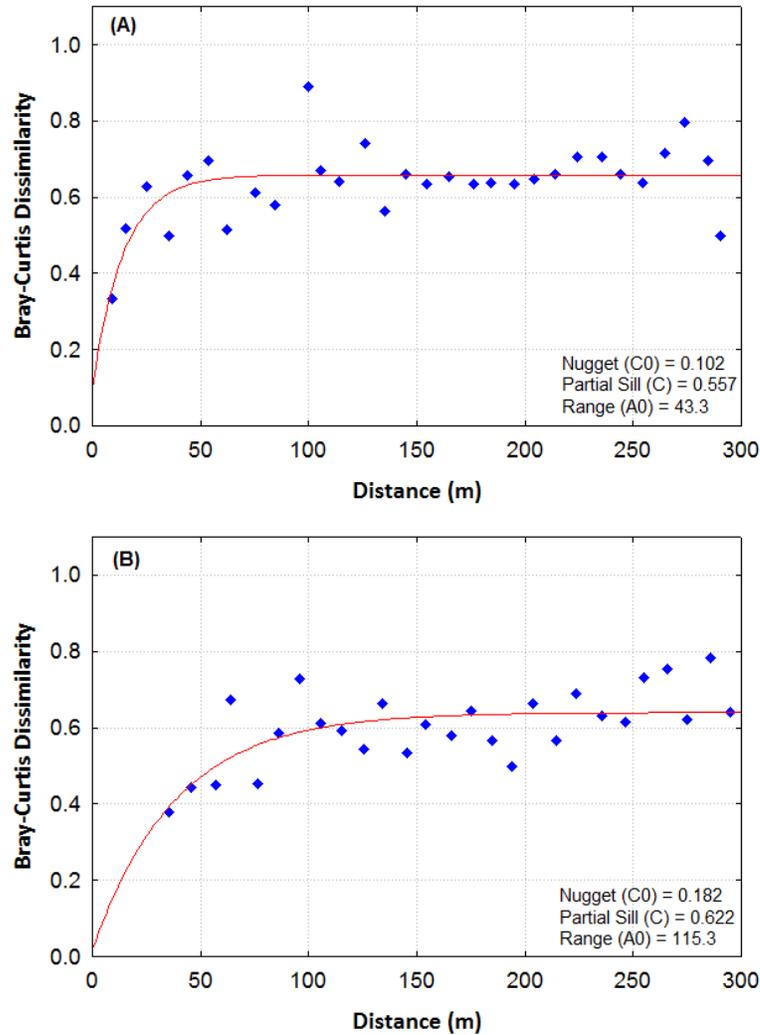


Figure 2.4: Plant species composition-based dissimogram for PSU-11. Exponential model was fitted to the distance (lag distance = 10 m) and Bray-Curtis dissimilarity data. (A) Along the flow direction. (B) Perpendicular to flow direction.

The variation in strength of spatial structure within PSUs was an indication of the condition of R&S patterning at the local level. In central Shark Slough (P018) and southern Blue Shanty Area (P401) within ENP, and in PSUs located in southern (PSU 34 & 52) and central (PSU 55 & 68) WCA-3A, nugget:sill values were lower in the flow direction than in perpendicular to the flow (Table 2.1), suggesting the presence of relatively conserved R&S landscape pattern in those areas. In contrast, R&S pattern seem to be degraded in northern Blue Shanty Area (P403) and PSUs located in northern WCA-3A (PSU 11, 39 & 51) and in WCA-2A (PSU 61) where nugget:sill values for the PSUs were relatively high in the flow direction. In some of the PSUs, deteriorated condition of R&S patterning was indicated by both values of anisotropy <1 and a lack of spatial structure in flow direction. In contrast, models showing some degree of spatial autocorrelation in the flow direction but not perpendicular to it suggested the healthy state of R&S pattern in the region. Interestingly, the region that lies in DPM area, between L67A and L67C, had anisotropy

value >1 and lower nugget:sill value in flow direction than in perpendicular to flow, both indicating relatively conserved R&S landscape. Though the strength of spatial structure there was not very strong (nugget:sill values >63%).

Discussion

Patchiness in plant community distribution is known to occur in a wide range of environments. While small-scale distribution patterns are usually influenced by interactions between organisms or between organisms and microhabitat, large scale spatial patterns in plant species composition result from environmental processes acting at different spatial scales and their feedbacks on community assembly. These patchy landscapes with multiple stable states are sensitive to changes in environmental drivers that can disrupt the positive feedback links responsible for maintaining the patterned landscape (Suding et al. 2004). In a large part of the R&S landscape, the historical vegetation pattern has disintegrated due to management-induced changes in local and regional hydrology (Science Coordination Team 2003; Larsen et al. 2011). Hence, an analysis of spatial structure in plant community composition revealed the presence of mixed conditions, with many more PSUs with degraded conditions than the PSUs with intact historical pattern.

The method of detecting spatial pattern in plant community composition used in this section is more or less similar in concept with distance decay of similarity used for different biological communities or functional groups (Nekola and White 1999; Soinen et al. 2007; Duque et al. 2009; Astroga et al. 2012). The basis for using this approach is that similarity in species composition declines with environmental or geographical distance, though the rate of decline in similarity may vary among taxonomic groups, lifeforms and functional groups (Nekola and White 1999). Several reasons for distance decay of similarity in ecological communities have been cited, including environmental heterogeneity and niche-based processes, limited dispersal or ecological drift (Nekola and White 1999; Hubbell 2001). However, these processes are not mutually exclusive and the rate of decline in similarity with distance in most communities may be jointly controlled by all these processes, but the relative strength of their effects may vary (Soinen et al. 2007). The strength of the effects of environmental heterogeneity on spatial variation in species composition also depends on landscape configurations (Jobbágy et al. 1996). In the R&S landscape that contains both physical and biological heterogeneity, there are environmentally distinct linear features like sawgrass-dominated ridges separated by a network of sloughs with submerged and floating species. Thus, similarity in species composition may decline for a distance, but again increase when pairs of distant points share similar topographic positions, e.g., ridges and sloughs. Our decision to limit the analysis by truncating the distance to 400m or less seems justified, as in several PSUs, most spatial structure was evident within 50-100 m distance. While the effective range varied between 14.8 and 427.3 m, its value was ≤ 100 m in more than two thirds of models (Table 2.1). Small effective ranges also suggest that vegetation communities are patchy and species composition changes within short distances, perhaps due to fragmentation of the landscape.

In the conserved R&S landscape where there is an apparent geometrical anisotropy, the direction of maximum spatial continuity (i.e. maximum range) is likely to prevail in the direction

of linearly oriented ridges. For such conditions, a small effective range is not expected along the flow direction, but only perpendicular to it. Hence, presence of small ranges in both directions in majority of PSUs indicated that the linear features are disintegrated, and the communities are instead in homogeneous patches. Similarly, an absence of spatial structure in the flow direction, but its presence to some degree in the direction perpendicular to flow in majority of PSUs within ENP, WCA2A, northern WCA3A, and WCA3B was also an indication of the degraded status of R&S patterning. Moreover, in the degraded part of R&S, it is likely that topographic degradation, i.e. loss of the difference in ground elevation between ridge and slough preceded the conversion of linear structures to homogeneous patches (*see the previous chapter*). Thus, the size of homogeneous vegetation patches would depend on the extent of homogenization of variation in ground elevation. In this study, the wide range of estimated nugget effect (C0) showed that the size and magnitude of variation in community patches differ within and among PSUs. The low nugget effect evident in several PSUs reveals an homogeneous plant species composition, a sign of R&S degradation. At the same time, a higher nugget effect could also be an indication of the presence of microhabitat heterogeneity at the present sampling resolution. When the nugget effect was paired with the sill, the low nugget to sill ratio indicating strong spatial structure was present along the flow direction in 11 PSUs, such as PSU 55 and southern regions of Blue Shanty. PSU 55, located in central WCA3A, was designated as relatively conserved R&S also by other metrics (XX) described in the previous chapter. In contrast, in several PSUs the nugget:sill ratio was >25% suggesting moderate to weak spatial structure in the present sampling resolution. In extreme cases, when the nugget effect approximates the sill (i.e., nugget:sill ratio ~1) such as in P403, the pattern approaches randomness. The third pattern, linearity, was observed in in one or both directions in at least 9 PSUs though the pattern was weak in many of them. In general, it results when spatial dependence occurs at scales greater than that found in the sample space. In this analysis, the distance was truncated at 400m, and a linear pattern within this distance suggests relatively long range. Thus, it is quite possible that linear features in those PSUs are well conserved. However, in PSUs 26 and 53, the slope was steeper in the direction perpendicular to flow (Ratio = 0.73) suggesting a degraded R&S condition. This contrasts with results described on the basis of bimodality in ground elevation (*see the previous chapter*).

The analysis in this section is based on the Bray-Curtis dissimilarity calculated from species abundance data collected in a series of 1 m² plots. Since small plots like these contain only a few individuals in many cases, the analytical result may be affected. An approach that is relatively robust and includes species-based probabilistic measure of decline in similarity has been suggested (Palmer, 2005). However, this method has weaknesses too, especially in that it is based on presence-absence data, and places more weight on common species, even if some rare species may make a significant contribution to the spatial pattern. Furthermore, the analysis used here is solely based on vegetation composition. Thus, an analysis that can combine dissimilarity in species composition based-dissimogram and variation in elevation-based variograms to detect congruence between spatial dependency of community composition and elevation may be more instructive. Finally, we conclude that while there is a need to improve the methodology, the present analysis of spatial variability in species composition may advance understanding of spatial patterns in existing community composition, and thus in maintaining the R&S pattern.

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